

1 **Tamm review: Does salvage logging mitigate subsequent forest disturbances?**

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3 Alexandro B. Leverkus<sup>1\*</sup>, Brian Buma<sup>2</sup>, Joseph Wagenbrenner<sup>3</sup>, Philip J. Burton<sup>4</sup>, Emanuele  
4 Lingua<sup>5</sup>, Raffaella Marzano<sup>6</sup>, Simon Thorn<sup>1</sup>

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6 <sup>1</sup> University of Würzburg, Rauhenebrach, Germany.

7 <sup>2</sup> University of Colorado, Denver, Colorado, USA.

8 <sup>3</sup> USDA Forest Service, Pacific Southwest Research Station, Arcata, California, USA.

9 <sup>4</sup> University of Northern British Columbia, Terrace, B.C., Canada.

10 <sup>5</sup> University of Padova, Legnaro (PD), Italy.

11 <sup>6</sup> University of Torino, Grugliasco (TO), Italy.

12

13 \*Correspondence to: [leverkus@ugr.es](mailto:leverkus@ugr.es)

14

15 **Abstract**

16 After natural forest disturbances such as wildfires, windstorms and insect outbreaks,  
17 salvage logging is commonly applied to reduce economic losses and mitigate subsequent  
18 disturbance risk. However, this practice is controversial due to its potential ecological  
19 impacts, and its capacity to mitigate or increase the risk of subsequent disturbances  
20 remains unclear. Salvage logging removes and alters the legacies remaining after natural  
21 disturbances, and it produces additional management legacies. Consequently, salvage  
22 logging has the potential to alter the functional connection between natural disturbances  
23 and also produce new functional connections to additional disturbances. We reviewed the  
24 efficacy of salvage logging in mitigating the risk of subsequent wildfire, insect outbreaks,  
25 hydrologic disturbances, mass movements, windthrow, browsing, and microclimatic  
26 stress. We asked: (1) *Does salvage logging modify resistance to subsequent disturbances?*  
27 (2) *Through what mechanisms do such effects operate?* Based on 96 publications, salvage  
28 logging can reduce total ecosystem fuels but increase small ground fuels and produce  
29 drier fuels in the short term, reduce bark beetle host trees and beetle-tree connectivity  
30 (though with little evidence for outbreak mitigation), magnify erosion and flood impacts of  
31 disturbance but with uncertain watershed-scale implications, increase susceptibility to  
32 windthrow at artificially created stand edges, remove the protective function of  
33 deadwood in preventing rockfall and avalanches, alter browsing pressure by modifying  
34 forage availability and hiding cover for herbivores and predators, and increase  
35 microclimatic stress due to greater radiation and temperature fluctuations. We propose a  
36 decision-making framework to evaluate the suitability of salvage logging to manage

37 subsequent disturbances. It contemplates the likelihood and impacts of both salvage  
38 logging and the subsequent disturbances. In summary, salvage logging does not  
39 necessarily prevent subsequent disturbances, and sometimes it may increase disturbance  
40 likelihood and magnitude. Forecasting the suitability of salvage logging for management  
41 goals requires assessing the mechanisms through which salvage logging effects operate  
42 under local conditions, balanced with its impacts as a disturbance itself. Managing to  
43 foster the highest-priority functions and services –such as biodiversity conservation, pest  
44 mitigation or economic return– across different parts of disturbed forest landscapes based  
45 on decision-making procedures such as the one proposed may constitute the best  
46 response to uncertain subsequent disturbances.

47

48 **Keywords:** beetle outbreak, compound disturbance, disturbance interaction, fire  
49 prevention, linked disturbance, pest control, post-disturbance management, salvage  
50 harvest, sanitation logging

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52

53 **1. Introduction**

54 Wildfires, insect outbreaks, windthrows, and other disturbances occur under natural  
55 conditions in the world's forests (White and Pickett, 1985). However, climate change,  
56 disturbance suppression, and changes in land use and management are modifying the  
57 characteristics of natural disturbances, often making them more frequent (Seidl et al.,  
58 2017). Further, disturbance events can trigger or buffer other disturbances, modifying  
59 their effects through complex disturbance interactions (Buma, 2015; Foster et al., 2016).  
60 Such interactions may have positive outcomes or produce unforeseen, negative ecological  
61 consequences, so it is important to evaluate whether post-disturbance management can  
62 modify ecological resistance to subsequent disturbances.

63 An initial disturbance can affect ecological resistance (defined here as a reduced  
64 vulnerability to disturbance given some forcing) to subsequent disturbances (Buma and  
65 Wessman, 2011). Reduced resistance can trigger an interaction chain as subsequent  
66 disturbances become more likely, extensive, or intense (Foster *et al.* 2016, Burton et al.  
67 2020). For instance, spruce trees weakened by windthrow are generally more susceptible  
68 to subsequent infestation by bark beetles (Seidl et al., 2016). Additionally, an initial  
69 disturbance can modify resilience (defined as a system's ability to recover given some  
70 disturbance) and thereby produce an interaction modification (Foster et al., 2016).  
71 Interaction modifications and interaction chains are driven by the legacies left behind by  
72 disturbances (Buma, 2015), such as surviving trees, snags and litter, and management that  
73 alters these legacies can affect the functional connection between multiple disturbances  
74 (Fig. 1).

75 Salvage logging consists of the removal of the trees affected by previous natural  
76 disturbances (Lindenmayer et al., 2008). While such management usually follows  
77 economic objectives, it also often professes to reduce the risk of subsequent disturbances  
78 (Müller et al., 2019). Salvage logging can alter the functional connection between  
79 disturbances by removing, modifying, and redistributing the biological legacies left by the  
80 initial disturbance. For example, logging after windthrow often aims to reduce beetle  
81 infestations by removing host material that can trigger population booms (Dobor et al.,  
82 2019), and post-beetle logging is often undertaken to reduce fuel loads and avoid  
83 subsequent wildfires (Donato et al., 2013).

84 Despite the above, several studies have questioned the actual efficacy of salvage  
85 logging in reducing the risk of subsequent disturbances, such as wildfire (Donato *et al.*  
86 2006) and insect outbreaks (Grodzki *et al.* 2006). Other studies indicate that salvage  
87 logging may trigger other interaction chains, including a higher risk of erosion  
88 (Wagenbrenner et al., 2016), avalanches (Wohlgemuth et al., 2017), herbivory (Castro,  
89 2013), and windthrow (Dobor et al., 2019). Several mechanisms may influence the risk of  
90 various subsequent disturbances depending on how, and over what timeframes, the  
91 logging intervention affects disturbance legacies and forest recovery. Further, broad-scale  
92 salvage logging contributes to a disturbance interaction chain itself (Leverkus et al., 2018a)  
93 and can impact ecosystem regeneration and functions (Lindenmayer et al., 2008). It is thus  
94 essential to understand the outcomes of post-disturbance logging in the context of  
95 interacting disturbances, but a comprehensive review on this topic is currently missing.

