

1 **BIOCRUST COVER AND SUCCESSIONAL STAGES INFLUENCE SOIL BACTERIAL**
2 **COMPOSITION AND DIVERSITY IN SEMIARID ECOSYSTEMS**

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10 **ABSTRACT**

11 Biocrusts are an important constituent of landscape in drylands, which enrich the upper millimeters of the
12 soil with organic matter and initiate the biogeochemical cycles. However, little is known about the
13 influence of the biocrust on soil bacterial community structure and diversity. Different biocrust types
14 representing a successional gradient were studied. This gradient, from the earliest to the latest
15 successional stages, consisted in incipient cyanobacterial biocrust < mature cyanobacterial biocrusts <
16 biocrust dominated by the *Squamarina lentigera* and *Diploschistes diacapsis* lichens < Biocrust
17 characterized by the *Lepraria isidiata* lichen. Moreover, in each biocrust type, four different percentage
18 of biocrust-cover were also selected. The soil diversity gradually increased with biocrust successional
19 stage and percentage of biocrust-cover. The biocrusts-cover had an important role in the total abundance
20 of bacteria generally increasing in soils colonized by the highest percentages of cover. Biocrust
21 successional stage was the most important factor significantly influencing in 108 soil bacteria genera,
22 whereas biocrust-cover showed significant differences in only 10 genera. Principal Component Analysis
23 showed contrasting microbial composition across the biocrust successional gradient. Some bacterial taxa

24 were dominant in the soil colonized by different biocrust types. Thus, *Leptolyngbya sp.*, *Rubrobacter*,
25 *Solirubrobacter*, *Geodermatophilus*, etc., were more abundant in incipient cyanobacteria; *Nostocales*,
26 *Chroococcidiopsaceae*, *Coleofasciculaceae* etc., in soils colonized by mature cyanobacterial biocrusts;
27 *Truepera*, *Sphingobacteriaceae*, *Actinophytocola*, *Kribella*, etc., in soils colonized by *D. diacapsis* and *L.*
28 *isidiata* and *Bryobacter*, *Ohtaekwangia*, *Opitutus*, *Pedosphaeraceae*, etc., in soils colonized by *L.*
29 *isidiata*. Several soil bacteria taxa showed significant correlations ($p < 0.05$) with chemical soil properties
30 (pH, total nitrogen, total organic carbon, available phosphorous and electrical conductivity). We discuss
31 the role of biocrusts influencing these variables by their effects on chemical soil parameters, and also in
32 soil moisture and the presence of certain metabolites secreted by biocrusts, which could favor a more
33 selective environment for certain bacteria.

34 **Keywords:** microbial communities, lichen, cyanobacteria, Illumina MiSeq, chemical soil properties,
35 drylands.

36

37 1. INTRODUCTION

38 Biocrusts constitute one of the most important and extensive dryland landscapes worldwide (Belnap
39 and Eldridge, 2003), and have developed multiple essential roles in ecosystem functioning (Belnap and
40 Lange, 2003; Li, 2012; Weber et al., 2016). Study of the microbial communities associated with biocrusts
41 has recently attracted great interest, due to their possible synergistic relationships (Grube and Berg, 2009).
42 Some authors have shown that the size and activity of microbial communities in biocrusts and
43 immediately underlying soils is higher than in bare soils (Miralles et al., 2012a). The microbial species
44 composition in biocrusts encompasses widely diverse bacterial taxa (Cardinale et al. 2008; Grube et al.
45 2009; Hodkinson and Lutzoni 2009; Bates et al. 2011; Hodkinson et al. 2012; Moquin et al., 2012;
46 Aschenbrenner et al., 2016; Zhang et al., 2016; Liu et al., 2017). This great diversity in the composition of
47 microbial communities in lichen biocrusts could be merely opportunistic because of the immediately
48 surrounding soil environment (Cardinale et al., 2006), or due to the presence of certain groups of bacteria

49 associated with different types of biocrusts with species-specific patterns (Grube et al., 2009). Some
50 studies have provided data supporting that hypothesis, showing that lichen species appear to be the
51 strongest predictor of community composition, and that highly-structured lichen-associated bacterial
52 communities reflect different functional roles in symbiosis with the lichens (Grube et al., 2009; Bates et
53 al., 2011). Biocrusts composed of lichens and cyanobacteria certainly secrete a wide variety of secondary
54 metabolites, such as oxalates, and acid compounds, such as gyrophoric acid, lecanoric acid, usnic acid,
55 and so on (Cockell and Knowland, 1999; Wynn-Williams and Edwards, 2000; Dickensheets et al., 2000;
56 Edwards, 2007; Miralles et al., 2012b), which could have significantly different (acidotolerant) bacterial
57 communities than species without such compounds. On the contrary, hydrophilic parts of the lichens
58 composed of a large amount of extracellular polysaccharides, could also promote the growth of anaerobic
59 bacteria (Grube and Berg, 2009). Moreover, some of the secondary metabolites produced by lichen-
60 dominated biocrusts have antimicrobial activity (Francolini et al., 2004; Boustie and Grube, 2005;
61 Hodkinson et al. 2012; Kosanić and Ranković, 2015), providing a selective environment in which some
62 complex bacterial communities thrive (Cardinale et al., 2008; Grube et al., 2009; Grube et al., 2015).
63 Differences in chemistry, structure and growth of lichens dominating biocrusts (i.e. crustose, foliose or
64 fruticose lichens) also create a wide diversity of ecological niches for additional microorganisms (Grube
65 and Berg, 2009; Grube et al., 2009; Hodkinson et al., 2012). Therefore, the presence of selective bacteria
66 communities in biocrusts and the leaching of certain secondary metabolites from the biocrusts to the
67 immediately underlying soil could influence and select the composition of the soil microbial
68 communities. However, there are hardly any studies focused on analyzing specific microbial communities
69 of soils which could be influenced by biocrusts.

70 Biocrusts could also indirectly influence the microbial communities in the immediately underlying soil
71 due to changes in it as it is colonized by the biocrusts (Pointing and Belnap, 2012; Miralles et al., 2012a).
72 Several studies have shown that soil microbial communities are governed by physical and chemical
73 properties of soils, such as pH, soil organic carbon or amount of salts in the soils (Lauber et al., 2009;
74 Goldfarb et al., 2011; Griffiths et al., 2011; Kuramae et al., 2012; Canfora et al., 2014; Sánchez-Marañón

75 et al., 2017). As biocrusts colonize the soil surface, changes in physical, chemical and biochemical soil
76 properties occur in the underlying layers (Pointing and Belnap, 2012; Miralles et al., 2012a, c). Biocrust
77 fungal and cyanobacterial filaments stabilize the soil by penetrating the soil surface and providing particle
78 cohesion (Pointing and Belnap, 2012). They favor soil aggregation and porosity by increasing water
79 infiltration and retention (Belnap, 2006), and increase carbohydrate-C, polyphenol-C and labile-C content
80 in underlying soils (Miralles et al., 2013). They also influence the nitrogen cycle, including N fixation
81 (Elbert et al., 2012) and nitrification (Castillo-Monroy et al., 2010), drive soil hydrolytic enzyme activity
82 (Miralles et al., 2012c) and reduce bulk density by increasing porosity (Miralles et al, 2011). These
83 changes in soil properties are in turn closely related to the different biocrust types, which are often related
84 to successional stages. Thus, biocrusts release organic acid secondary metabolites in abundance (Elix and
85 Stocker-Wörgötter, 2008), and some of their polysaccharides are characteristic of certain lichen groups
86 (Carbonero et al., 2002). Late successional stages of biocrusts predominated by lichens produce higher
87 labile-C, carbohydrate and polyphenol contents than early successional biocrusts (Miralles et al., 2013),
88 and show high N-cycle enzyme activity that hydrolyzes low-molecular-weight substrates, indicating a
89 gradual replacement of atmospheric N₂-fixing organisms by other autotrophs which are not (Miralles et
90 al., 2012a). At the same time, provision of those nutrients by biocrusts as well as the differences in carbon
91 sources and N dynamics between late and early successional biocrusts, could be shaping the bacterial
92 community in the immediately underlying soil on a very fine scale. Nevertheless, very little is known
93 about the taxonomic composition of bacterial communities in the geological substrate colonized by
94 biocrusts, or which specific groups of bacteria could be associated with different types of biocrusts, and
95 even less about how different soil bacterial communities are affected by the successional stages and cover
96 of the biocrusts, or the factors that determine these changes in their soil microbial communities. We
97 hypothesized that biocrust-cover and its successional stage may affect the total abundance of soil bacteria,
98 diversity and specificity of the soil bacterial taxa associated with the each biocrust. The development of
99 the biocrust would also have a positive effect on soil microbial communities. Thus, the composition of the
100 soil microbiota would depend on the type and cover of biocrust.

101 The aim of this study was to explore the changes in bacterial communities in soils colonized by four
102 biocrust types related by their succession, each representing a further step in their natural succession, and
103 different percentages of biocrust cover in the Tabernas Desert, a typically crusted location in semiarid
104 southeast Spain. The detailed objectives were to study: (i) total abundance of soil bacteria, diversity and
105 taxonomic composition at phylum and genus levels associated with biocrust types at their different
106 successional stages, and each of them associated in turn with different biocrust covers, (ii) the
107 relationships between soil bacterial taxa associated with each biocrust type and key chemical soil
108 parameters, and (iii) dominant soil bacteria taxa associated with the different biocrust types.

109

110 **2. MATERIALS AND METHODS**

111 **2.1. Site description**

112 The study was conducted at an experimental site (El Cautivo) located in the Tabernas Desert
113 (Almería, SE Spain), a Neogene-Quaternary depression characterized as one of the most extensive
114 badlands in Spain. This basin is located in the interior of the Betic System, delimited by the Filabres
115 Range to the north, the Alhamilla Range to the southeast, the Gador Range to the southwest and the Sierra
116 Nevada Range to the west. The altitude in the study site ranges from 240 to 385 m.a.s.l. The main
117 geological materials in the basin are Neogene marine sediments, most of which are calcaric-gypsiferous
118 mudstones and calcaric sandstones. The climate is semiarid Thermo-Mediterranean with hot dry summers
119 and mild temperatures throughout the rest of the year (the mean annual temperature is 18°C, with an
120 absolute maximum of 45°C and absolute minimum of -5.5°C; Lázaro et al., 2004). The mean annual
121 rainfall is 235 mm, most of which falls in winter. Most rainfall events are of low magnitude (less than
122 10% over 20 mm), with occasional high-intensity events associated with thunderstorms, but low-intensity
123 rainfall events lasting several hours are frequent as well. The main types of soil at the study site are
124 Epileptic and Endoleptic Leptosols, Calcaric Regosols and Eutric Gypsisols, according to the World
125 Reference Base for Soil Resources (FAO-ISRIC-ISSS, 1998). Soil pH is basic with very high calcium

126 carbonate content, and the main soil texture fractions are silt followed by sand and clay (Chamizo et al.,
127 2012). The study area landscape consists of asymmetric NW–SE valleys comprising a mosaic of zones
128 with vascular plants, biocrusts and bare substrate. The NE-facing hillslopes often have two parts: Near the
129 top, with gradients of nearly 30°, they are carpeted by Endoleptic Regosols (FAO–ISRIC–ISSS, 1998) or
130 Lithic–xeric Torriorthent (Soil Survey Staff, 1999). The soils are densely covered by lichens (mostly
131 *Squamarina lentigera* (Web.) Poelt., *Diploschistes diacapsis* (Ach.) Lumbsch, *Lepraria*
132 *isidiata* (Llimona) Llimona & Crespo and *Fulgensia fulgida* (Nyl.) Szatala) and cyanobacteria. Near the
133 bottom of the NE-facing hillslope, with gentle slope gradients, the soils are relatively thick, formed by
134 Haplic Calcisols (FAO–ISRIC–ISSS, 1998) or Xeric Haplocalcid (Soil Survey Staff, 1999) and covered
135 by patches of vascular plants in which annuals and perennials are very often spatially discriminated (with
136 predominance of *Stipa capensis* Thunb., *Helianthemum almeriense* Pau, *Hammada articulate* (Moq.) O.
137 Bolós & Vigo, *Artemisia barrelieri* Besser, and *Salsola genistoides* Poiret), while biocrusts cover
138 interplant spaces. The SW-facing slopes are steeper (slope gradients up to 70°) with poorly developed
139 soils, Epileptic Regosols (FAO–ISRIC–ISSS, 1998) or Lithic Torriorthent (Soil Survey Staff, 1999), and
140 very often bare, or with very sparse individuals of very few species of perennial and annual plants (*S.*
141 *genistoides* and *Moricandia foetida* Bourgeau ex Cosson) and very occasional small biocrust patches
142 (Lázaro et al. 2000).

