

## **Silica micro- and nanoparticles reduce the toxicity of surfactant solutions**

Francisco Ríos\*, Alejandro Fernández-Arteaga, Mercedes Fernández-Serrano, Encarnación Jurado, Manuela Lechuga

Department of Chemical Engineering, Faculty of Sciences, University of Granada, Campus Fuentenueva s/n. 18071 Granada. Spain.

\* e-mail: rios@ugr.es. Phone: +34-958243308. Fax: +34-958248992.

### **Abstract**

In this work, the toxicity of hydrophilic fumed silica micro- and nanoparticles of various sizes (7 nm, 12 nm, and 50  $\mu\text{m}$ ) was evaluated using the luminescent bacteria *Vibrio fischeri*. In addition, the toxicity of an anionic surfactant solution (ether carboxylic acid), a nonionic surfactant solution (alkyl polyglucoside), and a binary (1:1) mixture of these solutions all containing these silica particles was evaluated. Furthermore, this work discusses the adsorption of surfactants onto particle surfaces and evaluates the effects of silica particles on the surface tension and critical micellar concentration (CMC) of these anionic and nonionic surfactants. It was determined that silica particles can be considered as non-toxic and that silica particles reduce the toxicity of surfactant solutions. Nevertheless, the toxicity reduction depends on the ionic character of the surfactants. Differences can be explained by the different adsorption behavior of surfactants onto the particle surface, which is weaker for nonionic surfactants than for anionic surfactants. Regarding the effects on surface tension, it was found that silica particles increased the surface activity of anionic surfactants and considerably reduced their CMC, whereas in the case of nonionic surfactants, the effects were reversed.

### **Keywords**

1 **Silica micro- and nanoparticles reduce the toxicity of surfactant solutions**

2 Francisco Ríos\*, Alejandro Fernández-Arteaga, Mercedes Fernández-Serrano, Encarnación  
3 Jurado, Manuela Lechuga

4 Department of Chemical Engineering, Faculty of Sciences, University of Granada, Campus  
5 Fuentenueva s/n. 18071 Granada. Spain.

6 \* e-mail: rios@ugr.es. Phone: +34-958243308. Fax: +34-958248992.

7 **Abstract**

8 In this work, the toxicity of hydrophilic fumed silica micro- and nanoparticles of various sizes  
9 (7 nm, 12 nm, and 50  $\mu\text{m}$ ) was evaluated using the luminescent bacteria *Vibrio fischeri*. In  
10 addition, the toxicity of an anionic surfactant solution (ether carboxylic acid), a nonionic  
11 surfactant solution (alkyl polyglucoside), and a binary (1:1) mixture of these solutions all  
12 containing these silica particles was evaluated. Furthermore, this work discusses the  
13 adsorption of surfactants onto particle surfaces and evaluates the effects of silica particles on  
14 the surface tension and critical micellar concentration (CMC) of these anionic and nonionic  
15 surfactants. It was determined that silica particles can be considered as non-toxic and that  
16 silica particles reduce the toxicity of surfactant solutions. Nevertheless, the toxicity reduction  
17 depends on the ionic character of the surfactants. Differences can be explained by the  
18 different adsorption behavior of surfactants onto the particle surface, which is weaker for  
19 nonionic surfactants than for anionic surfactants. Regarding the effects on surface tension, it  
20 was found that silica particles increased the surface activity of anionic surfactants and  
21 considerably reduced their CMC, whereas in the case of nonionic surfactants, the effects were  
22 reversed.

23 **Keywords**

24 Silica nanoparticles; toxicity; nonionic surfactants; anionic surfactants; nanofluid.

## 25 **1. Introduction**

26 In recent years, nanoparticles (NPs) have attracted a large amount of scientific attention  
27 because of their potential applications in biomedicine, pharmacy, materials, catalysis,  
28 cleaning, electronics and pollutant removal [1-5]. Due to their widespread use and production,  
29 it is unavoidable that they will be released into the environment and sewage [6], where they  
30 can be toxic for biota or affect waste water treatment processes. In light of a growing concern  
31 about the environmental impact of new materials, the toxicity, hazards, fate, and  
32 environmental management of nanoparticles are being widely studied [6-11].

33 Silica micro- and nanoparticles occupy a prominent position in scientific research [12-13],  
34 where they are the basis for many applications due to their stability, exceptional  
35 physicochemical properties, low toxicity and ability to be functionalized with a range of  
36 molecules and polymers. Moreover, silica particles have a uniform size and shape and are  
37 resistant to alkali and acids [6, 14-16]. Silica nanoparticles are often used together with  
38 surfactants in nanofluids and foam stabilizers, as well as being paired for uses such as the  
39 immobilization of enzymes, oil recovery, or the removal of dyes [17-21].

40 The environmental impact of surfactants has been extensively studied since they can be  
41 recalcitrant, toxic to several organisms, or detrimental to autochthons or aerobic and  
42 anaerobic microorganisms in wastewater treatment plants [22-29]. However, toxicity  
43 interactions in mixtures of surfactants and nanoparticles remain underexplored. The  
44 predictability of joint effects is of great importance for a proper environmental risk  
45 assessment and is urgently required due to the increasing development of nanofluids,  
46 nanomaterials, and nanoproducts. Previously, Oleszczuk et al. [30] studied the toxicity of  
47 ZnO, TiO<sub>2</sub> and Ni nanoparticles with cetyl trimethylammonium bromide (CTAB), triton X-  
48 100 (TX100), and 4-dodecylbenzenesulfonic acid (SDBS) and concluded that the presence of

49 surfactants considerably reduced the toxicity of the nanoparticles tested. On the other hand,  
50 Barrena et al. [31] and Stampoulis et al. [32] reported on the increase in toxicity to plants of  
51 Au, Ag and Fe<sub>3</sub>O<sub>4</sub> nanoparticles in the presence of surfactants.

52 This study reports on the joint toxicity of silica micro- and nanoparticles with an anionic  
53 surfactant (ether carboxylic acid), a nonionic surfactant (alkyl polyglucoside) and a mixture of  
54 the two, whose environmental impacts have been previously studied [23-25, 33-37]. Bacteria  
55 *Vibrio fischeri* were selected as test organisms since bacteria comprise the widest category of  
56 organisms in toxicity assessments. Bacteria *V. fischeri* have been said to have the most  
57 sensitive assay while testing a wider range of chemicals compared to other microorganism  
58 assays [38], and the comparability between tests based on them is excellent [39]. In addition,  
59 with the aim of increasing the understanding of cleaning efficiency and other aspects such as  
60 wettability and emulsifying capacity, the effect of micro- and nanoparticles on the surface  
61 tension and critical micellar concentration (CMC) of surfactants has been analyzed.

## 62 **2. Materials and methods**

### 63 *2.1. Silica micro- and nanoparticles*

64 Hydrophilic fumed silica nanoparticles (Aerosil 380 and Aerosil 200) and hydrophilic silica  
65 microparticles (Sipernat 50) were purchased from Evonik Industries AG (Essen, Germany).  
66 They have different mean diameters ( $D_m$ ) and specific surface areas ( $S$ ), covering a wide  
67 range of sizes and applications. Table 1 shows the values of  $D_m$ ,  $S$ , and the tamped density  
68 ( $d$ ) provided by the supplier. Zeta potentials (ZP) of the nanoparticles were measured in Milli-  
69 Q<sup>®</sup> water using a Zetasizer Nano (Malvern Instruments Ltd, Worcestershire, United  
70 Kingdom) (Table 1). Three measures were performed to obtain a mean ZP and its confidence  
71 interval (95%).

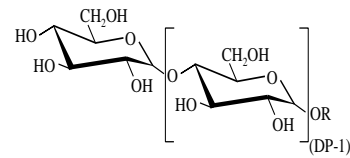
72 **Table 1 Characteristic parameters of silica micro- and nanoparticles**

Name	Abbreviation	D <sub>m</sub> , nm	S, m <sup>2</sup> /g	d, g/L	ZP, mV
<b>Aerosil® 380</b>	A380	7	380 ± 30	50	-36.0 ± 2.27
<b>Aerosil® 200</b>	A200	12	200 ± 25	50	-25.5 ± 1.89
<b>Sipernat® 50</b>	S50	50000	500	180	---

73 **2.2. Surfactants**

74 Two commercial surfactants and a binary mixture (1:1, mass basis) of them were tested: an  
 75 anionic surfactant, ether carboxylic acid (EC), supplied by KAO Corporation (Tokyo, Japan),  
 76 and a nonionic surfactant, alkyl polyglucoside (APG), manufactured by Henkel KgaA  
 77 (Düsseldorf, Germany) and provided by Sigma-Aldrich (St. Louis, MO, USA). Table 2  
 78 summarizes their characteristics.