96           Here, we review whether salvage logging modifies the likelihood and  
97 characteristics of subsequent disturbances. The questions addressed are: (1) *Does salvage*  
98 *logging modify resistance to subsequent disturbances?* (2) *Through what mechanisms do*  
99 *such effects operate?* This paper is structured around each subsequent disturbance type  
100 (including wildfire, insect outbreaks, flooding and major erosional events, mass movement  
101 disturbances, windthrow, browsing, and microclimatic stress), which are individually  
102 addressed in the following section. For each subsequent disturbance, we address the  
103 mechanisms through which salvage logging may interfere in the functional connection  
104 between the initial disturbance and the specified subsequent disturbance, or produce a  
105 functional connection mediated through salvage logging itself. After reviewing the effects  
106 of salvage logging on the likelihood and magnitude of subsequent disturbances, we briefly  
107 address the mechanisms through which the resilience to those disturbances could also be  
108 affected. We end by developing a management decision-making framework that includes  
109 several considerations for managing the risk of subsequent disturbances. Our review  
110 encompasses Mediterranean, temperate, and boreal forests, in which most of the  
111 relevant literature is concentrated (Leverkus et al., 2018b; Thorn et al., 2018).

112

## 113 **2. Salvage logging effects on subsequent disturbances**

114 In this section, we briefly indicate how initial disturbance by wildfire, insect outbreak or  
115 windthrow is functionally connected to the risk of each of several possible subsequent  
116 disturbances (each addressed in one subsection). We then address the mechanisms  
117 through which salvage logging can interfere in this connection or produce new

118 mechanisms of functional connection (Fig. 1; illustrated in Figs. 2 & 3). We specify those  
119 cases where the mechanisms are specific to particular initial disturbances. Note that, as  
120 highlighted previously (Leverkus et al., 2018b; Thorn et al., 2018), most of the publications  
121 on salvage logging have measured effects over less than five years. Whereas this limits our  
122 ability to understand the dynamics of subsequent disturbance risk, we highlight the  
123 temporal patterns that have been described.

124

### 125 *2.1 Wildfire*

126 Natural disturbances such as windthrow and insect outbreaks alter the conditions for  
127 subsequent wildfires by moving some fraction of the standing live biomass pool to the  
128 dead pool on the ground (Fig. 2a). This also occurs after wildfires, as trees tend not to  
129 burn completely and fuel gradually accumulates on the ground as they collapse (Molinas-  
130 González et al., 2017). The dense collections of fine, dead fuels after initial disturbances  
131 are potentially more flammable than the more dispersed, live fuels in an intact canopy  
132 (Cannon et al., 2017), and they can produce longer and more intense burns for some time  
133 (Buma and Wessman 2011). Thus, salvage logging often focuses on fuel removal.

134 Salvage logging reduces the amount of standing live and dead fuels, and of  
135 downed coarse deadwood after windthrow, thereby reducing total ecosystem fuels and  
136 the risk and severity of crown fire in the short term (Fraver et al., 2011). But it can also  
137 alter fire risk via fuel geometry (vertical/horizontal orientation), fuel status (live/dead),  
138 microclimate, and altered fuel trajectories (Fig. 2b). A large-scale assessment across the  
139 northwestern USA (Peterson et al., 2015) and a global meta-analysis (Leverkus et al.,

140 2020) highlight that salvage logging effects on surface fuels depend on an interaction  
141 between fuel type and time, explained as follows. With the exception of windthrows,  
142 downed coarse fuels are initially unaffected by salvage logging, as the delay in snag  
143 collapse in naturally-disturbed sites and the removal of trunks from salvaged sites yield an  
144 initial absence of downed deadwood in both scenarios (Leverkus et al., 2020). The gradual  
145 collapse of dead trees progressively increases coarse fuels in unlogged areas (Peterson et  
146 al., 2015), thereby potentially increasing the severity of subsequent fire (Buma and  
147 Wessman, 2011). But whereas coarse fuels can increase the ground-level impact of fire  
148 (Monsanto & Agee 2008, Buma and Wessman 2011), it is fine fuels that primarily drive key  
149 fire characteristics such as rate of spread and flame length (Dunn and Bailey, 2015).

150         After windthrow, salvage logging can reduce the amount of fine fuels through  
151 intensive, whole-tree removal approaches (Johnson et al., 2013), but it may also increase  
152 fine fuels via mechanical abrasion during tree removal and the accumulation of slash  
153 (branches, tops and bark) during initial on-site log processing (Donato et al., 2006; Gilmore  
154 et al., 2003). This can increase fire risk compared to unsalvaged scenarios. Fine surface  
155 fuels may remain constant for decades after beetle outbreaks or fire –thereby suppressing  
156 fire likelihood for at least a decade (Buma et al., 2020)– while they immediately increase  
157 after salvage logging for up to 4-5 years (Fig. 4a; Peterson, Dodson & Harrod 2015;  
158 Leverkus *et al.* 2020). At later stages, the effect of logging is a reduction in small fuels due  
159 to faster decay in salvaged stands and the addition of dead branches from the canopy to  
160 the surface in unsalvaged stands (Fig. 4b; Peterson, Dodson & Harrod 2015).



161 Shrub and tree regeneration can outweigh dead fuels and drive fire risk as the  
162 stand develops. Salvage logging can affect vegetation, and ultimately alter flammability,  
163 through numerous mechanisms. The shrub layer can be impacted when operations kill the  
164 initial flush of regeneration (Donato et al., 2013), thereby reducing live fuel loads. But soil  
165 compaction due to mechanical action can also affect vegetation growth, thereby reducing  
166 biomass. Contrarily, highly flammable, early seral species can be favoured by salvage  
167 logging (Campbell et al., 2016). Collins *et al.* (2012) modelled that after a mountain pine  
168 beetle (*Dendroctonus ponderosae*) outbreak, enhanced spruce regeneration in unsalvaged  
169 stands would increase the likelihood of active crown fire as the stand matured. Similar  
170 conclusions were reached after the Summit Fire in Oregon, where a reburn was similarly  
171 severe across treatments because regeneration outbalanced logging-induced increases in  
172 fine fuels and eliminated any effects of salvage logging (McIver and Ottmar, 2018).  
173 However, greater flammability and fire severity can also be driven by actions associated  
174 with salvage logging, such the establishment of densely stocked conifer plantations (Fig. 3,  
175 4c; Thompson, Spies & Ganio 2007).

176 Low fuel moisture can explain higher combustion of large fuels and quicker fire  
177 spread after salvage logging (Dunn and Bailey, 2015). Salvage logging may reduce shading  
178 after fire and beetle outbreaks and thus increase ground temperature (Griffin et al., 2013;  
179 Lindenmayer et al., 2009), ultimately producing drier fuels and greater potential fire  
180 spread and intensity (Hood et al., 2017). By drying the ground, exposing mineral soil to the  
181 heat, and compacting litter, salvage logging can lead to hotter smoldering and ultimately  
182 to higher fire severity at the ground level (Fraver et al., 2011). But such effects, while

183 potentially locally important, are likely to be outweighed by those of weather (Fernandes  
184 et al., 2014) and influenced by how salvage logging modifies fuel connectivity across the  
185 landscape –topics that have received less attention than stand-scale effects in the  
186 reviewed literature.