143

144 **2.2. Experimental design**

145 The selected study area is a representative semiarid Mediterranean ecosystem characterized by its
146 abundant biocrust coverage of over 50% of the surface. The biocrusts selected for this study were the
147 most common types representing ecological successional stages (Lázaro et al., 2008), from earliest to
148 latest, as follows: (1) Incipient cyanobacterial biocrust (Incipient, IC) < (2) Mature cyanobacterial
149 including early lichens (Cyanobacterial, MC) < (3) Biocrust dominated by the lichens *Squamarina*
150 *lentigera* and *Diploschistes diacapsis* (Squamarina-Diploschistes, SD) < (4) Biocrust characterized by the
151 *Lepraria isidiata* (Lepraria, LI) lichen. In each biocrust type, four points representing four cover levels

152 were selected along a gradient of biocrust cover. The four coverage levels were defined by the following
153 cover ranges: less than 20%, 20-50%, 50-80% and over 80%. Three replicates were taken from every
154 cover level of each biocrust, according to a factorial design totalling $4 \times 4 \times 3 = 48$ samples. The samples
155 were taken from the upper soil horizon to a depth of 3 cm; lichen thalli were extracted previously to keep
156 the soil right below them. The samples were sieved to 2 mm and part of them was reserved for chemical
157 analysis and the other part was frozen to -80°C for biological analysis.

158

159 **2.3. Chemical soil properties**

160 Basic soil chemical property analyses were performed following standard procedures. Electrical
161 conductivity (EC) of aqueous extract 1/5 (w/v) was measured using a digital conductivity meter (Basic
162 30, Crison, Carpi, Italy), and pH was determined in an aqueous solution 1/2 (w/v) in a micropH 2002
163 Crison pHmeter (Crison, Barcelona, Spain). Soil Organic Carbon (SOC) was measured via wet oxidation
164 with potassium dichromate and 590-nm absorbance determined with a spectrometer (Barahona et al,
165 2005). Total nitrogen content (TN) was determined using a Variomax CN analyzer (Elementar
166 Analysensysteme GmbH, Hanau, Germany) and assimilable phosphorus (AP) was determined by the
167 Watanabe and Olsen (1965) method.

168

169 **2.4. DNA extraction and qPCR**

170 Deoxyribonucleic acid (DNA) was extracted from 0.5 g soil aliquots using a PowerSoil[®] DNA
171 Isolation kit following the manufacturer's protocol (Mo-Bio Laboratories, Carlsbad CA). DNA size and
172 quality were verified by electrophoresis in 1% (w/v) agarose gel, and its concentration and purity were
173 estimated spectroscopically using Nanodrop 2000. Quantitative polymerase chain reaction (qPCR)
174 analysis was performed using iTaq[™] (BioRad, Los Angeles) universal SYBR[®] Green supermix (5 mM
175 dNTPs, 5 mM MgCl_2 , 1 mM SYBR[®] I and 1 U/uL hot-start Taq polymerase). qPCR parameters
176 consisted of initial denaturation at 95°C for 10 min, followed by 40 95°C cycles for 15 s and 60°C for
177 1 min. Triplicate PCR reactions (analytical replicates) were performed with a total 20- μL volume

178 containing 5 μL \times iTaq™ universal SYBR® Green supermix, 2 μL of each primer (10 μM) and 10 ng of
179 template DNA. Specific primers from conserved regions of the bacterial 16S rRNA gene (Steven et al.,
180 2014) were used for quantitative PCR analysis. Results were expressed as number of copies of bacterial
181 rRNA gene per gram of soil.

182

183 **2.5. High-throughput sequencing and bioinformatics analysis**

184 Illumina MiSeq platform (Reagent Kit v3 -2x300 cycles) was used for pair-ended sequencing of
185 amplicon libraries of 16S V4-V5 rRNA gene. The 16S rRNA data was processed with MOTHUR
186 Software v.1.39.5 (Schloss et al., 2009), following the MiSeq SOP. Identification and exclusion of
187 chimeric readings used Chimera UCHIME. Alpha-diversity was examined using operational taxonomic
188 units (OTU) defined at 3% dissimilarity with the distance-based greedy clustering algorithm (DGC)
189 implemented in VSEARCH. The rarefaction curves were calculated with increments of 100 sequences
190 computed at 97% similarity. The Good's coverage index and the number of observed OTUs (Sobs) were
191 calculated, as well as Shannon and InvSimpson diversity indices, Chao richness index and Evenness
192 Pielou (J') index. Finally, the taxonomy was estimated using the RDP Bayesian classifier with an 80%
193 homology level in the fixrank option, using the Silva v.132 database as a reference. The results are
194 expressed as relative abundance of each taxon in the sample, with respect to its total number of valid
195 sequences. Taxa with total abundance over 0.1% of the total number of sequences of all the samples, were
196 retained for statistical analysis.

197

198 **2.6. Statistical analysis**

199 The relative abundance of bacterial genera (variables) in the soil samples was analysed to assess the
200 significance of changes in the soil microbial composition along both gradients of biocrusts (cover and
201 successional). Permutational multivariate analysis of variance (PERMANOVA, Anderson, 2001) was
202 used to check the effects of two factors, biocrust successional stage and biocrust cover, and their
203 interaction. The first factor encompasses four biocrust types representing a gradient of successional

204 stages, from latest to earliest as follows: LI > SD > MC > IC. The second factor encompasses four
205 different cover levels of each biocrust type/stage. PERMANOVA analysis uses permutation tests to find
206 the P values, does not rely on the assumptions of traditional parametric ANOVA, and can handle
207 experimental designs such as the one employed here (Anderson, 2001). The similarity matrix of the
208 samples was carried out using the Bray Curtis dissimilarity index for multivariate PERMANOVAs. Two-
209 way Univariate Permutational Analysis of Variance (PERANOVA) with Euclidean distances was also
210 performed to check the effects of the both factors mentioned on each individual variable. PERMANOVA
211 and PERANOVA analyses were carried out using PERMANOVA+ for the PRIMER statistical package
212 (PRIMER-E Ltd., Plymouth Marine Laboratory, Ivybridge, UK) and R software (R Foundation for
213 Statistical Computing, Vienna, Austria). Pairwise comparisons were also performed for chemical soil
214 properties and diversity parameters.

215 A multivariate method based on Principal Component Analysis (PCA) was used to analyse
216 contrasting microbial compositions of soils under different biocrust types representing a successional
217 gradient and the relationships between the variables (abundance of bacterial genera) and the affinity of the
218 soil microbial communities with each biocrust type. The PCA displayed sample scores (soils colonized by
219 different biocrust types) in the space defined by the original variables (bacterial genus) on new axes
220 represented by principal components that explain the most variability between the variables. PCA was
221 conducted using Statgraphic Centurion XVII (Statpoint Technologies, Inc., Warrenton).

222 Box Plots were performed for the different biocrust types. The middle line in the boxes represents the
223 median and the asterisk the mean. Boxes include 50% of the data between the first and third quartiles
224 (interquartile range), and whiskers include values that deviate from the first and third quartile by a
225 distance less than 1.5 times the interquartile range. Values with a deviation over 1.5 times the interquartile
226 range are represented as circles.

227 To study the relationships between the abundance of each bacterial taxon and chemical soil
228 properties, the Pearson's correlation coefficient (r) was calculated, as well as its degree of significance
229 (p).

230 Box plot and Pearson's correlation analyses were done using R software (R Foundation for Statistical
231 Computing, Vienna, Austria).

232

233 **3. RESULTS**

234 **3.1. Chemical properties in soils colonized by different biocrusts**

235 In general, desert soils colonized by biocrusts were poor in SOC, TN and AP, with a slightly alkaline
236 pH. Biocrust types significantly influenced SOC and pH while biocrust cover significantly influenced
237 SOC, TN and EC, but the interaction between both factors did not significantly influence the chemical
238 soil parameters (Supplementary Table 1). Soils colonized by lichen-dominated biocrusts showed
239 significantly higher SOC, pH and EC than in cyanobacteria-dominated biocrusts (Table 1). The increasing
240 percentage of biocrust cover increased the SOC and TN content in LI, SD and MC. On the contrary, the
241 soils with the lowest percentage of biocrust cover showed higher EC in LI, SD and IC (Table 1).

242

243 **3.2. Bacterial abundance, alpha-diversity and richness of bacterial communities in soils colonized** 244 **by different biocrust types and covers**

245 Rarefaction curves (Supplementary Fig. 1) did not show any apparent saturation in this survey, which
246 covered 87% to 97% of within-community (alpha) diversity (Table 2). Sequencing the 16S rRNA gene
247 (V4-V5) amplicons with the Illumina MiSeq system resulted in a total of 2129955 sequences after
248 eliminating those nonaligned and chimeras with an average length of 413bp. High-quality sequences from
249 each soil sample were subsampled to 22373 sequences prior to calculating alpha-diversity parameters.
250 Biocrust types significantly influenced all diversity indices, but biocrust cover and interaction between the
251 two factors did not significantly influence the diversity indices (Supplementary Table 2). In general, the
252 pattern shown by Sobs, Chao, InvSimpson, Shannon, and Evenness (J') indices were similarly
253 significantly higher in soils colonized by LI, followed by SD, and finally MC and IC, which were not
254 significantly different from each other (Table 2). The Good's coverage index was significantly higher in

255 LI than in MC and IC, but SD was not significantly different from the rest of the biocrusts types. Neither
256 were any significant differences found between biocrust covers (Table 2). The total abundance of bacteria
257 estimated by qPCR of the 16S rRNA gene (number of copies per gram of soil) was the highest in IC
258 followed by MC, SD and LI (Figure 1). The total abundance of bacteria decreased progressively from the
259 highest percentage of biocrust cover to those with the lowest biocrust cover in LI, SD and IC.
260 Nevertheless, in the MC biocrusts, the total abundance of bacteria per gram of soil was greater at the
261 sampling points with an intermediate cover percentage (Figure 1).

262

263 **3.3. Taxonomic composition of microbial communities in soils colonized by different biocrust types** 264 **and covers**

265 Metagenome analysis revealed the same dominant phyla at different proportions in soils under
266 different biocrust types (Fig. 2). The Bayesian classifier with SILVA reference identified 29 phyla. The
267 most abundant phyla in LI were *Proteobacteria*, *Bacteroidetes* and *Planctomycetes* follow by
268 *Actinobacteria* and *Acidobacteria*, whereas in SD *Proteobacteria*, *Bacteroidetes*, *Actinobacteria* and
269 *Acidobacteria* were the most abundant. The *Planctomycetes* phylum was much less abundant in SD
270 biocrusts than in LI (Fig. 2). The abundance percentage of the *Cyanobacteria* phylum was very small in
271 soils colonized by both lichenic biocrusts (2% in LI and 6% in SD). Nevertheless, *Cyanobacteria* was the
272 most abundant phylum in MC follow by the *Bacteroidetes* and *Proteobacteria* phyla, whereas in IC, the
273 dominant phyla were *Bacteroidetes*, *Proteobacteria*, *Actinobacteria* and *Cyanobacteria*, where this last
274 phylum was much less abundant than in MC (Fig. 2).