79 **Table 2 Characteristics of the surfactants used**

Surfactant	EC-R <sub>12-14</sub> E <sub>3</sub>	APG-R <sub>8-14</sub> DP <sub>1.3</sub>
<b>Chemical Name</b>	Laureth-4 Carboxylic Acid	Coco Glucoside
<b>Commercial Name</b>	AKYPO® RLM-25	Glucopon 650EC
<b>Structure</b>	R <sub>12-14</sub> -(CH <sub>2</sub> -CH <sub>2</sub> O) <sub>3</sub> -O-CH <sub>2</sub> -COO <sup>-</sup>	
<b>CMC<sup>a</sup>, g/L</b>	29.08	33.2
<b>Active matter<sup>b</sup>, %</b>	93.1	48.6

80 R: alkyl chain length, n-C<sub>1</sub>H<sub>2i+1</sub>-

81 E: degree of ethoxylation

82 DP: average number of glucose units per molecule

83 a: measured at 25°C in Milli-Q® water

84 b: determined using infrared radiation [25]

85 **2.2. Sample preparation**

86 A 250-mg sample of silica particles was dispersed by sonication for 30 minutes (Sonorex RK  
 87 106 S, Bandelin, Berlin, Germany) in 1 L of Milli-Q® water containing 2% NaCl.  
 88 Subsequently, surfactant was added to obtain the required surfactant concentration (1.0 to

89  $4 \cdot 10^3$  mg/L). The CMC and toxicity of solutions of surfactants and silica particles were  
90 determined as described in the following sections. Moreover, the Zeta potential of the  
91 nanoparticles was measured in a saline medium (2% NaCl) and in surfactant solutions, using  
92 a Zetasizer Nano (Malvern Instruments Ltd, Worcestershire, United Kingdom) (Fig. 1).

93 Silica nanoparticles A380 and A200 were analyzed by ultra-high resolution scanning  
94 transmission electron microscope (S/TEM) and high-angle annular dark-field imaging  
95 (HAADF) FEI TITAN G2 60-300. In addition, silica nanoparticles A380 were analyzed by a  
96 transmission electron microscopy (TEM) in saline medium (2% NaCl) and in surfactant  
97 solutions with bacteria *Vibrio fischeri* using a 200 kV microscope CM20 Philips. For this, a  
98 droplet of the suspension was polymerized in pure Embed 812, after being treated with  
99 glutaraldehyde overnight and osmium tetroxide for 2 hours. Blocks were cut (50-70 nm  
100 thickness) with a diamond blade DIATOME and applied onto 300 mesh Cu grills. Ultrafine  
101 cuts were contrasted with uranyl acetate and dried in a coal evaporator.

### 102 2.3. Critical micellar concentration

103 The critical micellar concentration (CMC) is defined as the concentration of surfactant above  
104 which micelles form and any additional surfactant added to the system will go into micelles.  
105 The CMC value was estimated from plots of surface tension as a function of surfactant  
106 concentration (1.0 to  $4 \cdot 10^3$  mg/L) in a semi-log plot [40]. Surface tension has a rapid linear  
107 decrease followed by a slow decrease, and the break point in the plot shows the emergence of  
108 micelles. Surfactant solutions were prepared with the toxicity test conditions (i.e., 2% NaCl  
109 and 250 mg/L silica particles). Surface tension measurements were performed using the  
110 Wilhelmy Plate Method (BS EN 14370:2004) with a Krüss K11 tensiometer (Krüss GmbH,  
111 Hamburg, Germany) equipped with a 2-cm platinum plate. Five successive measurements  
112 were collected, and the standard deviation did not exceed  $\pm 0.1$  mN/m.

#### 113 2.4. Toxicity tests with bacteria *Vibrio fischeri*

114 The toxicity test with the photobacterium *V. fischeri* (strain NRRL-B-11177) was  
115 administered using the LumiStox<sup>®</sup> 300 system according to UNE-EN ISO 11348-2:2009  
116 guidelines (UNE-EN ISO 11348-2:2009). The bioluminescence of *V. fischeri* is inhibited by  
117 toxicants; this light inhibition can be quantified by a calibrated light meter and comparison  
118 with the light emitted by a blank sample without toxicant. Photobacteria were provided  
119 (dehydrated and frozen at -18°C) by Dr. Bruno Lange GmbH & Co., (Düsseldorf, Germany).  
120 Bacteria were reactivated in a 8 g/L C<sub>6</sub>H<sub>12</sub>O<sub>6</sub>·H<sub>2</sub>O, 20 g/L NaCl, 2.035 g/L, MgCl<sub>2</sub>·6H<sub>2</sub>O,  
121 0.30 g/L KCl and 11.9 g/L solution. Nine surfactant concentrations with the same particle  
122 concentration (250 mg/L) and a control were inoculated with the reactivated bacteria. The pH  
123 of test solutions was adjusted to 7.0 ± 0.2, with either 1N HCl or 1M NaOH, before the assay  
124 was initiated. NaCl was added to set a final chloride concentration of 2% w/w in the samples.  
125 Samples were tested in duplicate in 3-ml vessels. The light emission at the start and after 15  
126 min of contact with the toxicant was measured at a constant temperature (15°C) using a  
127 LumiStox<sup>®</sup> 300 luminometer.

128 EC<sub>50</sub> and EC<sub>20</sub> (the concentrations of surfactant that inhibited 50% and 20% of the  
129 luminescence, respectively) were calculated following the procedure described by Ríos et al.  
130 [28].

131 Three replicates tests were performed to obtain a mean EC<sub>50</sub> and its confidence interval  
132 (95%).

#### 133 2.5 Differential Scanning Calorimetry (DSC) experiments

134 To corroborate that surfactants adsorb on the nanoparticles, some Differential Scanning  
135 Calorimetry (DSC) experiments of solutions of surfactants and silica particles were carried  
136 out. These experiments are a tool to help elucidate the adsorption.

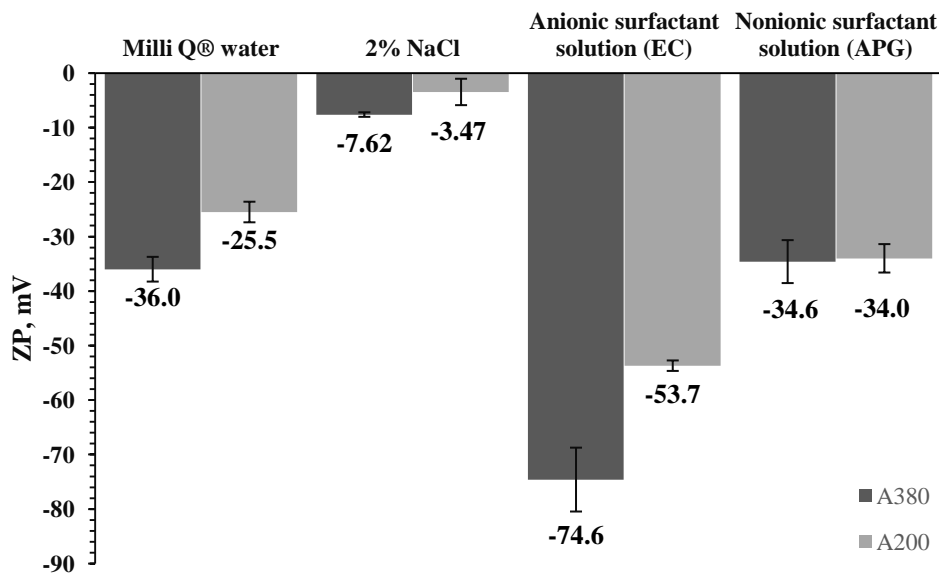
137 DSC experiments, in the temperature range from -5 to 100 °C using a scanning rate of 1  
138 °C/min, were performed with a DSC-1 instrument (Mettler Toledo). This equipment  
139 possesses a resolution of up to 0.04 μW. Samples were tightly sealed, and an empty pan was  
140 used as a reference. The amount of sample necessary to carry out the experiments was 30 μl.  
141 The absence of any changes in the signal at different scan rates indicates that the energetic  
142 transitions (related to adsorption) examined are under strict thermodynamic control as  
143 described by Chowdhry et al [41].