187           In sum, the efficacy of salvage logging in altering fire risk depends on how it affects  
188 the temporal trajectory of the fuel bed after disturbance. Coarse surface fuels are  
189 generally reduced immediately after logging of windthrows, yet salvaging of fire- and  
190 insect-affected stands increases small fuels in the early years and reduces large fuels at  
191 later stages. Salvage logging can increase fire intensity by producing drier fuels and if  
192 accompanied by reforestation. Thus, salvage logging alters the composition of fuels which  
193 can affect fire behaviour and impact, but rarely appears justified as a way to reduce fire  
194 likelihood.

195

## 196 *2.2 Insect outbreaks*

197 Insect outbreaks –among which we address those produced by bark beetles (Coleoptera,  
198 Curculionidae, Scolytinae)– are a common disturbance in many temperate and boreal  
199 forests. They are particularly common after other disturbances due to weakened or dead  
200 host trees (Seidl et al., 2016; Stadelmann et al., 2013). As a result, disturbed forest stands  
201 (Fig. 4d) are often salvage logged with the goal of preventing outbreaks. Salvage logging is  
202 expected to reduce insect impacts by reducing host density and reducing connectivity  
203 between host trees (Fettig *et al.* 2007). For instance, salvage logging reduced the mortality

204 of surviving spruces by half in a post-windstorm study in Sweden (Schroeder and  
205 Lindelöw, 2002).

206           Once outbreaks reach an epidemic stage and expand to undisturbed forests,  
207 harvesting –sometimes termed sanitation logging – is still a generalised response to halt  
208 beetle expansion. Depending on particular logging methods, the mechanisms expected to  
209 stop outbreak progression involve downsizing beetle populations by removing their  
210 broods (Jönsson et al., 2012), reducing brood survival (Billings, 2011), and forcing insect  
211 dispersal to unfavourable seasons (Billings, 2011). Still, most existing indications on the  
212 mechanisms through which logging should prevent outbreaks at the stand scale are based  
213 on unsystematic observations from practice (reviewed in Fettig *et al.* 2007; Billings 2011;  
214 Six, Biber & Long 2014). Further, some studies show that stand edges generated by  
215 sanitation logging increase attractiveness for bark beetles (Fig. 3; Grodzki *et al.* 2006) and  
216 susceptibility to windthrow and insolation, which, again, favour the expansion of bark  
217 beetles as a *further* disturbance (beetles → salvage → increased windthrow → increased  
218 beetles; Modlinger & Novotný 2015).

219           At the scale of landscapes, logging aims to reduce the number of infestation spots  
220 (Gawalko, 2004) and the connectedness between host and beetle populations (Seidl et al.,  
221 2016). However, management of bark beetles at the stand scale was a poor predictor of  
222 outbreak progression in a 23-year time series in the Bavarian Forest National Park,  
223 Germany (Seidl et al., 2016). Similarly, tree mortality due to a bark beetle outbreak in the  
224 Tatra Mountains of Slovakia could not be reduced despite intensive pest management  
225 measures (Grodzki et al., 2006); the logging represented a significant disturbance at the

226 scale of the landscape, and it was unfavourable weather that ultimately reduced the  
227 outbreak (Havašová et al., 2017). In a large-scale assessment of the European spruce bark  
228 beetle in Switzerland, Stadelmann *et al.* (2013) concluded that sanitation logging (i.e., of  
229 infested trees) reduced infestation spots, yet with an effect size that was an order of  
230 magnitude smaller than that of five variables that increased spot numbers. Intensive  
231 salvage logging of large windthrows reduced the number of infestation spots (as derived  
232 from a positive effect of unsalvaged spruce volume), but outbreaks still occurred after a  
233 severe storm regardless of management (Stadelmann et al., 2013).

234         The effectiveness of logging in mitigating beetle outbreaks appears to be non-  
235 linear. As found in a Swedish study, beetle populations may limit population growth at low  
236 beetle population sizes (with a threshold of 200 females m<sup>-2</sup>), whereas under greater  
237 population densities the brood material is the limiting factor (Jönsson et al., 2012). In a  
238 large-scale modelling study in central Europe (Dobor et al., 2019), removing 100% of  
239 windthrown or infested spruce trees reduced the number of newly infested spruce trees  
240 (albeit with a greater risk of wind disturbance), yet if as little as 5% of trees were left,  
241 logging had no effect on bark beetle dynamics. In a subsequent model (Dobor et al.,  
242 2020), salvaging along roads acted as “beetle breaks” with the potential to avoid their  
243 spread. However, climate change minimised the effectiveness of salvage logging because  
244 warming temperatures increased beetle populations (Seidl et al., 2016).

245         In sum, the existing body of literature describing well-designed studies addressing  
246 the efficacy of salvage logging in preventing the explosive population growth of forest-  
247 damaging insects is sparse. This is, at least partially, caused by the inherent nature of bark

248 beetle outbreaks and the fact that unlogged control areas are scarce due to legal mandate  
249 for salvage logging (Biedermann et al., 2019). Salvage and sanitation logging can reduce  
250 infestation rates, particularly with low beetle densities and if all infested trees are  
251 removed –an often unrealistic scenario after large disturbances. The efficacy of salvage  
252 logging is lower if outbreaks are widespread, and mostly determined by regional to  
253 landscape-scale factors such as weather.

254

### 255 *2.3 Flooding and major erosional events*

256 Initial disturbances that lessen canopy cover can increase the risk of hydrologically-related  
257 disturbance by reducing transpiration and increasing net precipitation and rainfall energy  
258 at the ground. Additionally, fire can substantially reduce soil infiltration capacity through  
259 changes in soil structure, infilling of pores by ash and clay, and increases in soil water  
260 repellency (Robichaud et al., 2010). Through these processes, initial disturbances can  
261 produce a shift from subsurface to overland flow, increase the sensitivity of hydrologic  
262 processes to extreme rainfall events, and increase the risk of soil erosion and the  
263 magnitude of peak flows (Schnorbus et al., 2010). Salvage logging, through mechanisms  
264 operating at different spatial scales, can mitigate or boost the risk and magnitude of  
265 floods and erosion (Fig. 2, 3).

266           Ground-based salvage equipment can compact soil, damage understory  
267 vegetation, and magnify erosion (Gerber et al., 2002; McIver and McNeil, 2006). Such  
268 effects, combined with increases in net precipitation after canopy removal, can reduce soil  
269 water holding capacity and increase soil saturation (Fig. 4e; Prats *et al.* 2019). After

270 moderate-severity disturbances, such as insect outbreaks with partial canopy mortality,  
271 salvage logging further decreases transpiration and infiltration by killing remaining  
272 overstory and understory vegetation (Winkler et al., 2008). This may result in wetter  
273 ground, increased overland flow, greater propensity to soil disturbance (e.g., rut  
274 formation) by machinery, and ultimately increased surface runoff (Wagenbrenner et al.,  
275 2015).

276 Increased runoff is more likely to concentrate into rills, thereby increasing  
277 transport capacity and the connectivity between hillslopes and stream networks (Bryan,  
278 2000). This hydrologic connectivity particularly increases when the skid trails and roads  
279 form highly connected networks. This occurs most often in burnt areas because of the  
280 large-scale consumption of the understory and litter layers (Sosa-Pérez and Macdonald,  
281 2017). The increased connectivity would persist until the understory and organic forest  
282 floor recover, which, after severe fire, could be on the order of decades.