275 Phylogenetic analysis showed a total of 136 taxa with relative abundance over 0.1% identified at the
276 lowest classification level (subgroup-to-genus). PERMANOVA results showed that differences in soil
277 bacterial communities between biocrusts types (different successional stages) were highly significant ($P <$
278 0.001), while differences between soil bacteria communities by biocrust cover or biocrust type/biocrust
279 cover interaction were not significant ($p > 0.05$; Table 3). Nevertheless, univariate PERANOVA showed
280 that both factors analyzed and their interaction, significantly influenced some soil bacteria classified at the

281 lowest classification level available (genus subgroup). However, the biocrust-successional stage was the
282 most important factor with significant influence in 108 genera of soil bacteria, whereas biocrust-cover
283 differences were only significant in 10 genera, and the interaction between both factors only in five
284 genera of soil bacteria (Supplementary Table 3).

285 The PCA biplot of the two first principal components calculated from the taxa with highest
286 significance in the univariate PERANOVA test (Supplementary Table 3) explained 57.52% of the total
287 variability, ranging from 37.46% to 20.06% between the first and second components, respectively.
288 Interestingly, the bacteria eigenvectors coincided with the biocrust types/successional stages (LI, SD, MC,
289 IC), each forming one of four independent clusters in different regions of the factor space. The horizontal
290 axis differentiates biocrust succession from the cyanobacteria, the first colonizers, followed by the better-
291 developed biocrust stages from SD to LI. The vertical axis differentiates the first step of bare soil
292 colonization by cyanobacteria from an incipient (IC) to a more developed colonization stage (MC)
293 (Figure 3).

294 The bacterial genera (variables) with the largest positive loadings on PC1 (*Bdellovibrio*, *Lineage IIb*
295 *genera*, *Ferruginibacter*, *Microscillaceae*, *Opitutus*, *Pirellula*, *Subgroup 6 unclass*, *Subgroups 6 and 7*
296 *genera*, *Pedosphaeraceae*, *TRA3-20 genera*, *Planctomycetales*, *Bryobacter*, *Azospirillaceae unculture*,
297 *Tepidisphaerales*, *Chitinophagaceae*, *Gemmata*) mostly influenced the scores of soils colonized by the
298 lichen LI at their different percentages of biocrust cover (replicates of LI1, LI2, L3 and L4 in Fig. 3). On
299 the contrary, the variables with negative loadings (*Blastocatella*, *Coleofasciculacea*, *Nostocales*
300 *uncultured*, *Chroococcidiopsis*, *Rhodocytophaga*) influenced mainly the scores of soils dominated by
301 cyanobacteria (MC1, MC2, MC3, MC4 and some replicates of IC1). The progressive increase in the
302 genera represented on the right led to a decrease in those on the left and vice versa (Supplementary Table
303 4). The cluster in an intermediate position between the MC and LI clusters, grouped mainly soil samples
304 colonized by lichens SD (samples SD1, SD2, SD3 and SD4 and some replicates of LI4). The position of
305 the SD samples, near the origin of ordinates, although expanding slightly to the right on PC1, indicates
306 that they are mainly influenced by the bacterial genera in the LI and MC clusters. Likewise, the genera

307 with loadings with opposite signs (inverse correlation) in PC2 (Fig. 3) showed differences between soils
308 with an incipient colonization of cyanobacteria (mainly replicates IC1 to IC4) and soils colonized by
309 mature cyanobacteria (MC), also indicating changes in the content of soil microbial communities in those
310 soils. *Nocardioideae*, *Beijerinckiaceae*, *Geodermatophilaceae*, *Geodermatophilus*, *Rubellimicrobium*,
311 *Rubrobacter*, *Sphingomonas* and *Blastococcus* were especially abundant in IC1, IC2, IC3, IC4 (Fig. 3).

312 In general, the box plot analysis corroborated that the soil bacterial taxa associated with each biocrust
313 type in PCA also showed higher relative abundance or were almost exclusive in the soils colonized by
314 those same biocrust types (Fig. 4). Thus, some bacteria were more abundant in LI or gradually increased
315 from the early successional (IC) to late successional stages of biocrusts (LI) (Fig. 4a). Other soil bacteria
316 were more abundant in SD (Fig. 4b) or in soils colonized by both lichen types (LI and SD; Fig. 4c), and
317 on the contrary, other bacteria were more abundant or almost exclusive in MC (Fig. 4d) or IC (Fig. 4f), or
318 decreased from the latest successional to the earliest successional development stages (Fig. 4f).

319

320 **3.4. Relationships between bacterial communities and edaphic properties in soils colonized by** 321 **different biocrust types.**

322 The soil parameters with the largest number of significant correlations ($p > 0.05$) with soil bacterial
323 abundance at their lowest taxonomic level (subgroup-to-genus) were pH, SOC, TN and EC, although
324 these correlations were not very high (r ranged from 0.30 to 0.55) (Supplementary Table 5). Soil bacteria
325 positively correlated with pH were more abundant in LI and SD, whereas the bacteria communities
326 correlated negatively with soil pH were more abundant in MC and IC. SOC and TN were positively
327 correlated with the abundance of bacteria per gram of soil, and in general, the soil bacteria positively
328 correlated with SOC and TN content were more abundant in LI, SD and MC biocrusts. On the contrary,
329 the soil bacteria communities correlated positively and negatively with EC, as well as those soil bacteria
330 communities correlated negatively with SOC and TN content, were ubiquitous in soils colonized by all
331 biocrust types. EC correlated negatively with the abundance of bacteria per gram of soil (Supplementary
332 Table 5).

333

334 4. DISCUSSION

335 4.1. Influence of the biocrust-cover and their successional stage in soil parameters, total abundance 336 of soil bacteria, diversity and soil microbial composition.

337 The results show that the diversity of soils colonized by biocrusts was parallel to the biocrust
338 successional stage. The Sobs, Chao, Shannon, InvSimpson and Evenness indices gradually increased
339 from soils colonized by cyanobacteria up to soils colonized by later successional biocrusts dominated by
340 the lichen *L. isidiata* (Table 2). Moreover, in general, the abundance of bacteria per gram of soil increased
341 in soils with more biocrust cover in all biocrust types (Fig. 1). The increase in diversity and richness in
342 soils colonized by the highest lichen covers (LI and SD) could be due, on the one hand, to higher TN, AP
343 and SOC contents in the soils colonized by these biocrust types than soils colonized by cyanobacteria
344 (MC and IC) (Table 1). The results show that SOC and TN correlated significantly positively ($p>0.05$)
345 with bacterial abundance (Supplementary Table 5). Moreover, the soils below *L. isidiata*, *D. diacapsis*
346 and *S. lentigera* lichens were also characterized by their higher biomass-C, carbohydrate and polyphenol
347 contents and enzymatic activity of microbial communities (Miralles et al., 2012a, 2012c, 2013).
348 Moreover, biocrusts colonized by lichens in the Tabernas Desert have higher soil moisture than soils
349 colonized by cyanobacteria (Chamizo et al., 2012). Therefore, that increase in nutrients and soil moisture
350 in soils colonized by lichens, especially in those where the biocrust cover is greater, could provide ideal
351 conditions for establishing greater richness and diversity of soil bacteria communities.

352 The majority phyla in soils colonized by all biocrust types were Proteobacteria, Bacteroidetes,
353 Actinobacteria and Acidobacteria, including the Planctomycetes phylum in LI and the Cyanobacteria
354 phylum in MC and IC, although in the Tabernas Desert, they showed dissimilar proportions at the
355 different successional stages of the biocrusts. The different proportions of dominant phyla could alter
356 microbial community functions in the biocrust succession process, which in turn could promote their
357 development (Liu et al., 2017). In general, these phyla have also been found to be predominant in

358 biocrusts and their underlying soils in other arid and semi-arid ecosystems in Nyngan, New South Wales,
359 Australia and the Colorado Plateau (Steven et al., 2013; Liu et al., 2017). The most abundant phylum in
360 MC biocrusts was Cyanobacteria, while this phylum was much less abundant in soils just below lichens
361 (Fig. 2). This low percentage of Cyanobacteria in soils colonized by lichens (6% and 2% in SD and LI
362 respectively) could be due to the changes in physical, chemical and biochemical soil properties during
363 succession, making new biocrusts more competitive by receiving more positive feedback from the new
364 soil properties, and eventually replacing the previous one (Lázaro et al., 2008). Consequently, the
365 percentage of cyanobacteria, which are primo-colonizers, could be progressively decreasing as the
366 successional state increases in the biocrust characterized by lichen *L. isidiata*. The Cyanobacteria phylum
367 was much more abundant in MC than in IC biocrusts, whereas, other bacterial groups belonging to the
368 Bacteroidetes, Proteobacteria and Actinobacteria phyla were the most abundant in the IC (Fig. 2).

369 In the Tabernas Desert, the successional biocrust stage exerted a stronger influence on more bacterial
370 genera than the biocrust-cover or the interaction between them (Table 2), although biocrust cover also
371 exerted a positive effect on the abundance of bacteria per gram of soil in all biocrusts types (Fig. 1). Our
372 results suggest that the effect of the biocrust type could influence the existence of a contrasting microbial
373 composition across a successional gradient from early successional stages of biocrusts colonized by
374 cyanobacteria (MC and IC; richer in the bacteria *Blastocatella*, *Nostocales-unclassified*,
375 *Chroococcidiopsaceae*, etc., Fig. 3) to soils colonized by lichens *D. diacapsis* and *S. lentigera* (SD) and
376 late-successional biocrusts characterized by *L. isidiata* (LI; richer in the bacteria *Subgroupes 6 genera*
377 *and unclassified*, *Pedosphaeraceae*, *TRA3-20 genera*, etc., Fig. 3). The SD sample cluster in the middle
378 of PC1 and close to the ordinate axis in the biplot (Fig. 3), suggests that these soils were sharing bacteria
379 with MC and LI biocrust types. Likewise, the axis aligned along PC2 in Fig. 3 also showed changes in the
380 soil microbial communities across a successional gradient from incipient colonization of cyanobacteria
381 (IC; richer in the *Nocardioides*, *Geodermatophilus*, *Rubrobacter* ... genera) to developed cyanobacteria
382 (MC; richer in *Blastocatella*, *Coleofasciculaceae*, *Chroococcidiopsaceae*, etc.). These differences in the
383 microbial community composition in soils colonized by biocrusts at different successional stages could be

384 explained by several interacting factors related to changes in soil properties due to the presence of
385 biocrusts and the ecophysiology of each biocrust type. A first explanation could be that the biocrusts can
386 substantially modify the local soil environment, affecting soil stability, nutrient content, soil texture and
387 pH (Belnap and Gardner, 1993; Belnap et al., 2003). Aggregation simultaneous with the accumulation of
388 carbon and nitrogen, begins with the higher soil moisture below the biocrusts (Chamizo et al., 2012).
389 Atmospheric nitrogen-fixing organisms increase the total nitrogen content in biocrusts, whereas the
390 excretion of organic compounds by the colonizing organisms contributes to accumulation of SOC (Mager
391 and Thomas, 2011). Soil parameters, such as pH and nutrient availability are considered the most
392 important factors driving the structure of soil bacteria communities (Lauber et al., 2009; Goldfarb et al.,
393 2011; Griffiths et al., 2011; Kuramae et al., 2012; Sánchez-Marañón et al., 2017). Our results also showed
394 significant correlations between several soil bacteria taxa and chemical soil parameters (Table 4). In our
395 study area, the soil bacteria which in general presented positive significant correlations ($p < 0.05$) with
396 SOC, TN and pH (Table 4) were also those with larger abundance of the late-successional biocrust types
397 (Fig. 4).