### 144 **3. RESULTS AND DISCUSSION**

#### 145 *3.1. Zeta potential of nanoparticles*

146 The Zeta potential of the nanoparticles was determined under the conditions of the toxicity  
147 tests, i.e., 2% NaCl with anionic (EC) or nonionic (APG) surfactant, and was compared with  
148 their ZP in Milli-Q® water (Fig. 1) to analyze the stability of dispersions. As expected, the ZP  
149 is considerably less negative in a saline medium, as the ions modify the surface electric  
150 potential of the silica particles. In anionic surfactant solutions, the ZP reaches more negative  
151 values, indicating greater stability of the particles. On the other hand, in nonionic surfactant  
152 solutions, the ZP shows no such remarkable changes. Fig. 1 shows a comparison of the ZPs at  
153 different conditions. In all cases, the ZP of A200 particles was less negative than the ZP of  
154 A380 particles, showing the greater stability of the smaller nanoparticles. The ZP of S50  
155 microparticles could not be measured due to their large size.





156

157 **Fig. 1. Comparison of Zeta potentials of nanoparticles in different conditions**

158 *3.2 Surface tension and critical micellar concentration*

159 The surface tensions of micro- and nanoparticle dispersions in Milli-Q® water at particle  
 160 concentrations in the range of 5-2500 mg/L at a constant temperature (15°C) were measured.  
 161 In all cases, the surface tension did not change with concentration. The surface tension was  
 162 approximately  $72.2 \pm 0.6$  mN/m, very close to the surface tension of pure water. Hence, silica  
 163 particles did not change the surface tension of water, probably due to their hydrophilic  
 164 character; they may not have a preference for the air-water interface. These results agree with  
 165 the values of surface tension measurements for Levasil® silica solutions found by Ma et al.  
 166 [42].

167 The surface tensions for the surfactant solutions and the micro- and nanofluids were measured  
 168 and the CMC was determined, using the conditions of the toxicity test (2% NaCl, 15°C).  
 169 CMC values and their variations are shown in Table 3. Two different results were obtained:  
 170 micro- and nanofluids with the nonionic surfactant (APG) had an increased CMC with respect  
 171 to the surfactant solution, whereas micro- and nanofluids with the anionic surfactant showed a

172 considerable reduction in their CMC. Comparison of the surface tension versus surfactant  
 173 concentration is shown in Fig. S1 in the supplementary material. In addition, it was observed  
 174 that the decreases in surface tension were the same for the three nanofluids using the same  
 175 surfactant and different nanoparticles (Fig. S2 in supplementary materials), and their CMC  
 176 values were similar or on the same order of magnitude.

177 The reduction of CMC and surface tension of the anionic surfactant is due to silica  
 178 nanoparticles increasing the surface activity of anionic surfactants [42]. As described  
 179 previously, the repulsive electrostatic forces between particles of the anionic surfactant favor  
 180 the diffusion of surfactant toward the interface, which leads to a decrease in the surface  
 181 tension [43]. According to Ma et al. [42], the presence of silica particles makes the Gibbs free  
 182 energy of adsorption and micellization more negative, and therefore, they promote the  
 183 adsorption and aggregation in micelles. In the case of the nonionic surfactant, adsorption and  
 184 electrostatic forces are much weaker, and the opposite effect is seen. Changes in the Gibbs  
 185 free energy of adsorption and micellization are negligible. Other authors [42, 44] also found a  
 186 decrease in the efficiency of nonionic surfactants with silica particles.

187 **Table 3 EC<sub>20</sub> and EC<sub>50</sub> of silica particles and surfactants (95% CI)**

Sample	CMC <sup>a</sup> , mg/L	CMC		Tox. reduction	
		variation %	EC <sub>20</sub> , mg/L	EC <sub>50</sub> , mg/L	%
<b>A380</b>	---	---	2104 ± 438	---	---
<b>A200</b>	---	----	1654 ± 398	---	----
<b>S50</b>	---	----	2434 ± 635	---	----
<b>APG</b>	63.42	----	4.38 ± 0.28	17.07 ± 0.87	----
<b>EC</b>	68.89	----	1.39 ± 0.06	3.35 ± 0.47	----
<b>A380 + APG</b>	87.91	41.45	6.22 ± 0.27	21.48 ± 1.97	25.84
<b>A200 + APG</b>	91.40	44.13	7.00 ± 0.56	24.53 ± 0.13	43.66

<b>S50 + APG</b>	91.36	44.05	5.71 ± 0.65	19.47 ± 1.76	14.04
<b>A380 + EC</b>	23.75	-65.53	4.08 ± 0.69	9.08 ± 1.23	171.13
<b>A200 + EC</b>	38.12	-44.66	3.26 ± 0.87	9.95 ± 0.86	197.12
<b>S50 + EC</b>	33.68	-51.11	2.64 ± 0.45	8.52 ± 1.50	154.50
<b>APG + EC</b>	74.66	---	1.12 ± 0.31	6.06 ± 0.85	---
<b>A380 + APG + EC</b>	65.23	-12.63	4.61 ± 0.74	11.37 ± 1.30	87.56
<b>A200 + APG + EC</b>	66.52	-10.90	4.53 ± 0.25	10.10 ± 1.25	66.73
<b>S50 + APG + EC</b>	67.37	-9.76	5.47 ± 0.31	11.13 ± 1.89	83.65