283 The amount of bare soil is a major control on erosion that management can  
284 influence (Fig. 4f; Robichaud, Ashmun & Sims 2010). Ground-based salvage logging  
285 typically increases bare soil by damaging vegetation and by displacing organic forest floor  
286 materials (Wagenbrenner et al., 2015). However, salvage logging can also mitigate post-  
287 fire erosion at small scales by increasing slash cover (Olsen, 2016) or creating irregular  
288 ground surfaces that reduce runoff speed (Collins and Dunne, 1988). The balance of  
289 different mechanisms can thus result in greater (Wagenbrenner et al., 2015) or equivalent  
290 erosion after salvage logging (Collins and Dunne, 1988), until vegetation regrowth  
291 outbalances initial and subsequent erosion effects.

292 Stream networks may receive greater sediment delivery through greater erosion  
293 and hydrologic connectivity. In streams where sediment loads are already high, this could  
294 lead to aggradation and loss of conveyance capacity, resulting in increased overbank flow  
295 and flooding. On the other hand, initial disturbances can increase wood delivery to stream  
296 channels (Jones and Daniels, 2008; Phillips and Park, 2009). Greater log recruitment can  
297 increase streambed stability, while it does not seem to increase flooding (Phillips and Park,  
298 2009). Salvage logging reduces the supplies of coarse wood to small streams unless  
299 riparian zones are protected, and this can produce relatively long periods of channel  
300 degradation and instability (Jones and Daniels, 2008), with concomitant risks of flooding.

301 In regions where snow melt dominates hydrographs, openings created by initial  
302 disturbances, and amplified by salvage logging, increase the radiation that reaches the  
303 snowpack, turbulent heat transfer due to higher local wind speeds, and latent heat  
304 transfer due to condensation and freezing at the snowpack surface (Alila et al., 2009). The  
305 resulting faster snow accumulation and melt rates can lead to earlier and greater peak  
306 flows (Schnorbus, 2011), and hence to greater flood risk. In western Canada, salvage  
307 logging has magnified the hydrological effects of a recent, large-scale mountain pine  
308 beetle infestation (Schnorbus, 2011), with concomitant shifts toward a greater frequency  
309 of extreme flows. However, as catchments are usually not completely affected by the  
310 initial disturbance or salvage logging, cumulative impacts at larger spatial scales may not  
311 always be detectable (e.g., Wagenbrenner *et al.* 2015; James & Krumland 2018).

312 In summary, salvage logging can affect various processes that, in turn, are related  
313 to hydrologic disturbances. By compacting soil, reducing rain-intercepting deadwood,

314 affecting advance and regenerating vegetation, altering snow-melt conditions, and  
315 amplifying rill networks, it can aggravate soil erosion, boost surface runoff, and increase  
316 the risk of flooding.

317

#### 318 *2.4 Mass movement disturbances*

319 On steep mountain slopes, forests protect against avalanches by intercepting snow in the  
320 tree crowns, affecting snow stratigraphy, and reducing the formation of homogeneous  
321 snowpack layers (Fig. 2, 4g; Teich *et al.* 2019). Major disturbances simplify the vertical  
322 structure of the forest and reduce anchoring to the ground. However, disturbed forests  
323 can still play a protective function (Lingua *et al.*, 2020). Standing dead and dying trees  
324 following bark beetle outbreaks or wildfires still intercept snow, produce heterogeneous  
325 snow stratigraphy, and buffer temperature fluctuations that weaken snow layers (Teich *et*  
326 *al.*, 2019). Following windthrows, the ground is suddenly covered by uprooted trees and  
327 boles, whose protective effect may last for several decades (Schönenberger *et al.*, 2005).

328 By removing deadwood after disturbances, salvage logging can trigger gravity-  
329 driven disturbances including avalanches, landslides, and rockfall (Fig. 2, 3; Lingua *et al.*  
330 2020). The reduction of canopy cover and fallen deadwood leads to snow profile  
331 characteristics similar to those of unforested sites (Teich *et al.*, 2019), including large,  
332 homogeneous snow packs and weaker layers resulting from wider temperature  
333 fluctuations, which are the main causes of slab avalanches (Frey and Thee, 2002). Clearing  
334 the windthrown areas also increases the risk of rockfall (Schönenberger *et al.*, 2005), as  
335 snags and deadwood otherwise obstruct falling rocks (Fig. 4h; Lingua *et al.* 2020). Salvage



336 logging in mountain forests thus increases the risk of gravity-driven disturbances during  
337 the decades preceding stand recovery, thereby prolonging the protection gap  
338 (Wohlgemuth et al., 2017).

339

#### 340 *2.5 Windthrow*

341 By creating forest edges and leaving isolated surviving trees and tree patches, natural  
342 disturbances can increase susceptibility to windthrow. Salvage logging increases wind  
343 speed (Fig. 2) and the amount of unforested area across the landscape, and it can thereby  
344 increase forest fragmentation and the extent of new forest edges (Grodzki et al., 2006),  
345 which are more susceptible to being windthrown (Modlinger and Novotný, 2015). But  
346 such effects can also be mediated by interactions with other disturbances: where salvage  
347 logging effectively dampens beetle outbreaks, it increases the availability of trees at risk of  
348 wind disturbance at long temporal scales (Dobor et al., 2019). Salvage logging can thus  
349 directly and indirectly increase the risk of windthrow in remaining forest patches, yet such  
350 effects have received little scientific attention.

351

#### 352 *2.6 Browsing*

353 Disturbances promote the growth of graminoids and resprouting shrubs, which in turn  
354 attract herbivores (particularly after fire; Foster *et al.* 2016). However, deadwood  
355 elements can act as physical impediments for large animals (Fig. 2), and thus as browsing  
356 refugia for regeneration (Castro, 2013). By opening space, salvage logging may increase  
357 browsing pressure and eventually reduce regeneration. However, different guilds of

358 herbivores may be attracted to areas managed in different ways (Hagge et al., 2019;  
359 Leverkus et al., 2013). In some places, the top-down limitation imposed by carnivores is  
360 more relevant than food availability, as shown both for ungulates (Hebblewhite et al.,  
361 2009) and rodents (Leverkus et al., 2013). Hiding cover can reduce perceived predation  
362 risk, resulting in higher damage to regeneration in uncleared sites. In contrast, browsing  
363 by some species shows no association with areas with contrasting amounts of deadwood  
364 (e.g., Kupferschmid & Bugmann 2005). Browsing pressure may also be affected by  
365 qualitative factors such as stem thickness, differential palatability of stems and leaves as a  
366 function of shade, and the use of habitats for purposes other than browsing (Faison et al.,  
367 2016). The ways in which salvage logging affects browsing thus depend on how it changes  
368 predation risk and the availability and quality of food, and on the local herbivore densities.  
369 But understanding whether such changes affect the long-term process of forest succession  
370 is limited by the short-term nature of studies.