398 The role of late successional biocrusts dominated by lichens producing higher increases in SOC, TN,
399 aggregate stability and water retention content than the cyanobacteria-dominated biocrusts, has previously
400 been reported (Housman et al., 2006; Chamizo et al., 2012). In the Tabernas Desert, the enzymatic
401 activities involved in C, N and P cycles progressively increased from biocrusts colonized by
402 cyanobacteria to biocrusts dominated by the lichen *D. diacapsis* up to the latest successional stages in
403 biocrusts-dominated by *L. isidiata*. The hydrolytic action of these enzymes could guarantee the presence
404 of readily metabolized low-molecular-weight substrates, which could be used by heterotrophic microbial
405 communities for nutrition and protection against desiccation (Miralles et al., 2012c). Moreover, it has
406 been also shown that in the Tabernas desert, labile C (estimated from the sum of unmineralized C and
407 CO₂-C emitted), carbohydrate-C and polyphenol-C contents were significantly higher in the biocrusts-
408 dominated by *D. diacapsis*, followed by the lichen *L. isidiata* than in biocrusts dominated by
409 cyanobacteria (Miralles et al., 2013). These authors reported other differences in the type of extractable

410 carbohydrates among those biocrust types, for example, mannitol was present in lichen biocrusts
411 dominated by *D. diacapsis* and *L. isidiata* whereas its presence in cyanobacteria was minimal. These
412 differences in the organic carbon pools could favor the proliferation of different soil bacteria specialized
413 in degrading specific organic compounds (Goldfarb et al., 2011). On the other hand, the effect of the
414 biocrusts on surface wetness and its influence on selecting soil microbial communities should not be
415 ignored either. Soil moisture has a strong effect on microbial communities (Moyano et al., 2013), because
416 microorganisms are dependent on water, and depending on the species, cannot sustain their normal cell
417 activity below a certain water potential (Roe and Conrad, 2013). Some authors (Kidron et al., 2009; 2014)
418 have found high correlations between daylight surface wetness duration and the available water content in
419 the upper centimeter of soil and biocrust chlorophyll content. Lazaro et al. (2008) found that in the
420 Tabernas desert, photosynthetically active radiation and surface temperature were considerably lower in
421 all communities dominated by lichens than in communities dominated by cyanobacteria, while the soil
422 moisture and the duration of dew deposition were higher in biocrusts colonized by lichens. Thus, the
423 combined effects of this set of changes in the physico-chemical and biochemical soil properties by
424 biocrusts, along with the differences in microclimate and soil moisture among the habitats corresponding
425 to the different successional stages, may drive the differentiation of soil microbial communities. Our
426 results showed that SOC, TN and AP content were significantly influenced by the biocrust cover
427 (Supplementary Table 1), increasing in soils colonized by the biocrusts as their percentage of cover
428 increased (Table 1). Therefore, as the biocrust cover increases, it could increase their effect in underlying
429 soil chemical properties, contributing to the selection of bacteria in their underlying layers through of the
430 indirect effect of biocrusts on the soil properties.

431 Nevertheless, it was also striking that the correlations between the bacterial genera and chemical soil
432 properties were not very high (Table 1), suggesting that in the Tabernas Desert, other factors could also
433 be influencing the structure of microbial communities in soils colonized by biocrusts. Different lichens
434 and cyanobacteria colonies synthesize biochemical metabolites, such as mono and dihydrate calcium
435 oxalates (whewellite and weddellite, respectively), carotenoids, chlorophyll, parietin, emodin, atranorin,

436 gyrophoric acid, lecanoric acid, fumarprotocetraric acid, rhizocarpic acid, calycin, usnic acid, and so on,
437 with diverse ecophysiological functions (Cockell and Knowland, 1999; Holder et al., 2000, Wynn-
438 Williams and Edwards, 2000; Dickensheets et al., 2000; Edwards, 2007; Miralles et al., 2012b). The
439 synthesis of these biomolecules is species-specific, and *L. isidiata*, *D. diacapsis*, *S. lentigera* and
440 cyanobacteria biocrust types are the most important in determining the nature and proportion of those
441 biomolecules in the Tabernas Desert (Miralles et al., 2017). Other authors have shown that some of the
442 secondary metabolites secreted by lichens have antibacterial or antifungal activities (Francolini et al.,
443 2004; Boustie and Grube, 2005; Hodkinson et al., 2012; Kosanić and Ranković, 2015). The differences
444 in metabolites by biocrust type could contribute to developing selective ecological niches in which some
445 bacterial communities could develop better than others.

446

447 **4.2. Dominant soil bacterial taxa associated with different biocrust types.**

448 Biocrust type could be a strong predictor of soil bacteria community composition in the Tabernas
449 Desert, as different bacterial taxa could be associated with soils colonized by each biocrust type (Fig. 4).
450 Thus, our results show that the soil bacterial groups, such as *Pirellula*, *Gemmata*, *Gemmataceae*
451 *uncultured*, *Planctomycetales uncultured ge*, *Pir4 lineage*, *Tepidisphaerales unclassified*,
452 *Pedosphaeraceae ge*, *Microscillaceae unclassified*, *TRA3.20 ge*, etc., were much more abundant in or
453 almost exclusive to LI biocrusts (Fig. 4a). Some soil bacteria genera, such as *WD2101 ge*, *Pirellulaceae*
454 *uncultured*, *Pirellulaceae unclassified*, *0319.6G20 ge*, and *Microscillaceae* also showed a clear pattern by
455 successional state of the four biocrust types analyzed, their relative abundance increasing proportionally
456 from the earliest successional stages (IC) to the latest (LI) (Fig. 4a). These results suggest that *L. isidiata*
457 lichen could exert a positive influence on those groups of soil bacteria in the Tabernas Desert.

458 The soils under the other biocrust types also showed other much more abundant or almost exclusive
459 bacterial groups, suggesting that they could also be favored by the overlying biocrusts. Thus, the SD
460 biocrusts seemed to exert a positive influence on the genera *Truepera*, *Sphingobacteriaceae unclassified*,
461 *Devosia*, *Actinophytocola*, and *Kribella* (Fig. 4b). The soils colonized by this biocrust shared other groups

462 of bacteria with relatively similar abundance in soils colonized by *L. isidiata* (as was the case of
463 *Chthoniobacter*, *Acidobacteria Subgroup 10*, *Bryobacter*, *Ohtaekwangia*, *Opitutus*, *Burkholderiaceae*
464 *unclassified*, *Steroidobacter*, etc. Fig. 4c) or in soils colonized by MC (e.g., *Flavitalea*, *Rhodocytophaga*).
465 As mentioned above, SD biocrusts occupy an intermediate successional position between *L. isidiata* (LI)
466 and cyanobacteria (MC) (Lázaro et al., 2008). Some authors have found that some lichen-associated
467 bacterial communities are not extensions of those found in surrounding soils, as different lichen species in
468 close spatial proximity harbor dissimilar bacterial communities, suggesting that some bacterial taxa are
469 widespread in different lichen species, probably reflecting their functional role in the lichen symbiosis
470 (Bates et al., 2011).

471 In MC the most favored genera were *Nostocales uncultured ge*, *Chroococciopsaceae uncultured* and
472 *Coleofasciculaceae unclassified* and also some bacteria, such as *Blastocatella* (Fig. 4d). However, IC
473 exerted a positive influence on the cyanobacterial genus *Leptolyngbya VRUC 135* (Fig. 4e). Other authors
474 also found that filamentous cyanobacteria *Leptolyngbya sp.* are abundant in poorly developed soil crusts
475 and decrease in well-developed lichenized biocrust (Pushkareva et al., 2015). The *Rubrobacter*,
476 *Solirubrobacter*, *Geodermatophilus*, *Longimicrobium*, *Geodermatophilaceae unclassified*, and *Nitrospira*
477 bacterial groups also showed a clear pattern by successional stage of the four biocrust types (Fig. 4f).
478 Those bacteria decreased in relative abundance from IC to LI, suggesting they are primo-colonizing taxa
479 which are outcompeted when improved soil conditions allow the development of other bacterial groups.
480 In the Tabernas Desert, biocrusts dominated by cyanobacteria are spatially distributed in the areas most
481 exposed to the sun and receiving the highest insolation (Miralles et al., 2012b). Therefore, the stress
482 caused by the higher temperatures, incident solar radiation, and long drought periods seems to be an
483 important factor explaining the occurrence of bacteria resistant to stress, such as the cyanobacteria
484 *Leptolyngbya VRUC 135* and *Rubrobacter* and *Geodermatophilus* bacteria in the incipient colonization of
485 cyanobacteria. These species have been described as thermophilic, being highly resistant to solar radiation
486 (Nakamori et al., 2014; Ferreira et al., 1999; Montero-Calasanz et al., 2014). Their resistance to stress
487 could offer them an advantage, because it would allow them to occupy the most unfavorable places,

488 where they have little competition with other bacteria that require more humid shady slopes for their
489 development.

490 Our results have shown that some soil bacteria could be favored by different biocrust types (at their
491 different successional stages), and similarly, the abundance of soil bacteria could be favored by biocrust
492 cover, possibly due to changes in the chemical soil properties and local environment fostered in large part
493 by the biocrust itself and its micro-climatic conditions, such as exposure to solar radiation, in which that
494 biocrust is successful. An additional explanation for the association between bacterial communities and
495 biocrust types could be that bacteria exert a functional role in symbiosis with biocrust organisms.
496 However, the functional roles of bacteria associated with biocrusts remain largely undetermined, so a
497 further study of the genes involved in the metabolism of biogeochemical cycles and their possible
498 connection with biocrusts would be needed.

499

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509

510 **Figure captions:**

511 -Figure 1. Total abundance (bacteria g⁻¹ soil) of bacteria estimated by qPCR of 16S rRNA in different
512 biocrust types and biocrust covers from 1, representing the maximum percentage of biocrust cover, to 4,
513 the lowest.

514 -Figure 2. Phylogenetic community composition. Relative abundance of dominant bacteria (>0.1% of
515 total reading) at the phylum level in soils colonized by different biocrust types.

516 -Figure 3. Principal component analysis. Biplots for the relative abundance of bacterial communities at
517 the lowest classification level (subgroup-to-genus) in soils colonized by four different biocrust types
518 representing a successional gradient (LI- lichen *L. isidiata*; SD- lichens *D. diacapsis* and *S. lentigera*;
519 MC-mature cyanobacteria; IC- incipient colonization of cyanobacteria), and each of them in turn
520 associated with different degrees of biocrust cover (from 1, representing the maximum percentage of
521 biocrust cover, to 4, the lowest; i.e.: LI1 indicates the soil sample colonized by the *L. isidiata* lichen with
522 the maximum percentage of biocrust cover).

523 -Figure 4. Bacterial taxonomy at the lowest classification level (subgroup-to-genus) associated with soils
524 colonized by four different biocrust types representing a successional gradient (n=12). Fig. 4a: soil
525 bacteria more abundant in soil colonized by *L. isidiata* lichen dominating biocrusts (LI); fig. 4b: soil
526 bacteria more abundant in soil colonized by *D. diacapsis* and *S. lentigera* lichens dominating biocrusts
527 (SD); fig. 4c: soil bacteria more abundant in soil colonized by both LI and SD; fig. 4d: soil bacteria more
528 abundant in soil colonized by mature cyanobacteria (MC); fig. 4f: soil bacteria more abundant in soil
529 colonized by incipient cyanobacteria (IC).