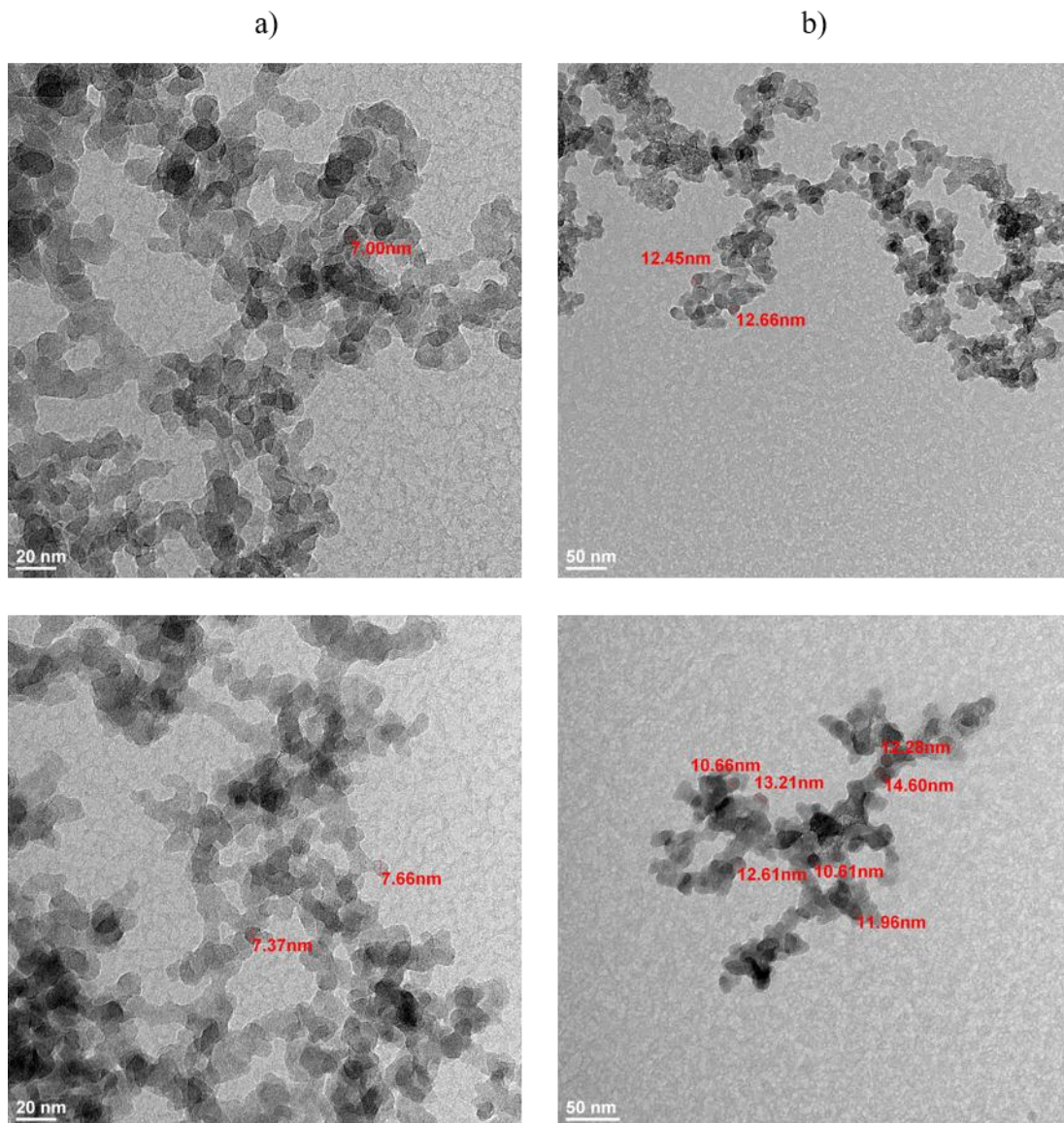
188 <sup>a</sup> 2% NaCl and 15°C

### 189 3.3. Toxicity of silica particles

190 The inhibition of the bioluminescence of *V. fischeri* after 15 min of exposure in a range of  
191 particle concentrations from 20 mg/L to 2500 mg/L was determined. Silica nanoparticles can  
192 be considered as non-toxic, since the percentage of inhibition barely exceeds 10%. In  
193 addition, these percentages were achieved at very high particle concentrations (>1000 mg/L),  
194 which are unlikely to occur in the environment and wastewater. Other studies also categorized  
195 silica particles as non-toxic to other organisms and safe to the environment [5, 7, 45]. Values  
196 of the EC<sub>50</sub> cannot be calculated because at higher concentrations of nanoparticles, the  
197 solutions become so dark that it interferes with the correct determination of the luminescence.  
198 Instead, values of EC<sub>20</sub> were estimated and are shown in Table 3.

199 The biological action of silica nanoparticles in microorganisms is related to their  
200 membranotropic properties [11]. Some studies have reported that they can easily penetrate  
201 cells and interact with lipid membranes, stimulating the generation of reactive oxygen species  
202 (ROS), which are responsible for the peroxidation of biomolecules [45-48]. Therefore, this  
203 mechanism can be supposed as the main toxicological mode of action (MoA) of silica  
204 nanoparticles to *V. fischeri*.

205 There is currently controversy about the dependence of toxicity on silica particle size and  
206 surface area [49]. In this study, we found that A200 (12 nm) were more toxic than A380 (7  
207 nm) toxic. On the other hand, S50 microparticles (50  $\mu\text{m}$ ) showed the lowest toxic effects.  
208 This fact agrees with the results from other studies in which nanoparticles under 100 nm  
209 induced more effects in cells than larger particles [50]. However, it contrasts with the results  
210 found by Adams et al. [51], who reported about similar antibacterial activity of silica particles  
211 ranging from 14 nm to 60  $\mu\text{m}$ . Adams et al. [51] also explained that nanoparticles tend to  
212 aggregate, and the actual and effective sizes of particles are highly variable and differ from  
213 their sizes in dry powders. The aggregation phenomenon was corroborated by means of a  
214 HRTEM image of A380 and A200 silica nanoparticles (Fig. 2), and the size distributions  
215 observed match with the mean diameter provided by the supplier (Table 1).

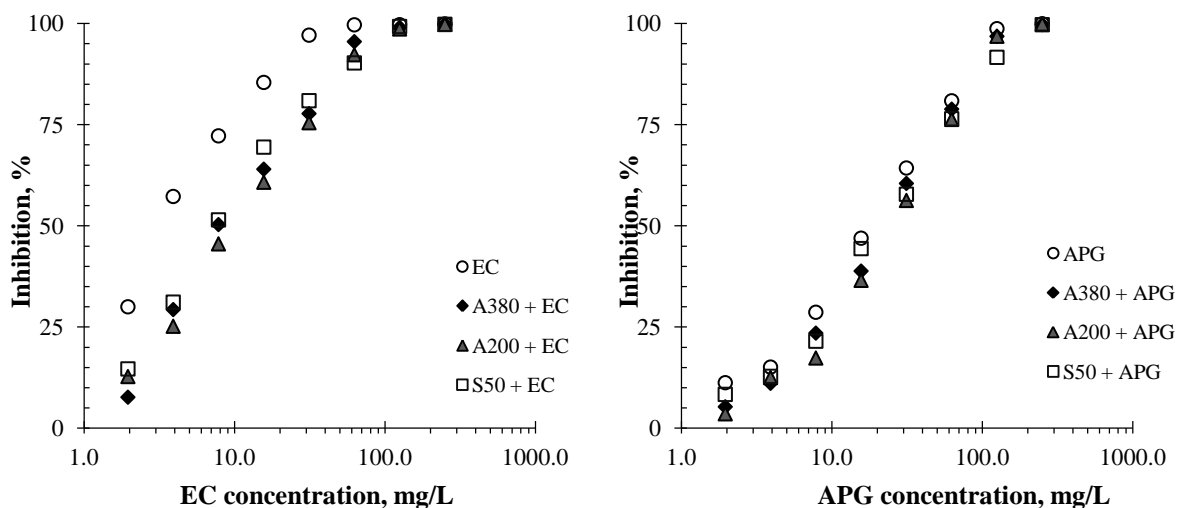


216

217 **Fig. 2. HRTEM images. a) A380 b) A200**

218 *3.4. Toxicity of surfactants and silica particles*

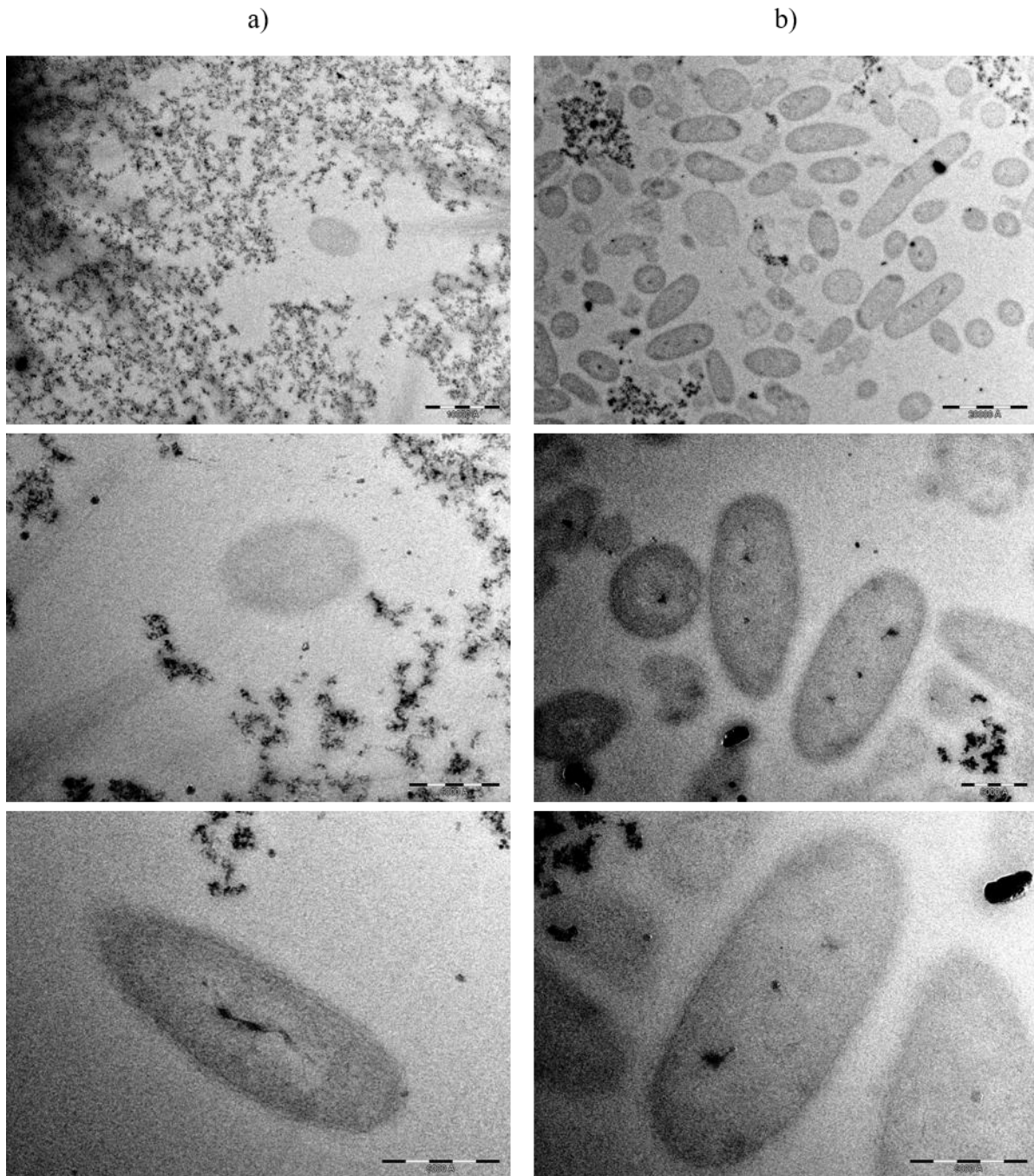
219 The luminescence inhibition of solutions of EC and APG in the presence of silica particles at  
 220 a constant concentration (250 mg/L) has been studied. Fig. 3 shows the inhibition percentages  
 221 at different surfactant concentrations, and Table 3 summarizes the calculated values of EC<sub>50</sub>  
 222 and EC<sub>20</sub>.