371

## 372 *2.7 Microclimatic stress*

373 Deadwood can ameliorate harsh microclimatic conditions and thereby improve the  
374 survival and growth of regeneration (Marzano et al., 2013). Dead-tree removal may  
375 increase near-surface daytime temperatures (Vlassova and Pérez-Cabello, 2016) and  
376 reduce nightly minimum temperatures (Fontaine et al., 2010), thus increasing diurnal soil  
377 temperature fluctuations. As microclimatic requirements for seedlings are usually more  
378 restrictive than for adult plant survival, the amplification of temperature ranges can be  
379 particularly detrimental for plants regenerating from seed. As a result, the few existing

380 studies suggest that salvage logging can worsen the physiological performance of  
381 seedlings by increasing abiotic stress in semi-arid environments (Castro et al., 2011;  
382 Marañón-Jiménez et al., 2013; Marzano et al., 2013). Yet more research on interactions  
383 between natural disturbance, salvage logging, and subsequent drought is critically needed  
384 under a warming climate.

385

### 386 **3. Salvage logging effects on resilience to subsequent disturbances**

387 Above, we reviewed the effects of salvage logging on ecological resistance to subsequent  
388 disturbances. However, salvage logging may also affect resilience to subsequent  
389 disturbances, here defined as forest recovery capacity. Such effects may occur through  
390 changes in community functional composition (Taboada et al., 2018) and in the behaviour  
391 of the subsequent disturbance (Buma and Wessman, 2011). For instance, after blowdown  
392 by tornado in a hardwood-pine forest in Mississippi, salvage logging disfavoured  
393 resprouter species of low flammability, and therefore produced communities less resilient  
394 to subsequent disturbances such as wildfires (Cannon and Brewer, 2013). In a burnt pine  
395 stand in Spain, salvage logging reduced bird-mediated seed dispersal of a key resprouter  
396 tree species (Leverkus and Castro, 2017) and reduced plant diversity and the cover of  
397 post-fire seeder species (Leverkus et al., 2014), with likely implications for resilience to  
398 further fire. Conversely, species with long-range seed dispersal mechanisms are generally  
399 less impacted (Buma and Wessman, 2012). However, multiple different responses of  
400 vegetation to salvage logging have been reported (see reviews in Royo *et al.* 2016 and  
401 Taeroe *et al.* 2019), and in some cases the compositional differences induced by salvage

402 logging can be erased by the occurrence of the subsequent disturbance (Rhoades et al.,  
403 2018; Taboada et al., 2018).

404 The effects of salvage logging on resilience via changes in subsequent-disturbance  
405 characteristics can also be mediated by species traits. In subalpine forests of the Rocky  
406 Mountains, severe blowdown increased coarse-fuel loadings and the extent of high-  
407 severity burn patches (Buma and Wessman, 2011). This reduced the regeneration of  
408 serotinous *Pinus contorta* – but not that of wind-dispersed trees – through greater seed  
409 mortality and through increased seed-dispersal distances (Buma and Wessman, 2012).  
410 Post-blowdown salvage logging, which occurred prior to the fire, mitigated this interaction  
411 through reductions in coarse fuels.

412 By modelling changes to subsequent-disturbance behaviour as a result of post-  
413 logging ecosystem alterations (e.g., Collins *et al.* 2012), and by combining that with  
414 knowledge of species-specific resilience mechanisms, one could anticipate the sometimes  
415 diverging effects of salvage logging on species and ecosystem resilience. This constitutes a  
416 key direction for further research and an important aspect for decision-making.

417

#### 418 **4. Management considerations**

419 Based on our review, neither is salvage logging universally successful in preventing the  
420 subsequent disturbances addressed in sections 2.1–2.7, nor are subsequent disturbances  
421 guaranteed in its absence. Rather, salvage logging can affect particular mechanisms that  
422 modify the risk and intensity of particular disturbances. The effects of salvage logging on  
423 natural disturbances vary in space, time, and magnitude and thus its efficacy in preventing

424 subsequent disturbance highly depends on local conditions. Further, managing to avoid  
425 subsequent disturbances via salvage logging involves applying one disturbance, namely  
426 logging, to avoid another. Thus, beyond solely attempting to reduce the risk that such  
427 disturbances occur, management decisions should also address the risks associated with  
428 the disturbance of management itself.

429         Our review – in accordance with most of the literature – simplified the effects of  
430 post-disturbance management by primarily considering the stand-scale effects of applying  
431 one treatment or another, whereas post-disturbance landscapes are generally more  
432 complex. The spatially intermingled combinations of different disturbance severities,  
433 slopes, aspects, proximities to riparian areas and roads, property ownerships, and other  
434 factors, generally result in heterogeneous risks associated with subsequent disturbances  
435 across the disturbed landscape, and in different management needs. The complexity of  
436 managing subsequent-disturbance risk is further exacerbated by the potential effects of  
437 the spatial configuration of salvage logging on landscape connectivity, as highlighted by  
438 some existing modelling studies (e.g., on the effect of the spatial configuration of salvage  
439 logging along roads or in large blocks on the progression of bark beetle outbreaks; Dobor  
440 et al., 2020). Management decisions require considering such complexity to prioritise  
441 different functions across the landscape.

442         Based on our review, and on some broader questions that deserve attention, we  
443 propose a management decision-making framework regarding subsequent disturbances  
444 (Fig. 5). The starting point for this framework is the occurrence of one natural disturbance,  
445 and the steps are as follows.

446           1. Evaluate whether, and how, the legacies of the initial disturbance modify the risk  
447 of subsequent disturbances (point 1 in Fig. 5). This requires mapping aspects such as  
448 deadwood, soil properties, and weakened trees across the area of the initial disturbance.  
449 Coupled with information on climate, slope, wind exposure, beetle populations, and other  
450 variables, the risk of particular subsequent disturbances across the landscape should be  
451 assessed prior to justifying the use of management to mitigate them. For instance,  
452 whereas fire severity can depend on the severity of previous disturbance (Buma and  
453 Wessman, 2011), in other cases such functional connection is inexistent (McIver and  
454 Ottmar, 2018).

455           2. In case of increased likelihood of subsequent disturbances, forecast their  
456 expected impacts (point 2 in Fig. 5). Whereas natural disturbances can produce some  
457 negative impacts (Thom and Seidl, 2016), there is increasing recognition of their role in  
458 maintaining biodiversity, landscape heterogeneity, and other ecosystem services (e.g.,  
459 Pausas & Keeley 2019). Managers may sometimes allow disturbance chains to happen  
460 without necessarily compromising management goals (e.g., Beudert *et al.* 2015),  
461 particularly in conservation-dedicated areas (Müller et al., 2019). But if management goals  
462 are primarily economic, subsequent disturbances may reduce wood quality further and  
463 compromise management objectives.