530 -Supplementary Fig. 1. Rarefaction curves representing the numbers of operational taxonomic units
531 (OTUs) versus the number of valid readings in soils colonized by four different biocrust types
532 representing a successional gradient (LI- lichen *L. isidiata*; SD- lichens *D. diacapsis* and *S. lentigera*;
533 MC-mature cyanobacteria; IC- incipient colonization of cyanobacteria), and each of them associated in
534 turn with different degrees of biocrust cover (from 1, representing the maximum percentage of biocrust
535 cover, to 4, representing the lowest; e.g., LI1 indicates the soil sample colonized by the *L. isidiata* lichen
536 with the maximum percentage of biocrust cover).

537

538 **REFERENCES**

539

540 Abed, R.M.M., Kharusi, S.A., Schramm, A., Robinson, M.D., 2010. Bacterial diversity,
541 pigments and nitrogen fixation of biological desert crusts from the Sultanate of Oman. *FEMS*
542 *Microbiology Ecology* 72,418-428.

543

544 Abed, R.M.M., Al-Sadi, A.M., Al-Shehi, M., Al-Hinai, S., Robinson, M.D., 2013. Diversity of
545 free-living and lichenized fungal communities in biological soil crusts of the Sultanate of Oman
546 and their role in improving soil properties. *Soil Biology and Biochemistry* 57, 695-705.

547

548 Anderson, M.J., 2001. A new method for non-parametric multivariate analysis of variance.
549 *Austral Ecology* 26, 32-46.

550

551 Aschenbrenner, I.A., Cernava, T., Berg, G., Grube, M., 2016. Understanding Microbial Multi-
552 Species Symbioses. *Frontiers in Microbiology* 7,180.

553

554 Barahona, E., Fernández, J., Mingorance, M.D., 2005. Determinación rápida de carbono
555 orgánico en suelos por oxidación vía húmeda. In: Jimenez, R., Álvarez, A.M. (Eds.), *Control de*
556 *la degradación de suelos. Simposio Nacional sobre Control de la Degradación de Suelos. Vol. 1.,*
557 *pp. 737-741.*

558

559 Bates, S.T., Cropsey, G.W.G., Caporaso, J.G., Knight, R., Fierer, N., 2011. Bacterial
560 communities associated with the lichen symbiosis. *Applied and Environmental Microbiology* 77,
561 1309-1314.

562

563 Belnap, J., Gardner, J.S., 1993. Soil microstructure in soils of the Colorado Plateau: the role of
564 the cyanobacterium *Microcoleus vaginatus*. *Great Basin Naturalist* 53, 40-47.

565

566 Belnap, J., Lange, O.L., 2003. *Biological Soil Crusts: Structure, Function, and Management*.
567 Springer - Verlag, Berlin.

568

569 Belnap, J., 2006. The potential roles of biological soil crusts in dryland hydrologic cycles.
570 *Hydrological Processes* 20, 3159-3178.

571

572 Boustie, J., Grube, M., 2005. Lichens-a promising source of bioactive secondary metabolites.
573 *Plant Genetic Resources* 3, 273-278.

574

575 Canfora, L., Bacci, G., Pinzari, F., Lo Papa, G., Dazzi, C, Benedetti, A., 2014. Salinity and
576 Bacterial Diversity: To What Extent Does the Concentration of Salt Affect the Bacterial
577 Community in a Saline Soil?. *PLoS ONE* 9(9): e106662.

578

579 Carbonero, E.R., Montai, A.V., Woranovicz-Barreira, S.M., Gorin, P.A.J., Iacomini, M., 2002.
580 Polysaccharides of lichenized fungi of three *Cladina* spp.: significance as chemotypes.
581 *Phytochemistry* 61, 681-686.

582

583 Cardinale, M., Puglia, A. M., Grube, M., 2006. Molecular analysis of lichen-associated bacterial
584 communities. *FEMS Microbiology Ecology* 57, 484-495.

585

586 Cardinale, M., Vieira de Castro, J., Müller, H., Berg, G., Grube, M., 2008. In situ analysis of the
587 bacterial community associated with the reindeer lichen *Cladonia arbuscula* reveals
588 predominance of Alphaproteobacteria. *FEMS Microbiology Ecology* 66, 63-71.

589

590 Castillo-Monroy, A.P., Maestre, F.T., Delgado-Baquerizo, M., Gallardo, A., 2010. Biological
591 soil crusts modulate nitrogen availability in semi-arid ecosystems: insights from a Mediterranean
592 grassland. *Plant and Soil* 333, 21-34.

593

594 Chamizo, S., Cantón, Y., Miralles, I., Domingo, F., 2012. Biological soil crust development
595 affects physicochemical characteristics of soil surface in semiarid ecosystems. *Soil Biology and*
596 *Biochemistry* 49, 96-105.

597

598 Cockell, C.S., Knowland, J., 1999. Ultraviolet radiation screening compounds. *Biological*
599 *reviews of the Cambridge Philosophical Society* 74, 311-345.

600

601 Dickensheets, D.L., Wynn-Williams, D.D., Edwards, H.G.M., Crowder, C., Newton, E.M., 2000.
602 A novel miniature confocal microscope/Raman spectrometer system for biomolecular analysis
603 on future Mars missions after Antarctic trials. *Journal of Raman Spectroscopy* 31, 633-635.

604

605 Edwards, H.G.M., 2007. A novel extremophile strategy studied by Raman spectroscopy.
606 *Spectrochimica Acta Part A: Molecular and Biomolecular Spectroscopy* 68, 1126-1132.
607

608 Elbert, W., Weber, B., Burrows, S., Steinkamp, J., 2012. Contribution of cryptogamic covers to
609 the global cycles of carbon and nitrogen. *Nature Geoscience* 5, 1-4.
610

611 Elix, J.A., Stocker-Wörgötter, E., 2008. Biochemistry and secondary metabolites. In: T.H. Nash
612 III (Ed.), *Lichen Biology* 2nd Edn. Cambridge University Press, Vol. 1., Chapter 7, pp. 104-133.
613

614 FAO–ISRIC–ISSS. 1998. World reference base for soil resources. FAO, Rome.
615

616 Ferreira, A.C., Nobre, M.F., Moore, E., Rainey, F.A., Battista, J.R., Da Costa, M.S., 1999.
617 Characterization and radiation resistance of new isolates of *Rubrobacter radiotolerans* and
618 *Rubrobacter xylanophilus*. *Extremophiles* 3, 235-238.
619

620 Francolini, I., Norris, P., Piozzi, A., Donelli, G., Stoodley, P., 2004. Usnic Acid, a natural
621 antimicrobial agent able to inhibit bacterial biofilm formation on polymer surfaces.
622 *Antimicrobial Agents and Chemotherapy* 48, 4360-4365.
623

624 Goldfarb, K.C., Karaoz, U., Hanson, C.A., Santee, C.A., Bradford, M.A., Treseder, K.K.,
625 Wallestein, M.D., Brodie, E.L., 2011. Differential Growth Responses of Soil Bacterial Taxa to
626 Carbon Substrates of Varying Chemical Recalcitrance. *Frontiers in Microbiology*, 2: 94.
627

628 Griffiths, R.I., Thomson, B.C., James, P., Bell, T., Bailey, M., Whiteley, A.S., 2011. The
629 bacterial biogeography of British soils. *Environmental Microbiology* 13, 1642-1654.

630

631 Grube, M., Cardinale, M., de Castro, J.V. Jr, Müller, H., Berg, G., 2009. Species-specific
632 structural and functional diversity of bacterial communities in lichen symbioses. *The ISME*
633 *Journal* 3, 1105-1115.

634

635 Grube, M., Berg, G., 2009. Microbial consortia of bacteria and fungi with focus on the lichen
636 symbiosis. *Fungal Biology Reviews* 23, 72-85.

637

638 Grube, M., Cernava, T., Soh, J., Fuchs, S., Aschenbrenner, I., Lassek, C., Wegner, U., Becher,
639 D., Riedel, K., Sensen, C.W., Berg, G., 2015. Exploring functional contexts of symbiotic sustain
640 within lichen-associated bacteria by comparative omics. *The ISME Journal* 9, 412-424.

641

642 Hodkinson, B.P., Lutzoni, F., 2009. Amicrobiotic survey of lichen-associated bacteria reveals a
643 new lineage from the Rhizobiales. *Symbiosis* 49, 163-180.

644

645 Hodkinson, B.P., Gottel, N.R., Schadt, C.W., Lutzoni, F., 2012. Photoautotrophic symbiont and
646 geography are major factors affecting highly structured and diverse bacterial communities in the
647 lichen microbiome. *Environmental Microbiology* 4, 147-161.

648

649 Holder, J.M., Wynn-Williams, D.D., Rull-Perez, F., Edwards, H.G.M., 2000. Raman
650 spectroscopy of pigments and oxalates in situ within epilithic lichens: *Acarospora* from the
651 Antarctic and Mediterranean. *New Phytologist* 145, 271-280.

652

653 Housman, D.C., Powers, H.H., Collins, A.D., Belnap, J., 2006. Carbon and nitrogen fixation
654 differ between successional stages of biological soil crusts in the Colorado Plateau and
655 Chihuahuan Desert. *Journal of Arid Environments* 66, 620-634.

656

657 Jorge-Villar, S.E., Edwards, H.G.M., Seaward, M.R.D., 2005. Raman spectroscopy of hot desert,
658 high altitude epilithic lichens. *Analyst* 130, 730-737.

659

660 Jorge-Villar, S.E., Edwards, H.G.M., 2010. Lichen colonization of an active volcanic
661 environment: a Raman spectroscopic study of extremophile biomolecular protective strategies.
662 *Journal of Raman Spectroscopy* 41, 63-67.

663

664 Kidron, G.J., Vonshak, A., Abeliovich, A., 2009. Microbiotic crusts as biomarkers for surface
665 stability and wetness duration in the Negev Desert. *Earth Surface Processes and Landforms* 34,
666 1594-1604.

667

668 Kidron, G.J., Benenson, I., 2014. Biocrusts serve as biomarkers for the upper 30 cm soil water
669 content. *Journal of Hydrology* 509, 398-405.

670

671 Kosanić, M., Ranković, B., 2015. Lichen secondary metabolites as potential antibiotic agents.
672 In: Ranković B. (Ed.), Lichen Secondary Metabolites. Berlin: Springer, pp. 81-104.
673

674 Kuramae, E.E., Yergeau, E., Wong, L.C., Pijl, A.S., van Veen, J.A., Kowalchuk, G.A., 2012.
675 Soil characteristics more strongly influence soil bacterial communities than land-use type. *FEMS*
676 *Microbiology Ecology* 79,12-24.
677

678 Lauber, C.L., Hamady, M., Knight, R., Fierer, N., 2009. Pyrosequencing-Based Assessment of
679 Soil pH as a Predictor of Soil Bacterial Community Structure at the Continental Scale. *Applied*
680 *and Environmental Microbiology* 75, 5111-5120.
681

682 Lázaro, R., Alexander, R.W., Puigdefábregas, J., 2000. Cover distribution patterns of lichens,
683 annuals and shrubs in the Tabernas Desert, Almería, España. In: Alexander, R.W., Millington,
684 A.C. (Eds.), *Vegetation Mapping: From Patch to Planet*. Wiley, Chichester, pp. 19-39.
685

686 Lázaro, R., Rodríguez-Tamayo, M.L., Ordiales, R., Puigdefábregas, J., 2004. El Clima. In: Mota,
687 J., Cabello, J., Cerrillo, M.I., Rodríguez-Tamayo, M.L. (Eds.), *Subdesiertos de Almería:*
688 *naturaleza de cine*. Consejería de Medio Ambiente, Junta de Andalucía. Almería, pp. 63-79.
689

690 Lázaro, R., Cantón, Y., Solé-Benet, A., Bevan, J., Alexander, R., Sancho, L.G., Puigdefábregas,
691 J., 2008. The influence of competition between lichen colonization and erosion on the evolution
692 of soil surfaces in the Tabernas badlands (SE Spain) and its landscape effects. *Geomorphology*
693 102, 252-266.