223

224 **Fig. 3. Dose-response curves of EC and APG at a constant concentration of silica micro- and nanoparticles**  
 225 **to *V. fischeri*.**

226 Given the non-polar nature of APG, its MoA is likely non-polar narcosis Class 1 [52],  
 227 whereas given the anionic character of EC, it is expected to act as a polar narcotic Class 2  
 228 [53]. Differences between the MoA of these surfactants and the MoA of silica particles may  
 229 indicate that they act independently from each other (response addition), which is to say that  
 230 the organism's response to the surfactant is the same whether or not particles are present. Fig.  
 231 4 shows TEM images of *V. fischeri* with silica nanoparticles A380 at the EC<sub>50</sub> concentration  
 232 determined for the nonionic surfactant APG (a) and the anionic surfactant EC (b) with A380.



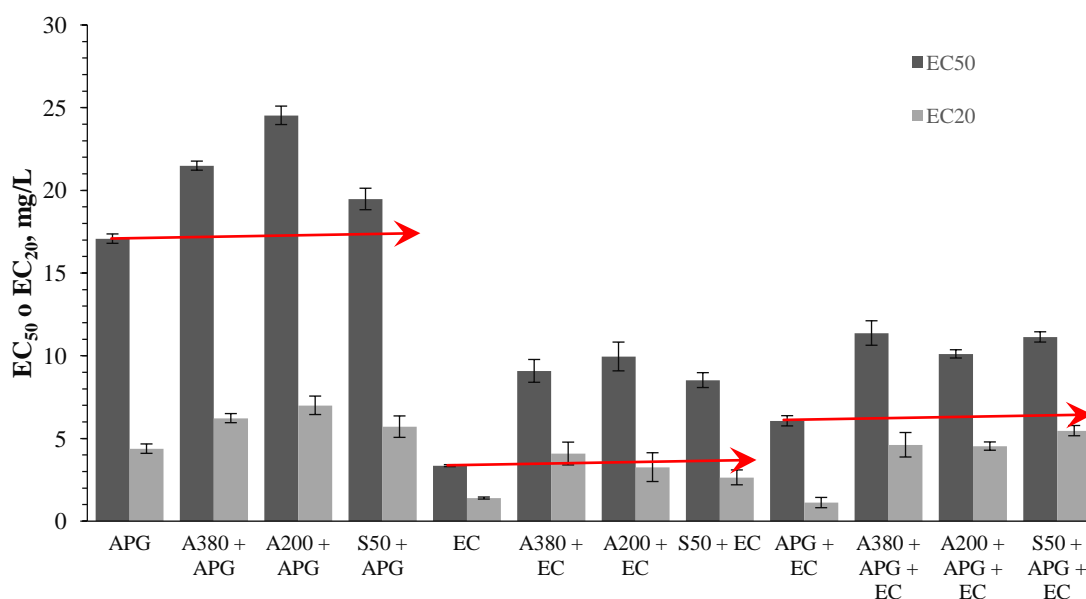
233

234 **Fig. 4. TEM images of bacteria *V. fischeri* with silica nanoparticles A380. a) APG solution (21.48 mg/L), b)**  
 235 **EC solution (9.08 mg/L).**

236 For both surfactants, the percentages of inhibition were lower in solutions with silica particles  
 237 than without them (Fig. 3). However, the differences were more pronounced in the case of the  
 238 anionic surfactant (about 25% at low concentrations). In the case of APG, these differences  
 239 are at most 10%. These deviations in the toxicity of surfactant micro- and nanofluids with

240 respect to the surfactant solutions can also be realized in the toxicity parameters  $EC_{50}$  and  
 241  $EC_{20}$  (Table 3, Fig. 5), which are above the values of the pure surfactant solutions.  
 242 Considering the  $EC_{50}$ , we calculated the toxicity reduction and show those percentages in  
 243 Table 3. This parameter makes the reduction of toxicity and the differences between the  
 244 surfactants more evident. The most remarkable case is EC + A200, for which  $EC_{50}$  increased  
 245 almost three times, and in all cases of anionic surfactant, the reduction of toxicity was higher  
 246 than 150%. In the case of the nonionic surfactant, toxicity reduction percentages ranged from  
 247 14.04 to 43.66%.

248 Additionally, we tested a binary (1:1) mixture of the anionic and nonionic surfactants (EC +  
 249 APG). Using the model of toxic units (TU) [28;54], where a TU is the sum of  $TU_i$  of the  
 250 individual components (e.g., the ratio between the surfactant concentration in a mixture ( $C_i$ )  
 251 and its toxicological acute endpoint ( $EC_{50i}$ )), it can be stated that there is no synergistic or  
 252 antagonistic effect and that the dose/concentration addition principle applies ( $TU=1 \pm 0.2$ )  
 253 [55]. In the case of micro- and nanofluids with a mixture of surfactants, a reduction in the  
 254 toxicity was also observed. Nevertheless, it is possible to think that silica particles and  
 255 surfactants act independently.



256



257 **Fig. 5. EC<sub>50</sub> and EC<sub>20</sub> of solutions of EC and APG with silica particles.**

258 When it comes to the differences in toxicity reduction depending on the particle size, A200  
259 particles promoted the highest reduction for both surfactants, while S50 gave the lowest.  
260 However, particle size has no clear influence on the toxicity of surfactant micro- and  
261 nanofluids, as it was explained before due to the particle aggregation.

262 Toxicity reduction in surfactant micro- and nanofluids is promoted by adsorption of surfactant  
263 on silica hydrophilic particles, and the differences in toxicity reduction percentages between  
264 the anionic and nonionic surfactant can be attributed to their distinct ionic characters.  
265 Adsorption of surfactants on nanoparticles has been widely studied in recent years [41; 56-  
266 59]. From these studies, it can be interpreted that adsorption onto hydrophilic silica particle  
267 surfaces represents an aggregation process akin to micelle formation in the bulk solution,  
268 which depends on the surfactant character and structure (e.g., the relative size of the  
269 hydrophilic group and hydrocarbonated chain) [60]. Two adsorption models have been  
270 examined in the literature: bilayer formation and individual micelles decorating nanoparticles  
271 [61-62].