464           3. If expected impacts from subsequent disturbances are negative, assess whether  
465 salvage logging can prevent or mitigate the functional connection between disturbances  
466 (point 3 in Fig. 5). The effects of salvage logging on the disturbance legacies that are  
467 functionally connected with subsequent disturbances, identified in point 1, can be

468 evaluated based on the results of our review (Figs. 2 & 3; Appendix S1). For instance,  
469 following small windthrows, quick salvage logging may effectively dampen subsequent  
470 beetle outbreaks. In other cases, the functional connection between natural disturbances  
471 may not be driven by deadwood legacies but by the subsequent regeneration (Thompson  
472 et al., 2007) or other mechanisms on which management may have little effect (Cannon *et*  
473 *al.* 2017). Furthermore, climatic drivers are sometimes more important than previous  
474 disturbance dynamics in defining the risk of fire (Fernandes et al., 2014; James et al.,  
475 2011), insect outbreaks (Dobor et al., 2020), and hydrologic disturbances (Gerber et al.,  
476 2002). This suggests that climatic drivers may become increasingly important in defining  
477 disturbance likelihood and extent as the climate changes, thereby overriding the effects of  
478 management (e.g., Dobor *et al.* 2020).

479         4. If salvage logging has the potential to mitigate subsequent disturbances,  
480 compare the risks (likelihood x impact) associated with those potential subsequent  
481 disturbances with the impacts of salvage logging (point 4 in Fig. 5). Salvage logging is itself  
482 a subsequent disturbance (Leverkus et al., 2018a), sometimes more intense than the  
483 initial disturbance (Modlinger and Novotný, 2015), and it can affect ecosystem functioning  
484 and biodiversity (Leverkus et al., 2020; Lindenmayer et al., 2008; Lindenmayer and Sato,  
485 2018; Thorn et al., 2020; but see Royo *et al.* 2016). Such impacts would need to be  
486 compared with the risks from the subsequent disturbance, derived from points 1 and 2  
487 above.

488         5. If salvage logging is not expected to have greater negative impacts than  
489 subsequent disturbances, assess whether salvaging produces new legacies that affect the

490 likelihood or characteristics of additional disturbances (point 5 in Fig. 5). Such legacies are  
491 included in this review (Figs. 2 & 3; Appendix S1), and they may include fine deadwood  
492 (logging slash) on the ground, compacted soil, skid trails, habitat for herbivores, etc. The  
493 new management legacies, plus the removal of natural-disturbance legacies, can produce  
494 additional feedbacks by favouring subsequent disturbances.

495           6. In case management legacies are connected with subsequent disturbances,  
496 define additional actions to mitigate them (point 6 in Fig. 5). Examples include mechanical  
497 treatments to reduce fuels such as whole-tree removal and slash burning (Gilmore et al.,  
498 2003), pre-emptive logging to remove host trees in advance of outbreaks (Fettig et al.,  
499 2007), bark scratching of downed trunks to reduce breeding substrate of bark beetles  
500 (Hagge et al., 2018), mulching to reduce bare soil and mitigate hydrologic impacts  
501 (Wagenbrenner et al., 2006), and planting in unsalvaged windthrow to speed up  
502 protection from avalanches (Wohlgemuth et al., 2017). Return to point 4 to contrast the  
503 combined effect of salvage logging plus mitigation actions with the forecasted impact of  
504 subsequent disturbances.

505           7. Following point 5, if disturbance plus management legacies are not functionally  
506 connected to further disturbances, additional decision-making criteria can be evaluated  
507 (point 7 in Fig. 5). Note that our framework allows to decide on the appropriateness of  
508 salvage logging to mitigate the impacts of subsequent disturbances, yet other legitimate  
509 aspects such as aesthetic, economic, and safety criteria (unrelated to subsequent  
510 disturbance) may also be relevant.

511



512           As the likelihood of occurrence and magnitude of subsequent disturbances tends  
513 to vary across disturbed landscapes, and as managing one disturbance can affect others in  
514 unknown ways (Dobor et al., 2019), we further suggest that, besides considering the  
515 framework of Fig. 5 to address resistance to subsequent disturbances, strategies be  
516 developed to enhance resilience, which would promote recovery after uncertain  
517 subsequent disturbances. This would increase the odds of reducing large-scale impacts of  
518 uncertain future events, and of management aimed at preventing them, in the world's  
519 forests experiencing shifting disturbance regimes.

520

## 521 **Acknowledgements**

522 ABL acknowledges postdoctoral funding from the Alexander von Humboldt Foundation  
523 and grant RTI2018-096187-J-100 from the Spanish Ministry of Science and Universities.  
524 Anika Grossmann assisted with the literature searches. The authors thank their many  
525 colleagues and students for their contributions to the insights reported here.

526

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836 **Figure captions**

837 **Figure 1.** Conceptual diagram of salvage logging in the context of multiple disturbances.

838 This review addresses how salvage logging can modify the likelihood and characteristics of  
839 subsequent disturbances, both by altering the functional connection between natural  
840 disturbances and by creating new pathways of functional connection with other  
841 disturbances.

842

843 **Figure 2.** Illustration of a forest stand affected by a) natural disturbance and b) natural  
844 disturbance and salvage logging. The effects of salvage logging (positive + or negative -) on  
845 the indicated elements and processes can increase (full symbol) or decrease (empty  
846 symbol) the likelihood, extent or magnitude of subsequent disturbances. Asterisks show  
847 particular controversy. For references, see Appendix S1.

848

849 **Figure 3.** Illustration of a forest landscape affected by a) natural disturbance and b)  
850 natural disturbance and salvage logging. The effects of salvage logging (positive + or  
851 negative -) on the indicated elements and processes can increase (full symbol) or decrease  
852 (empty symbol) the likelihood, extent or magnitude of subsequent disturbances. Asterisks  
853 show particular controversy. For a symbol legend see Figure 2. For references, see  
854 Appendix S1.

855

856 **Figure 4.** Ground fuel loads can greatly increase after post-fire logging (a), yet in the mid-  
857 term they increase more in unsalvaged areas (b) through snag fall. Actions associated with



858 salvage logging, such as planting, can produce additional effects on soils, flammability, and  
859 erosion risk (c). Recent windthrows are susceptible to bark beetle outbreaks (d). Salvage  
860 logging compacts soils and reduces infiltration (e), and it can increase erosion and  
861 sediment export risk (f). Standing dead trees reduce the risk of avalanches by preventing  
862 homogeneous snow layers (g) and of rockfall by halting downward movement (h).