694

695 Li, X.R., 2012. Eco-hydrology of Biological Soil Crusts in Desert Regions of China. China
696 Higher Education Press, Beijing.

697

698 Liu, L., Liu, Y., Hui, R., Xie, M., 2017. Recovery of microbial community structure of biological
699 soil crusts in successional stages of Shapotou desert revegetation, northwest China. *Soil Biology
700 and Biochemistry* 107, 125-128.

701

702 Liu Y.-R., Delgado-Baquerizo, M., Trivedi, P., He, J.-Z., Wang, J.-T., Singh, B.K., 2017.
703 Identity of biocrust species and microbial communities drive the response of soil
704 multifunctionality to simulated global change. *Soil Biology and Biochemistry* 107, 208-217.

705

706 Mager, D.M., Thomas, A.D., 2011. Extracellular polysaccharides from cyanobacterial soil
707 crusts: a review of their role in dryland soil processes. *Journal of Arid Environments* 75, 91-97.

708

709 Miralles, I., Cantón, Y., Solé-Benet, A., 2011. Two-Dimensional Porosity of Crusted Silty Soils:
710 Indicators of Soil Quality in Semiarid Rangelands?. *Soil Science Society of America Journal* 75,
711 1289-1301.

712

713 Miralles, I., Domingo, F., García-Campos, E., Trasar-Cepeda, C., Leirós, M.C., Gil-Sotres, F.,
714 2012a. Biological and microbial activity in biological soil crusts from the Tabernas desert, a sub-
715 arid zone in SE Spain. *Soil Biology and Biochemistry* 55, 113-121.

716

717 Miralles, I., Jorge-Villar, S.E., Cantón, Y., Domingo, F., 2012b. Using a mini-Raman
718 spectrometer to monitor the adaptive strategies of extremophile colonizers in arid deserts:
719 relationships between signal strength, adaptive strategies, solar radiation, and humidity.
720 *Astrobiology* 12, 743-753.

721

722 Miralles, I., Domingo, F., Cantón, Y., Trasar-Cepeda, C., Leirós, M.C., Gil-Sotres, F., 2012c.
723 Hydrolase enzyme activities in a successional gradient of biological soil crusts in arid and
724 semiarid zones. *Soil Biology and Biochemistry* 53, 124-132.

725

726 Miralles, I., Trasar-Cepeda, C., Leirós, M.C., Gil-Sotres, F., 2013. Labile carbon in biological
727 soil crusts of the Tabernas desert, SE Spain. *Soil Biology and Biochemistry* 58, 1-8.

728

729 Miralles, I., Jorge-Villar, S.E., van Wesemael, B., Lázaro, R., 2017. Raman spectroscopy
730 detection of biomolecules in biocrusts from differing environmental conditions. *Spectrochimica*
731 *Acta Part A: Molecular and Biomolecular Spectroscopy* 171, 40-51.

732

733 Moquin, S.A., Garcia, J.R., Brantley, S.L., Takacs-Vesbach, C.D., Shepherd, U.L., 2012.
734 Bacterial diversity of bryophyte dominant biological soil crusts and associated mites. *Journal of*
735 *Arid Environments* 87, 110-117.

736

737 Montero-Calasanz, M.C., Hezbri, K., Göker, M., Sghaier, H., Rohde, M., Spröer, C., Schumann,
738 P., Klenk, H-P., 2014. Description of gamma radiation-resistant *Geodermatophilus dictyosporus*

739 *sp. nov.* to accommodate the not validly named *Geodermatophilus obscurus* subsp. *dictyosporus*
740 (Luedemann, 1968). *Extremophiles* 3, 235-8.
741
742 Moyano, F.E., Manzoni, S., Chenu, C., 2013. Responses of soil heterotrophic respiration to
743 moisture availability: An exploration of processes and models. *Soil Biology and Biochemistry*
744 59, 72-85.
745
746 Nakamori, H., Yatabe, T., Yoon, K.S., Ogo, S., 2014. Purification and characterization of an
747 oxygen-evolving photosystem II from *Leptolyngbya sp.* strain O-77. *Journal of Bioscience and*
748 *Bioengineering* 118, 119-24.
749
750 Pointing, S.B., Belnap, J., 2012. Microbial colonization and controls in dryland systems. *Nature*
751 *Reviews Microbiology* 10, 551-562.
752
753 Pushkareva, E., Pessi, I.S., Wilmotte, A., Elster, J., 2015. Cyanobacterial community
754 composition in Arctic soil crusts at different stages of development. *FEMS Microbiology*
755 *Ecology* 91, 1-10.
756
757 Sánchez-Marañón, M., Miralles, I., Aguirre-Garrido, J.F., Anguita-Maeso, M., Millán, V.,
758 Ortega, R., García-Salcedo, J.A., Martínez-Abarca, F., Soriano, M., 2017. Changes in the soil
759 bacterial community along a pedogenic gradient. *Scientific Reports* 7, 14593.
760

761 Schloss, P.D., Westcott, S.L., Ryabin, T., Hall, J.R., Hartmann, M., Hollister, E.B., Lesniewski,
762 R.A., Oakley, B.B., Parks, D.H., Robinson, C.J., Sahl, J.W., Stres, B., Thallinger, G.G., Van
763 Horn, D.J., Weber, C.F., 2009. Introducing mothur: Open-Source, Platform-Independent,
764 Community-Supported Software for Describing and Comparing Microbial
765 Communities. *Applied and Environmental Microbiology* 75, 7537-7541.

766

767 Soil Survey Staff., 1999. Soil taxonomy: A basic system of soil classification for making and
768 interpreting soil surveys. Agric. Handb., No. 436. U.S. Dep. of Agric., Washington, DC.

769

770 Steven, B., Gallegos-Graves, LaV., Belnap J., Kuske, C.R., 2013. Dryland soil microbial
771 communities display spatial biogeographic patterns associated with soil depth and soil parent
772 material. *FEMS Microbiology Ecology* 86, 101-113.

773

774 Steven, B., Gallegos-Graves, L.V., Yeager, C., Belnap, J., Kuske, C.R., 2014. Common and
775 distinguishing features of the bacterial and fungal communities in biological soil crusts and shrub
776 root zone soils. *Soil Biology and Biochemistry* 69, 302-312.

777

778 Watanabe, F.S., Olsen, S.R., 1965. Test of an Ascorbic Acid Method for Determining
779 Phosphorus in Water and NaHCO₃ Extracts from the Soil. *Soil Science Society of America*
780 *Journal* 29, 677-678.

781

782 Weber, B., Budel, B., Belnap, J., 2016. *Biological Soil Crusts: An Organising Principle in*
783 *Drylands Ecological Studies* 226. Springer, New York.

784

785 Wynn-Williams, D.D., Edwards, H.G.M., 2000. Proximal analysis of regolith habitats and
786 protective biomolecules in situ by laser Raman spectroscopy: overview of terrestrial Antarctic
787 habitats and Mars analogs. *Icarus* 144, 486–503.

788

789 Zhang, B., Kong, W., Wu, N., Zhang, Y., 2016. Bacterial diversity and community along the
790 succession of biological soil crusts in the Gurbantunggut Desert, Northern China. *Journal of*
791 *Basic Microbiology* 56, 670-679.

Table 1. Chemical properties of soil samples colonized by different biocrust types and covers (average \pm standard derivation).

Biocrust type (A)	Biocrusts covers (a)	SOC (g kg ⁻¹)	NT (g kg ⁻¹)	AP (ppm)	EC (mS/cm)	pH
LI	1	20.50 \pm 15.27Aa	1.23 \pm 0.23Aa	0.18Aa	484 \pm 324Aa	8.02 \pm 0.47Aa
LI	2	12.20Aabc	0.94 \pm 0.08Aa	0.36Aa	224Aab	8.21 \pm 0.38Aa
LI	3	7.87 \pm 0.40Abcd	0.96 \pm 0.04Ab	0.02Aa	579 \pm 505Aa	8.23 \pm 0.29Aa
LI	4	10.13 \pm 4.96Abc	0.89 \pm 0.13Ab	0.11Aa	875Aac	8.34 \pm 0.34Aa
SD	1	19.47 \pm 1.99Aa	1.15 \pm 0.15Aa	0.96Aa	460 \pm 166Aa	8.61 \pm 0.36Aa
SD	2	12.42 \pm 4.13Aabc	1.13 \pm 0.13Aa	0.42Aa	307 \pm 18Aab	8.15 \pm 0.47Aa
SD	3	6.84 \pm 2.11Abcd	0.93 \pm 0.03Ab	0.03Aa	365 \pm 59Aa	8.24 \pm 0.10Aa
SD	4	8.08 \pm 2.55Abc	0.83 \pm 0.07Ab	0.30Aa	678Aac	8.13 \pm 0.30Aa
MC	1	14.05 \pm 3.89ABa	1.13 \pm 0.02Aa	0.14Aa	378 \pm 86Aa	8.02 \pm 0.02Ba
MC	2	0.93ABabc	1.04 \pm 0.02Aa	0.14Aa	316 \pm 30Aab	7.89 \pm 0.08Ba
MC	3	2.40ABbcd	0.96 \pm 0.06Ab	4.35Aa	350 \pm 22Aa	7.94 \pm 0.38Ba

MC	4			1.47 ±	352 ±	
		5.34 ± 1.55ABbc	0.81 ± 0.19Ab	0.23Aa	106Aac	7.84 ± 0.10Ba
IC	1			1.45 ±		8.05 ±
		8.64 ± 2.69Ba	0.95 ± 0.16Aa	0.27Aa	254 ± 80Aa	0.08ABa
IC	2			1.64 ±		8.12 ±
		9.12 ± 0.87Babc	1.04 ± 0.10Aa	0.21Aa	281 ± 38Aab	0.42ABa
IC	3			1.33 ±		8.01 ±
		6.10 ± 0.51Bbcd	0.81 ± 0.11Ab	0.10Aa	324 ± 214Aa	0.53ABa
IC	4			2.12 ±	633 ±	8.00 ±
		8.66 ± 4.32Bbc	0.80 ± 0.16Ab	0.29Aa	475Aac	0.16ABa

LI: lichen *Lepraria isidiata*; SD: Lichens *Squamarina lentigera* and *Diploschistes diacapsis*; MC: Mature Cyanobacteria; IC: Incipient Cyanobacteria; SOC: Soil Organic Carbon; NT: Total Nitrogen; AP: Available Phosphorous; EC: Electrical Conductivity. Biocrusts covers are ranged from 1, representing the maximum percentage of biocrust cover, to 4, the lowest. Different capital letters denoting significant differences between biocrust types. Different lowercase letters denoting significant differences between biocrust covers.