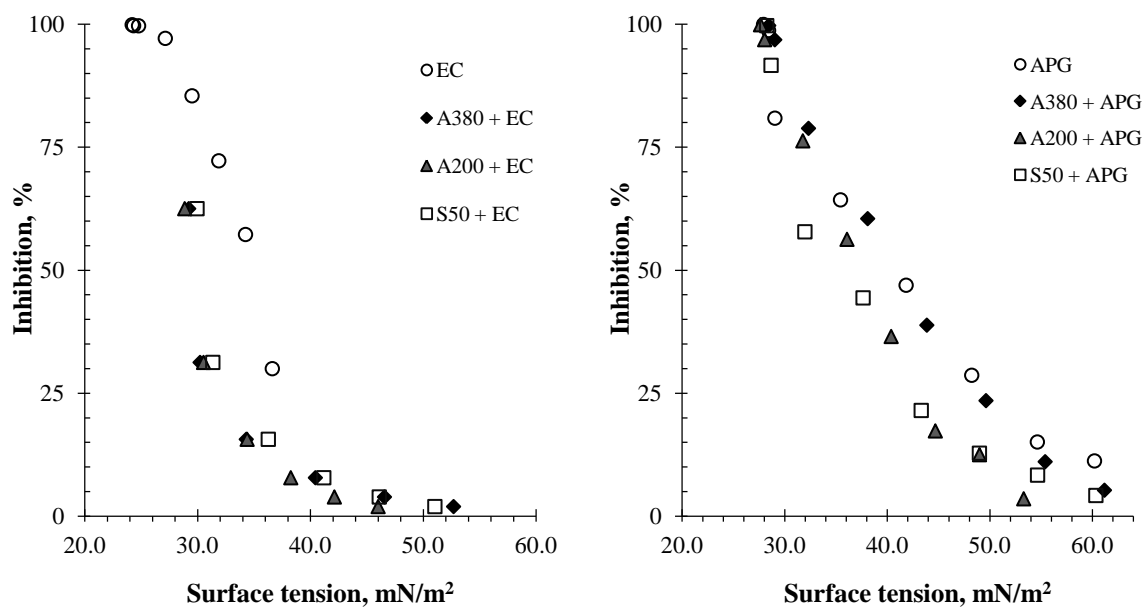
272 On the one hand, the anchoring of nonionic surfactant heads to the surface is due to weak  
273 interactions such as hydrogen bonding. When weak anchoring energies are present, micelle  
274 adsorption on the silica surface may be disfavored versus micelles in solution, implying little  
275 adsorption [58]. For example, Jurado et al. [18] investigated the interaction between silica  
276 micro- and nanoparticles and nonionic surfactants; they found that alkyl polyglucosides  
277 adsorbed slightly onto silica particles, and Lugo et al. [60] found low adsorption levels of  
278 another sugar surfactant (dodecyl- $\beta$ -maltoside) onto silica surfaces. On the other hand,  
279 nanoparticles increase the surface activity of anionic surfactants and induce electrostatic  
280 repulsion between particles. Ahualli et al. [63] confirmed the adsorption of anionic surfactants

281 onto silica particles, creating a supercharged system. This is supported by the decrease in the  
282 ZP of EC and nanoparticles found in this study (Fig. 1).

283 DSC experiments on several aqueous systems containing both the surfactants, and the three  
284 types of micro- and nanoparticles were carried out in order to prove the adsorption process,  
285 which promoted a reduction on the toxicity. The results for the studied systems showed a  
286 single, endothermic peak, corresponding to the adsorption process and typical of a first order  
287 transition (see Figure S3 in supplementary materials).

288 Fig. 6 shows the luminescence inhibition percentages versus the surface tension of the  
289 surfactant and silica particle solutions. The inhibition percentages for the anionic surfactant  
290 with silica particles increase more sharply than for the nonionic surfactant when the surface  
291 tension decreases. Furthermore, it can be appreciated that the decrease in the surface tension  
292 due to the co-occurrence of silica particles and EC does not imply an increase in the  
293 percentages of inhibition, but rather the opposite. That is, greater effectiveness of the anionic  
294 surfactant does not entail a greater effect on *V. fischeri*.

295 Adsorption of a surfactant onto nanoparticles decreases the availability of the surfactant to  
296 partition into membranes, which reduces the toxicity. Stronger adsorption of anionic  
297 surfactants than nonionic surfactants onto silica particles makes them less available and  
298 promotes a greater surfactant toxicity reduction. Moreover, particles containing adsorbed  
299 surfactant can act as carriers of surfactant toward the interface, since spontaneous adsorption  
300 of particles at the interface decreases the energy of the system [42;6].



301

302 **Fig. 6. Inhibition vs surface tension of solutions of EC and APG with silica particles.**

303 **4. Conclusions**

304 In this study, it was found that hydrophilic fumed silica micro- and nanoparticles can be  
 305 considered as non-toxic, showing percentages of inhibition that did not exceed 10% to  
 306 bacteria *V. fischeri*. Moreover trends linking particle size and toxicity could be observed,  
 307 which agree with data from the literature. In the case of mixtures of surfactant and silica  
 308 particles, silica particles reduce the toxicity of both the anionic surfactant ether carboxylic  
 309 acid (EC) and the nonionic surfactant alkyl polyglucoside (APG). However, the toxicity  
 310 reduction was much higher in the case of the anionic surfactant than the nonionic surfactant.  
 311 Differences can be explained by the adsorption of surfactant onto particle surfaces, which is  
 312 weak in the case of nonionic surfactants and stronger in the case of anionic surfactants,  
 313 causing a supercharged system. Adsorption of surfactants onto nanoparticles makes the  
 314 surfactant unavailable to partition into membranes and cause toxicity. To corroborate our  
 315 results, the surface tension and CMC of mixtures of surfactants and silica particles were  
 316 measured. As a result, it was found that silica particles increase the surface activity of the

317 anionic surfactant (EC) and reduce its CMC considerably, whereas the particles decrease the  
318 efficiency of the nonionic surfactant (APG) and increase its CMC.

## 319 Acknowledgements

320 The authors acknowledge the financial support provided by the Spanish Ministry of Economy  
321 and Competitiveness (Project CTQ2015-69658-R).

## 322 5. References

- 323 [1] Q.A. Pankhurst, J. Connolly, S.K. Jones, J. Dobson, Applications of magnetic  
324 nanoparticles in biomedicine, *J. Phys. D. Appl. Phys.* 36 (2003) R167–R181.  
325 doi:10.1088/0022-3727/36/13/201.
- 326 [2] K.S. Rao, K. El-Hami, T. Kodaki, K. Matsushige, K. Makino, A novel method for  
327 synthesis of silica nanoparticles, *J. Colloid Interface Sci.* 289 (2005) 125–131.  
328 doi:10.1016/j.jcis.2005.02.019.
- 329 [3] A.R. Tao, Nanomaterials: Nanoparticles meet their sticky ends, *Science* 351 (2016)  
330 561–562. doi:10.1126/science.aae0455.
- 331 [4] M. Soleimani, A. Khani, K. Najafzadeh,  $\alpha$ -Amylase immobilization on the silica  
332 nanoparticles for cleaning performance towards starch soils in laundry detergents, *J.*  
333 *Mol. Catal. B Enzym.* 74 (2012) 1–5. doi:10.1016/j.molcatb.2011.07.011.
- 334 [5] P. Wang, X. Wang, S. Yu, Y. Zou, J. Wang, Z. Chen, N.S. Alharbi, A. Alsaedi, T.  
335 Hayat, Y. Chen, X. Wang, Silica coated Fe<sub>3</sub>O<sub>4</sub> magnetic nanospheres for high  
336 removal of organic pollutants from wastewater, *Chem. Eng. J.* 306 (2016) 280–288.  
337 doi:10.1016/j.cej.2016.07.068.
- 338 [6] Z. Huang, G. Chen, G. Zeng, Z. Guo, K. He, L. Hu, J. Wu, L. Zhang, Y. Zhu, Z.  
339 Song, Toxicity mechanisms and synergies of silver nanoparticles in 2,4-  
340 dichlorophenol degradation by *Phanerochaete chrysosporium*, *J. Hazard. Mater.* 321  
341 (2017) 37–46. doi:10.1016/j.jhazmat.2016.08.075.
- 342 [7] G. Bystrzejewska-Piotrowska, J. Golimowski, P.L. Urban, Nanoparticles: Their  
343 potential toxicity, waste and environmental management, *Waste Manage.* 29 (2009)  
344 2587–2595. doi:10.1016/j.wasman.2009.04.001.
- 345 [8] S.K. Brar, M. Verma, R.D. Tyagi, R.Y. Surampalli, Engineered nanoparticles in  
346 wastewater and wastewater sludge - Evidence and impacts, *Waste Manage.* 30  
347 (2010) 504–520. doi:10.1016/j.wasman.2009.10.012.
- 348 [9] Y. Liu, M. Tourbin, S. Lachaize, P. Guiraud, Nanoparticles in wastewaters: Hazards,  
349 fate and remediation, *Powder Technol.* 255 (2014) 149–156.  
350 doi:10.1016/j.powtec.2013.08.025.
- 351 [10] S. Lopes, C. Pinheiro, A.M.V.M. Soares, S. Loureiro, Joint toxicity prediction of  
352 nanoparticles and ionic counterparts: Simulating toxicity under a fate scenario, *J.*  
353 *Hazard. Mater* 320 (2016) 1–9. doi:10.1016/j.jhazmat.2016.07.068.