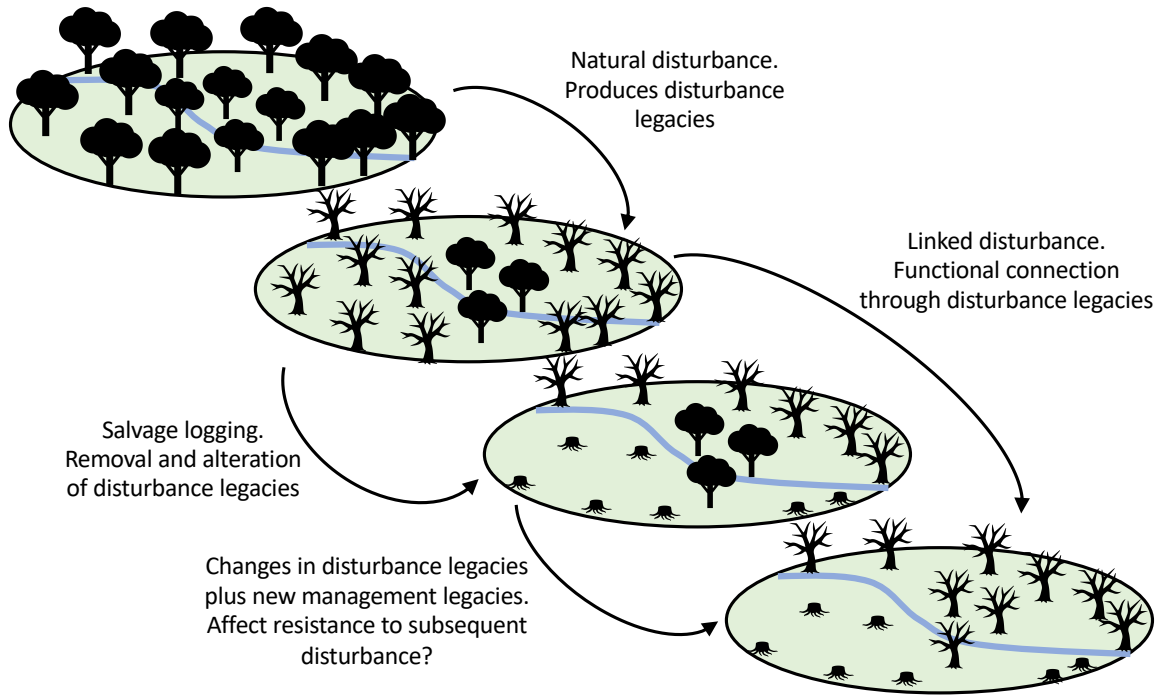
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864 **Figure 5.** Decision-making framework about salvage logging in relation to mitigating the  
865 risk of subsequent disturbances. Green boxes (with solid contour lines) indicate  
866 management assessments. The orange box (dotted contour line) indicates the ending  
867 point of the process, at which stage the needs and strategies to mitigate other impacts  
868 (from economic to biodiversity) need to be evaluated. This framework is best applied at  
869 scales small enough to encompass the spatial variation in disturbance legacies and risk of  
870 subsequent disturbances. SL = Salvage logging.

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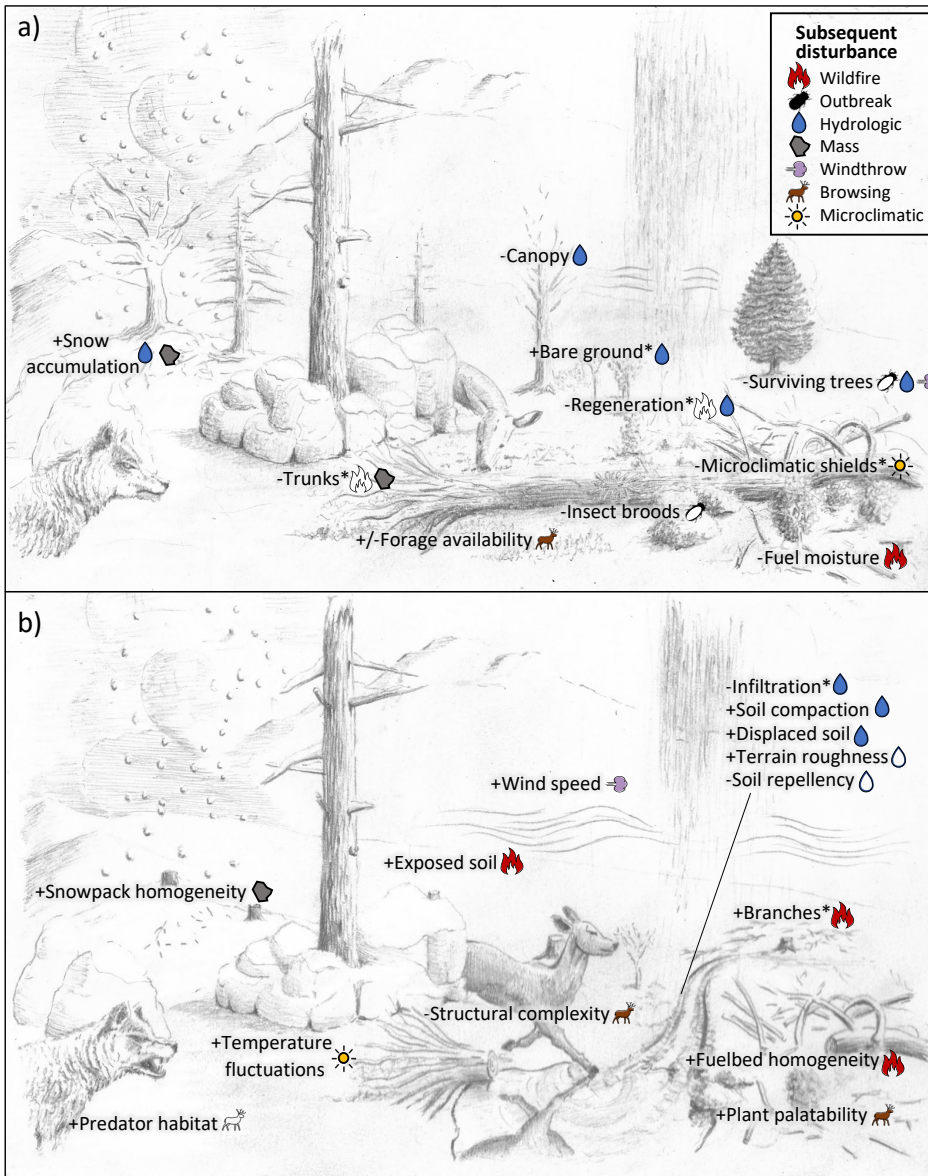
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873 **Figure 1**