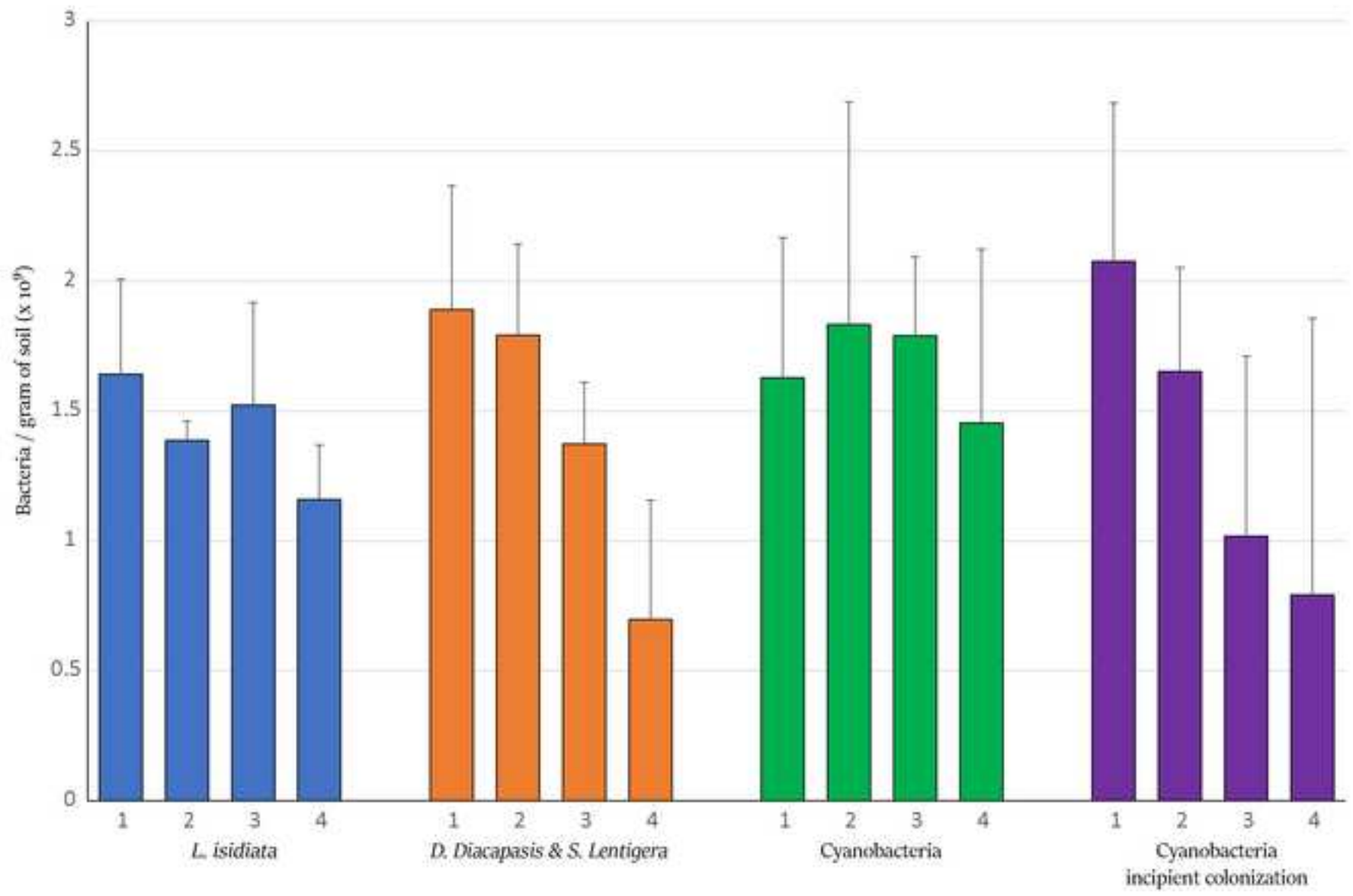
Table 2. Diversity indices for 16S rDNA sequences in soil samples colonized by different biocrust types and covers (average \pm standard derivation).

Biocrust types	Biocrusts covers	Coverage	Sobs	Chao	InvSimpson	Shannon	Evenness Pielou (J')
LI	1	0.90 \pm 0.02a	3876 \pm 534a	7862 \pm 1369a	201 \pm 30.2a	6.60 \pm 0.22a	0.80 \pm 0.01a
LI	2	0.89 \pm 0.00a	3972 \pm 79a	8534 \pm 502a	221 \pm 24.6a	6.68 \pm 0.03a	0.81 \pm 0.00a
LI	3	0.89 \pm 0.01a	4069 \pm 372a	8928 \pm 792a	239 \pm 30.0a	6.73 \pm 0.16a	0.81 \pm 0.01a
LI	4	0.91 \pm 0.01a	3725 \pm 244a	7264 \pm 1027a	188 \pm 26.3a	6.55 \pm 0.07a	0.80 \pm 0.01a
SD	1	0.90 \pm 0.01abc	3716 \pm 220b	8395 \pm 968ab	202 \pm 37.9b	6.58 \pm 0.03b	0.80 \pm 0.00b
SD	2	0.91 \pm 0.00abc	3473 \pm 99b	7040 \pm 326ab	152 \pm 47.7b	6.41 \pm 0.14b	0.79 \pm 0.02b
SD	3	0.90 \pm 0.00abc	3733 \pm 24b	8140 \pm 572ab	172 \pm 26.0b	6.54 \pm 0.07b	0.79 \pm 0.01b
SD	4	0.91 \pm 0.02abc	3379 \pm 468b	7540 \pm 1456ab	134 \pm 98.8b	6.14 \pm 0.42b	0.76 \pm 0.04b
MC	1	0.92 \pm 0.01bcd	3079 \pm 469c	7257 \pm 1171b	109 \pm 61.4c	5.94 \pm 0.47c	0.74 \pm 0.05c
MC	2	0.91 \pm 0.01bcd	3193 \pm 201c	7447 \pm 1174b	81 \pm 39.7c	5.92 \pm 0.28c	0.73 \pm 0.03c
MC	3	0.92 \pm 0.01bcd	3142 \pm 213c	6852 \pm 478b	81 \pm 27.9c	5.97 \pm 0.16c	0.74 \pm 0.02c
MC	4	0.92 \pm 0.01bcd	2907 \pm 164c	6623 \pm 597b	50 \pm 7.1c	5.74 \pm 0.03c	0.72 \pm 0.00c
IC	1	0.92 \pm 0.00bc	2894 \pm 133c	6513 \pm 473b	73 \pm 30.4c	5.81 \pm 0.23c	0.73 \pm 0.02c
IC	2	0.91 \pm 0.01bc	3304 \pm 403c	7302 \pm 1014b	104 \pm 78.2c	6.05 \pm 0.47c	0.75 \pm 0.05c
IC	3	0.91 \pm 0.02bc	3318 \pm 430c	7626 \pm 1938b	130 \pm 39.2c	6.18 \pm 0.23c	0.76 \pm 0.02c
IC	4	0.92 \pm 0.02bc	2900 \pm 657c	6175 \pm 1421b	131 \pm 47.0c	6.06 \pm 0.26c	0.76 \pm 0.02c

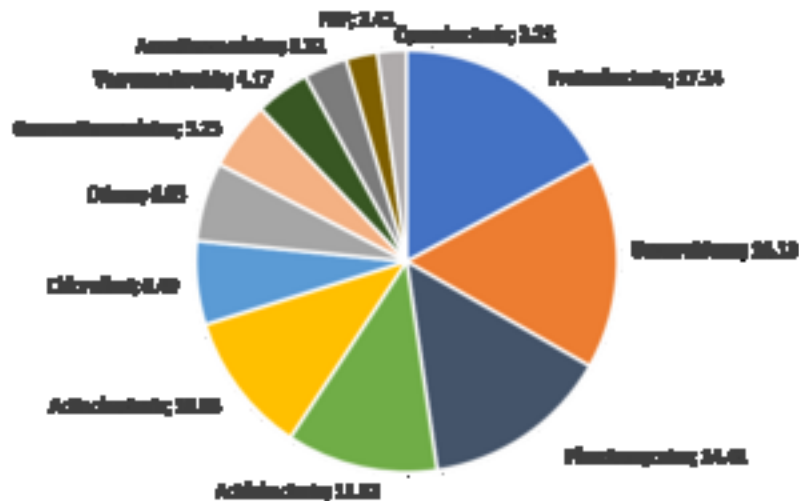
LI: lichen *Lepraria isidiata*; SD: Lichens *Squamarina lentigera* and *Diploschistes diacapsis*; MC: Mature Cyanobacteria; IC: Incipient Cyanobacteria. Biocrusts covers are ranked from 1, representing the maximum percentage of biocrust cover, to 4, the lowest. Lower letters denote significant differences among soils colonized by different biocrust types. No significant differences were found among biocrusts covers.

Table 3. Significant differences in soil bacterial communities found by PERMANOVA analysis, by biocrust type, biocrust cover and interaction of both.

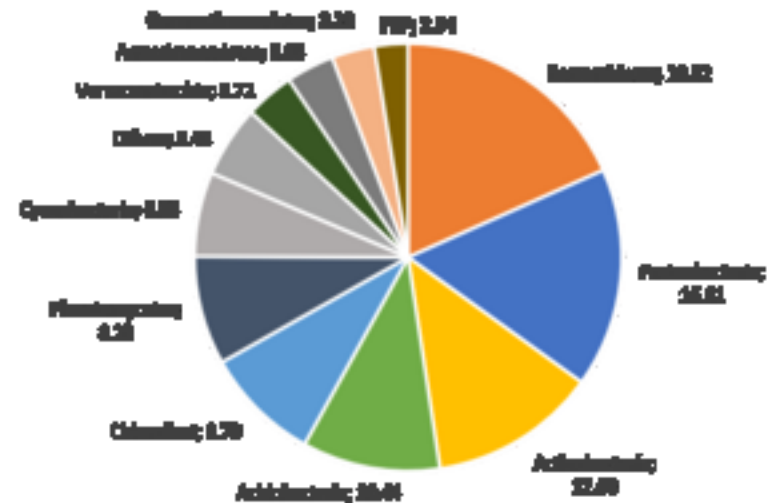
	df	SS	MS	Pseudo-F	P(perm)	perms
Biocrusts types	3	13501	4500.2	8.645	0.0001	9901
Biocrusts covers	3	1606	535.3	1.028	0.4127	9873
Biocrusts types x Biocrusts covers	9	3778	419.7	0.806	0.9036	9826
Res	32	16658				
Total	47	35542				



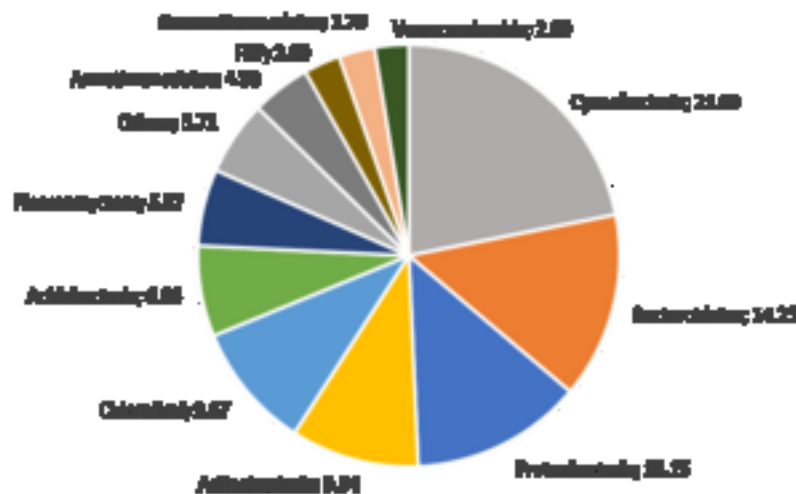
Phyla in biocrusts dominated by *Lepraria liliifera* lichen (3%)