- 354 [11] I.O. Skorochood, A.O. Roy, I.K. Kurdish, Influence of Silica Nanoparticles on  
355 Antioxidant Potential of *Bacillus subtilis* IMV B-7023., *Nanoscale Res. Lett.* 11  
356 (2016) 139. doi:10.1186/s11671-016-1348-2.
- 357 [12] I.I. Slowing, J.L. Vivero-Escoto, B.G. Trewyn, V.S.-Y. Lin, Mesoporous silica  
358 nanoparticles: structural design and applications, *J. Mater. Chem.* 20 (2010) 7924–  
359 7937. doi:10.1039/c0jm00554a.
- 360 [13] V. Mamaeva, C. Sahlgren, M. Lindén, Mesoporous silica nanoparticles in medicine-  
361 Recent advances, *Adv. Drug Deliv. Rev.* 65 (2013) 689–702.  
362 doi:10.1016/j.addr.2012.07.018.
- 363 [14] M.C. Llinàs, D. Sánchez-García, Silica nanoparticles: Preparation and applications in  
364 biomedicine | Nanopartículas de sílice: Preparación y aplicaciones en biomedicina,  
365 *Afinidad.* 71 (2014) 20–31.
- 366 [15] Y. Zhao, J. Li, S. Zhang, X. Wang, Amidoxime-functionalized magnetic mesoporous  
367 silica for selective sorption of U(VI), *RSC Adv.* 4 (2014) 32710–32717.  
368 doi:10.1039/c4ra05128a.
- 369 [16] G. Sheng, A. Alsaedi, W. Shammakh, S. Monaqueul, J. Sheng, X. Wang, H. Li, Y.  
370 Huang, Enhanced sequestration of selenite in water by nanoscale zero valent iron  
371 immobilization on carbon nanotubes by a combined batch, XPS and XAFS  
372 investigation, *Carbon* 99 (2016) 123–130. doi:10.1016/j.carbon.2015.12.013.
- 373 [17] M. Soleimani, A. Khani, K. Najafzadeh,  $\alpha$ -Amylase immobilization on the silica  
374 nanoparticles for cleaning performance towards starch soils in laundry detergents. *J.*  
375 *Mol. Catal. B Enzym.* 74 (2012) 1–5. doi:10.1016/j.molcatb.2011.07.011
- 376 [18] E. Jurado, O. Herrera-Márquez, A. Plaza-Quevedo, J.M. Vicaria, Interaction between  
377 nonionic surfactants and silica micro/nanoparticles. Influence on the cleaning of  
378 dried starch on steel surfaces, *J. Ind. Eng. Chem.* 21 (2015) 1383–1388.  
379 doi:10.1016/j.jiec.2014.06.011.
- 380 [19] M. Zargartalebi, R. Kharrat, N. Barati, Enhancement of surfactant flooding  
381 performance by the use of silica nanoparticles, *Fuel.* 143 (2015) 21–27.  
382 doi:10.1016/j.fuel.2014.11.040.
- 383 [20] Y. Zhu, X. Pei, J. Jiang, Z. Cui, B.P. Binks, Responsive Aqueous Foams Stabilized  
384 by Silica Nanoparticles Hydrophobized in Situ with a Conventional Surfactant,  
385 *Langmuir* 31 (2015) 12937–12943. doi:10.1021/acs.langmuir.5b03681.
- 386 [21] A.S. Patra, S. Ghorai, S. Ghosh, B. Mandal, S. Pal, Selective removal of toxic  
387 anionic dyes using a novel nanocomposite derived from cationically modified guar  
388 gum and silica nanoparticles, *J. Hazard. Mater.* 301 (2016) 127–136.  
389 doi:10.1016/j.jhazmat.2015.08.042.
- 390 [22] F. Aloui, S. Kchaou, S. Sayadi, Physicochemical treatments of anionic surfactants  
391 wastewater: Effect on aerobic biodegradability. *J. Hazard. Mater.* 164 (2009) 353–  
392 359. doi:10.1016/j.jhazmat.2008.08.009.
- 393 [23] E. Jurado, M. Fernández-serrano, F. Ríos, M. Lechuga, Aerobic Biodegradation of  
394 Surfactants, in: R. Chamy, F. Rosenkranz (Eds.), *Biodegradation - Life of Science,*  
395 *In-Tech, Rijeka, 2013, pp. 66-81.*

- 396 [24] M. Lechuga, M. Fernández-Serrano, E. Jurado, J. Núñez-Olea, F. Ríos, Acute  
397 toxicity of anionic and nonionic surfactants to aquatic organisms., *Ecotoxicol.*  
398 *Environ. Saf.* 125 (2016) 1–8. doi:10.1016/j.ecoenv.2015.11.027.
- 399 [25] F. Ríos, A. Fernández-Arteaga, M. Lechuga, E. Jurado, M. Fernández-Serrano,  
400 Kinetic study of the anaerobic biodegradation of alkyl polyglucosides and the  
401 influence of their structural parameters., *Environ. Sci. Pollut. Res. Int.* (2016) 1–8.  
402 doi:10.1007/s11356-016-6129-z.
- 403 [26] F. Ríos, M. Lechuga, M. Fernández-Serrano, A. Fernández-Arteaga, Aerobic  
404 biodegradation of amphoteric amine-oxide-based surfactants: Effect of molecular  
405 structure, initial surfactant concentration and pH. *Chemosphere* 171 (2017) 324–331.  
406 doi:10.1016/j.chemosphere.2016.12.070
- 407 [27] F. Ríos, M. Olak-Kucharczyk, M. Gmurek, S. Ledakowicz, Removal efficiency of  
408 anionic surfactants from water during UVC photolysis and advanced oxidation  
409 process in H<sub>2</sub>O<sub>2</sub>/UVC system. *Arch. Environ. Prot.* 43 (2017) 20–26.  
410 doi:10.1515/aep-2017-0003
- 411 [28] F. Ríos, A. Fernández-Arteaga, M. Lechuga, M. Fernández-Serrano,  
412 Ecotoxicological characterization of polyoxyethylene glycerol ester nonionic  
413 surfactants and their mixtures with anionic and nonionic surfactants. *Environ. Sci.*  
414 *Pollut. Res.* 24 (2017) 10121-10130. doi: 10.1007/s11356-017-8662-9
- 415 [29] F. Ríos, M. Lechuga, A. Fernández-Arteaga, E. Jurado, M. Fernández-Serrano,  
416 Anaerobic digestion of amine-oxide-based surfactants: biodegradation kinetics and  
417 inhibitory effects. *Biodegradation* 28 (2017) 303–312. doi: 10.1007/s10532-017-  
418 9797-6
- 419 [30] P. Oleszczuk, I. Joško, E. Skwarek, Surfactants decrease the toxicity of ZnO, TiO<sub>2</sub>  
420 and Ni nanoparticles to *Daphnia magna*, *Ecotoxicology* 24 (2015) 1923–1932.  
421 doi:10.1007/s10646-015-1529-2.
- 422 [31] R. Barrena, E. Casals, J. Colón, X. Font, A. Sánchez, V. Puentes, Evaluation of the  
423 ecotoxicity of model nanoparticles, *Chemosphere* 75 (2009) 850–857.  
424 doi:10.1016/j.chemosphere.2009.01.078.
- 425 [32] D. Stampoulis, S.K. Sinha, J.C. White, Assay-dependent phytotoxicity of  
426 nanoparticles to plants, *Environ. Sci. Technol.* 43 (2009) 9473–9479.  
427 doi:10.1021/es901695c.
- 428 [33] E. Jurado, M. Fernández-Serrano, J. Núñez-Olea, M. Lechuga, J.L Jiménez, F. Ríos,  
429 Effect of concentration on the primary and ultimate biodegradation of  
430 alkylpolyglucosides in aerobic biodegradation tests, *Water Environ. Res.* 83 (2011)  
431 154–161. doi:10.2175/106143010X12780288628336.
- 432 [34] E. Jurado, M. Fernández-Serrano, J. Núñez-Olea, M. Lechuga, F. Ríos, Ecotoxicity  
433 of anionic surfactants AKYPO®, in: J. Villacampa, C.A. Brebia (Eds.), *Ecosystems*  
434 *and Sustainable Development VIII*, WIT Trans. Ecol. Environ., WIT Press,  
435 Southampton, 2011, pp. 497–505. doi:10.2495/ECO110431.
- 436 [35] E. Jurado, M. Fernández-Serrano, M. Lechuga, F. Ríos, Environmental impact of  
437 ether carboxylic derivative surfactants, *J. Surfactants Deterg.* 15 (2012) 1–7.  
438 doi:10.1007/s11743-011-1278-z.