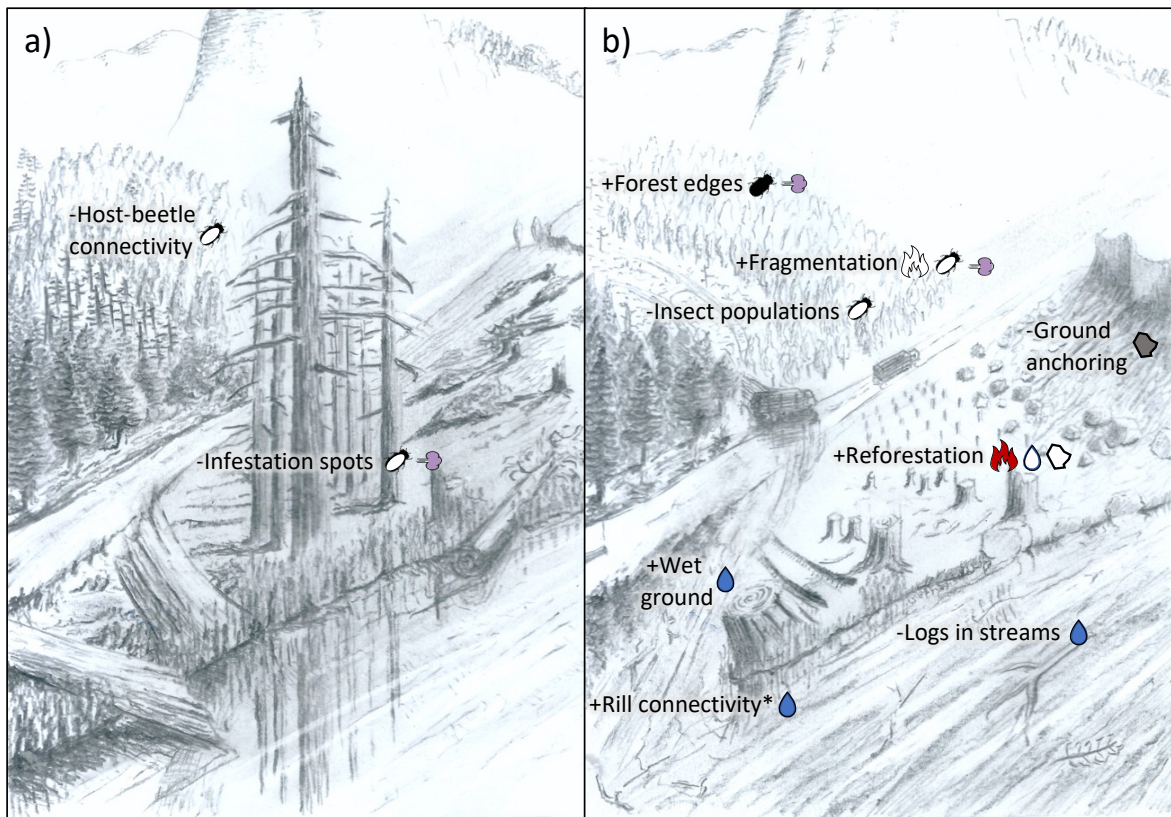
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875 **Figure 2**

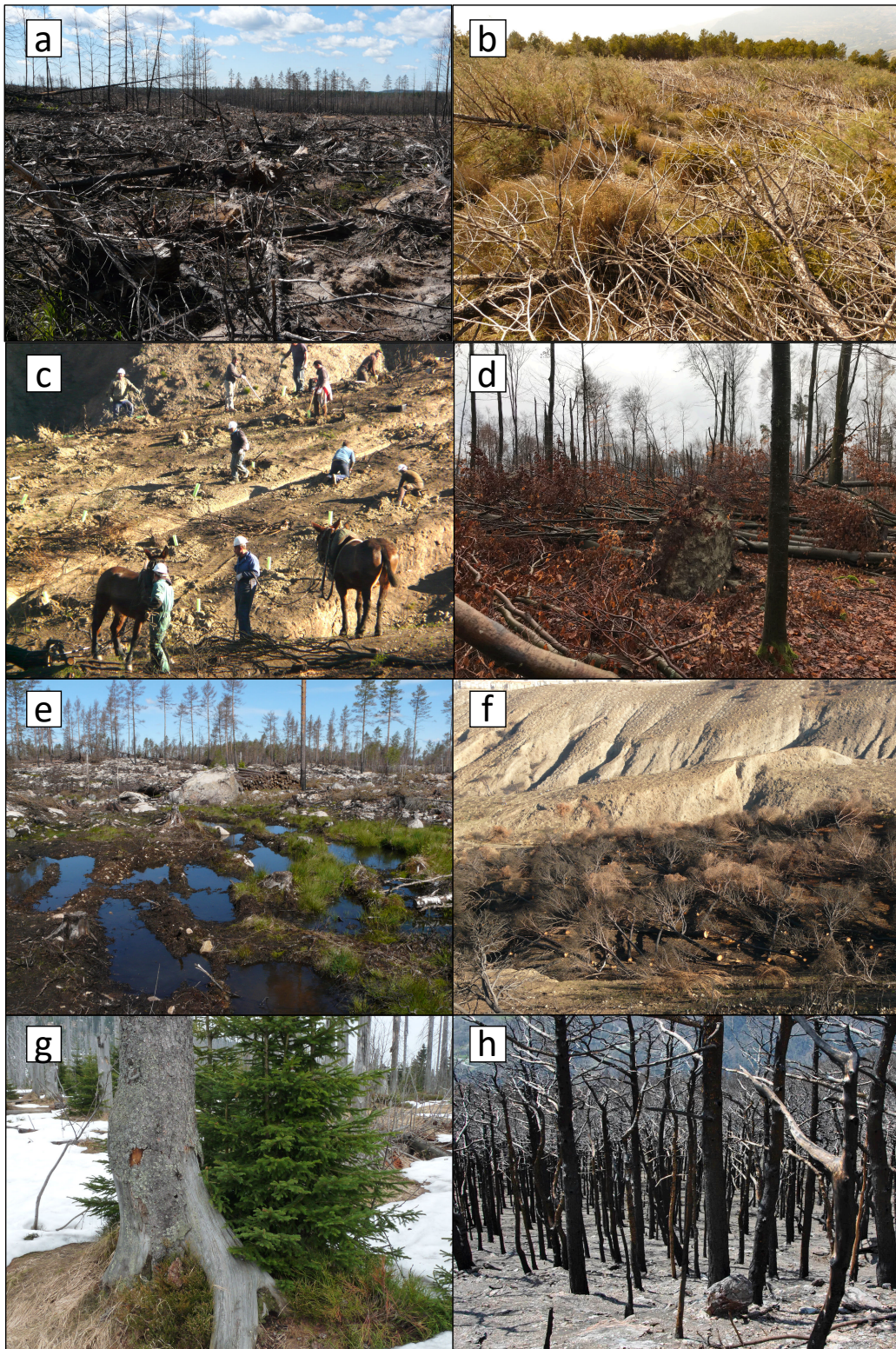


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880 **Figure 4**

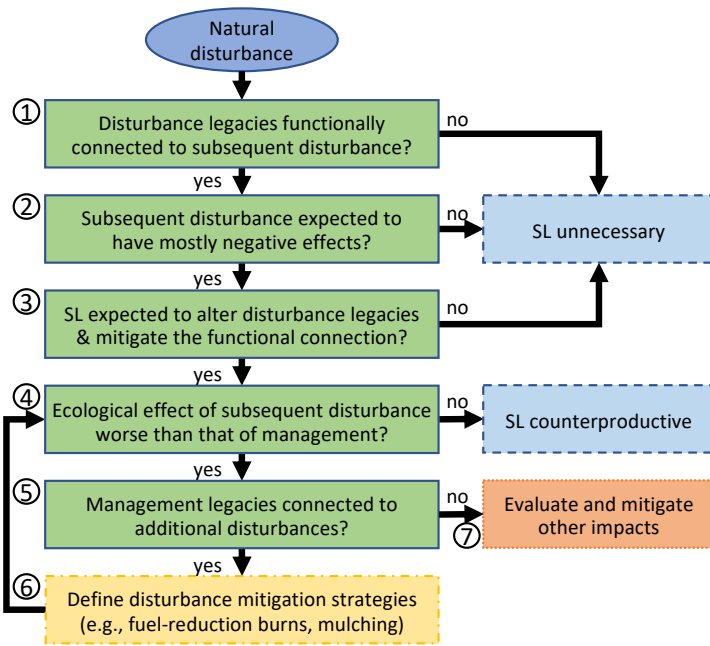


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884 **Figure 5**



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