- 439 [36] E. Jurado, M. Fernández-Serrano, J. Núñez Olea, M. Lechuga, J.L. Jiménez, F. Ríos,  
440 Acute toxicity of alkylpolyglucosides to *Vibrio fischeri*, *Daphnia magna* and  
441 microalgae: a comparative study., *Bull. Environ. Contam. Toxicol.* 88 (2012) 290–5.  
442 doi:10.1007/s00128-011-0479-5.
- 443 [37] M. Fernández-Serrano, E. Jurado, A. Fernández-Arteaga, F. Ríos, M. Lechuga,  
444 Ecotoxicological assessment of mixtures of ether carboxylic derivative and amine-  
445 oxide-based nonionic surfactants on the aquatic environment, *J. Surfactants Deterg.*  
446 17 (2014) 1161–1168. doi:10.1007/s11743-014-1621-2.
- 447 [38] S. Parvez, C. Venkataraman, S. Mukherji, A review on advantages of implementing  
448 luminescence inhibition test (*Vibrio fischeri*) for acute toxicity prediction of  
449 chemicals., *Environ. Int.* 32 (2006) 265–8. doi:10.1016/j.envint.2005.08.022.
- 450 [39] V.L.K. Jennings, M.H. Rayner-Brandes, D.J. Bird, Assessing chemical toxicity with  
451 the bioluminescent photobacterium (*vibrio fischeri*): a comparison of three  
452 commercial systems, *Water Res.* 35 (2001) 3448–3456. doi:10.1016/S0043-  
453 1354(01)00067-7.
- 454 [40] M. J. Rosen, J. T. Kunjappu, *Surfactants and Interfacial Phenomena*, 4th ed., Wiley,  
455 New York, 2012.
- 456 [41] B. Z. Chowdhry, M. J. Snowden, C. MacLeod, S. A. Leharne, *Thermochimica Acta*,  
457 359 (2000) 29–36. doi.org/10.1016/S0040-6031(00)00482-2
- 458 [42] H. Ma, M. Luo, L.L. Dai, Influences of surfactant and nanoparticle assembly on  
459 effective interfacial tensions, *Phys. Chem. Chem. Phys.* 10 (2008) 2207–2213.  
460 doi:10.1039/b718427c.
- 461 [43] M. Zargartalebi, N. Barati, R. Kharrat, Influences of hydrophilic and hydrophobic  
462 silica nanoparticles on anionic surfactant properties: Interfacial and adsorption  
463 behaviors, *J. Petrol. Sci. Eng.* 119 (2014) 36–43. doi:10.1016/j.petrol.2014.04.010.
- 464 [44] B.P. Binks, A. Desforges, D.G. Duff, Synergistic stabilization of emulsions by a  
465 mixture of surface-active nanoparticles and surfactant, *Langmuir* 23 (2007) 1098-  
466 1106. doi:10.1021/la062510y.
- 467 [45] C. Fruijtier-Pölloth, The toxicological mode of action and the safety of synthetic  
468 amorphous silica—A nanostructured material, *Toxicology* 294 (2012) 61–79.  
469 doi:10.1016/j.tox.2012.02.001.
- 470 [46] A. Nel, T. Xia, L. Mädler, N. Li, Toxic Potential of Materials at Nanolevels, *Science*  
471 311 (2006) 622–627. doi:10.1126/science.1114397.
- 472 [47] K. Unfried, C. Albrecht, L. Klotz, A. Von Mikecz, S. Grether-Beck, R.P.F Schins,  
473 Cellular responses to nanoparticles: Target structures and mechanisms,  
474 *Nanotoxicology* 1 (2007) 52–71. doi:10.1080/00222930701314932.
- 475 [48] E.-J. Park, K. Park, Oxidative stress and pro-inflammatory responses induced by  
476 silica nanoparticles in vivo and in vitro, *Toxicol. Lett.* 184 (2009) 18–25.  
477 doi:10.1016/j.toxlet.2008.10.012.
- 478 [49] I.-Y. Kim, E. Joachim, H. Choi, K. Kim, Toxicity of silica nanoparticles depends on  
479 size, dose, and cell type, *Nanomedicine Nanotechnology, Biol. Med.* 11 (2015)  
480 1407–1416. doi:10.1016/j.nano.2015.03.004.
- 481 [50] H. Jaganathan, B. Godin, Biocompatibility assessment of Si-based nano- and micro-  
482 particles, *Adv. Drug Deliv. Rev.* (2012) 1800-1819. doi:10.1016/j.addr.2012.05.008.

- 483 [51] L.K. Adams, D.Y. Lyon, P.J.J. Alvarez, Comparative eco-toxicity of nanoscale  
484 TiO<sub>2</sub>, SiO<sub>2</sub>, and ZnO water suspensions, *Water Res.* 40 (2006) 3527–3532.  
485 doi:10.1016/j.watres.2006.08.004.
- 486 [52] D.W. Roberts, S.J. Marshall, Application of Hydrophobicity Parameters to  
487 Prediction of the Acute Aquatic Toxicity of Commercial Surfactant Mixtures, *SAR*  
488 *QSAR Environ. Res.* 4 (1995) 167–176. doi:10.1080/10629369508029914.
- 489 [53] D.W. Roberts, J.F. Costello, Mechanisms of action for general and polar narcosis: A  
490 difference in dimension, *QSAR Comb. Sci.* 22 (2003) 226-233.
- 491 [54] Scientific Committee on Health and Environmental Risk, Scientific Committee on  
492 Emerging and Newly Identified Health Risks, Scientific Committee on Consumer  
493 Safety, 2012. Toxicity and assessment of chemical mixtures. European Commission.  
494 doi:10.2772/21444.
- 495 [55] R. Altenburger, M. Nendza, G. Schüürmann, Mixture toxicity and its modeling by  
496 Quantitative Structure–Activity Relationships, *Environ. Toxicol. Chem.* 22 (2003)  
497 1900–19015. doi:10.1897/01-386.
- 498 [56] M.K. Matsson, B. Kronberg, P.M. Claesson, Adsorption of alkyl polyglucosides on  
499 the solid/water interface: Equilibrium effects of alkyl chain length and head group  
500 polymerization, *Langmuir* 20 (2004) 4051–4058. doi:10.1021/la035959p.
- 501 [57] D.M. Lugo, J. Oberdisse, A. Lapp, G.H. Findenegg, Effect of nanoparticle size on  
502 the morphology of adsorbed surfactant layers, *J. Phys. Chem. B.* 114 (2010) 4183–  
503 4191. doi:10.1021/jp911400j.
- 504 [58] S. Kumar, V.K. Aswal, J. Kohlbrecher, Size-Dependent Interaction of Silica  
505 Nanoparticles with Different Surfactants in Aqueous Solution, *Langmuir* 28 (2012)  
506 9288–9297. doi:10.1021/la3019056.
- 507 [59] B. Bharti, J. Meissner, U. Gasser, G.H. Findenegg, Surfactant adsorption and  
508 aggregate structure at silica nanoparticles: Effects of particle size and surface  
509 modification, *Soft Matter* 8 (2012) 6573–6581. doi:10.1039/c2sm25648g.
- 510 [60] D. Lugo, J. Oberdisse, M. Karg, R. Schweins, G.H. Findenegg, Surface aggregate  
511 structure of nonionic surfactants on silica nanoparticles, *Soft Matter.* 5 (2009) 2928–  
512 2936. doi: 10.1039/B903024g.
- 513 [61] G. Despert, J. Oberdisse, Formation of micelle-decorated colloidal silica by  
514 adsorption of nonionic surfactant, *Langmuir* 19 (2003) 7604–7610.  
515 doi:10.1021/la0300939.
- 516 [62] S. Kumar, V.K. Aswal, Tuning of nanoparticle-surfactant interactions in aqueous  
517 system, *J. Phys. Condens. Matter.* 23 (2011) 035101. doi:10.1088/0953-  
518 8984/23/3/035101.
- 519 [63] S. Ahualli, G.R. Iglesias, W. Wachter, M. Dulle, D. Minami, O. Glatter, Adsorption  
520 of anionic and cationic surfactants on anionic colloids: supercharging and  
521 destabilization, *Langmuir* 27 (2011) 9182–9192. doi:10.1021/la201242d.
- 522 [64] V. N. Paunov, and Bernard P. Binks, N.P. Ashby, Adsorption of charged colloid  
523 particles to charged liquid surfaces, *Langmuir* 18 (2002) 6946–6955.  
524 doi:10.1021/la0203584.
- 525