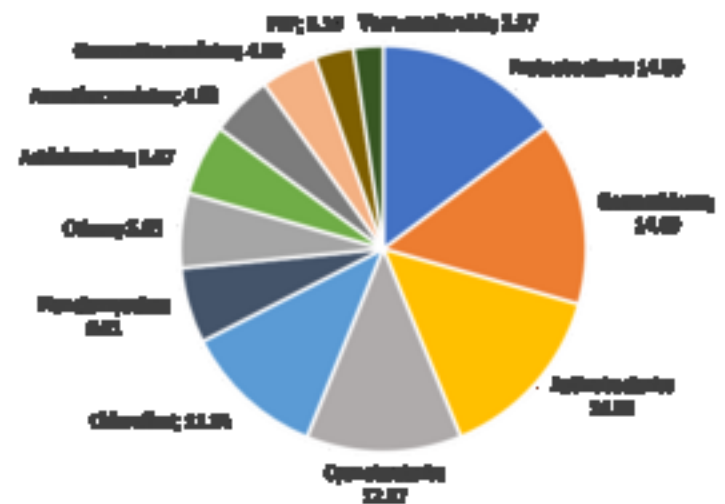
Phyla in biocrusts dominated by *D. discopula* and *X. fastigata* lichens (3%)



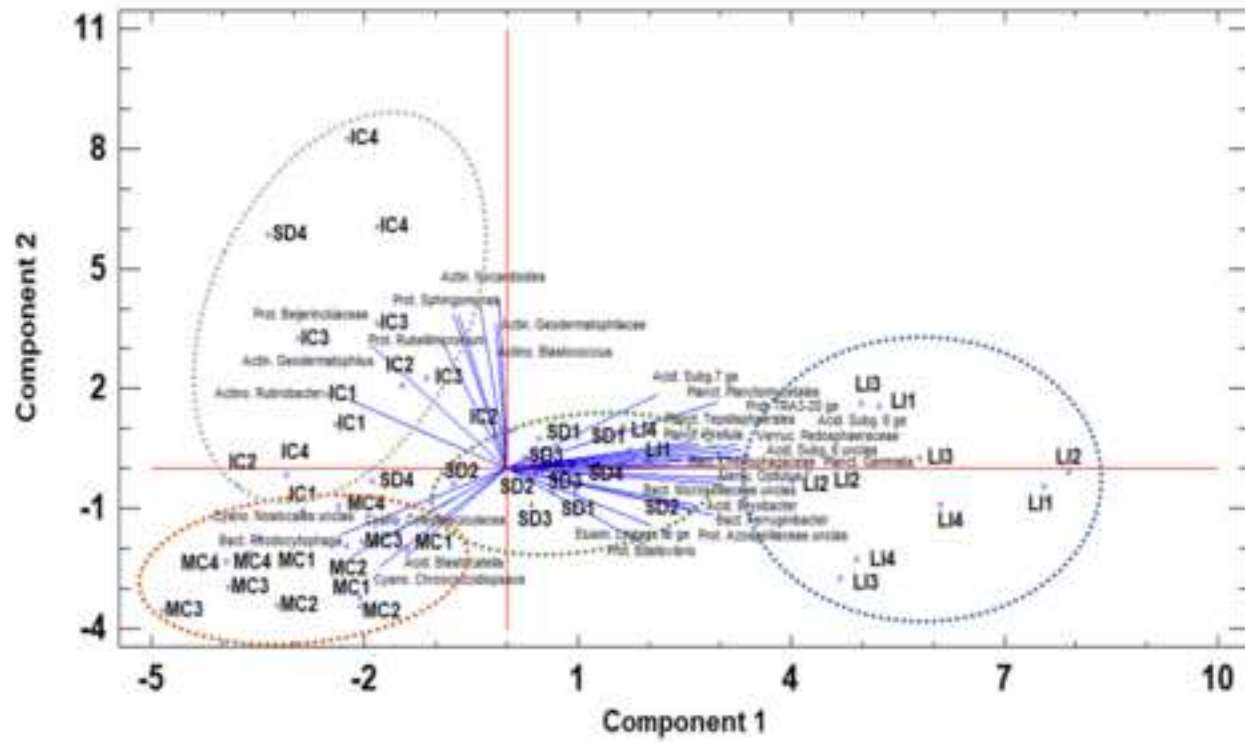
Phyla in biocrust dominated by *Cyanobacteria* (3%)

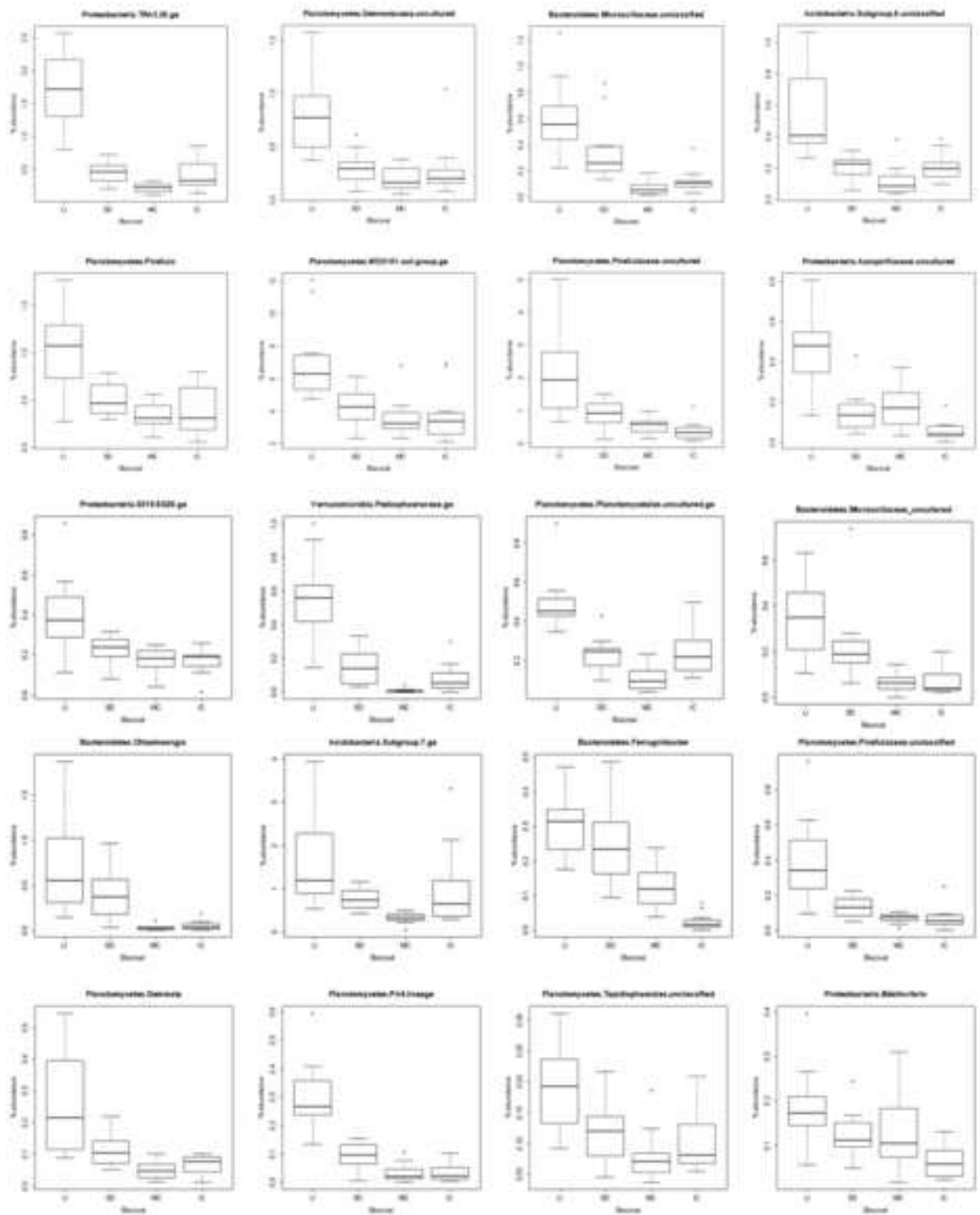


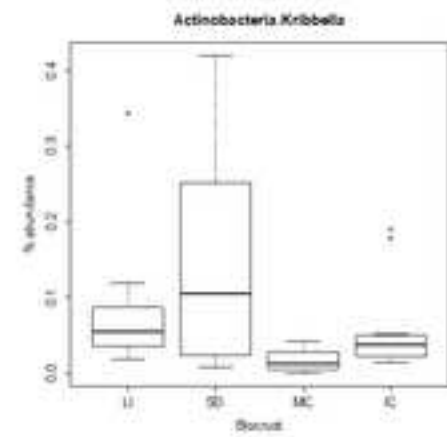
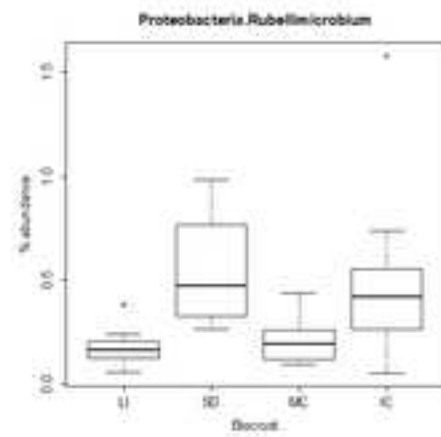
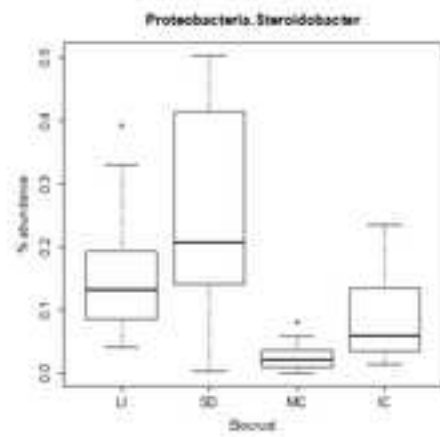
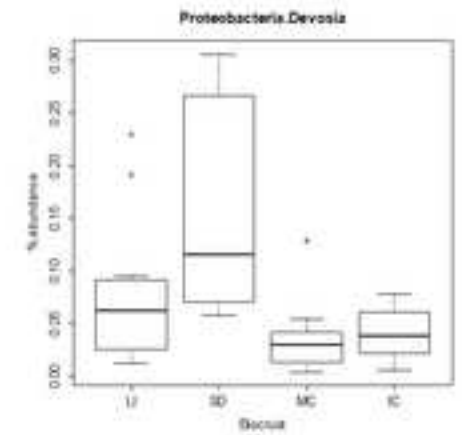
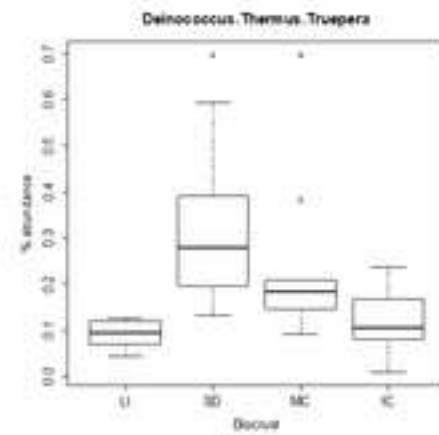
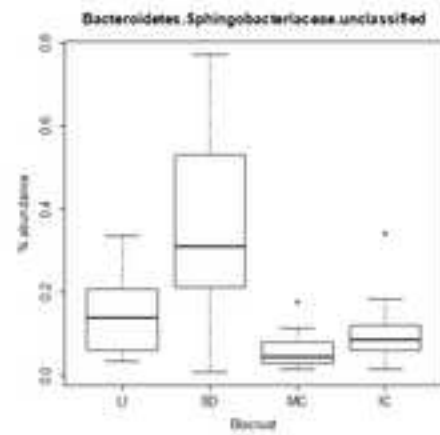
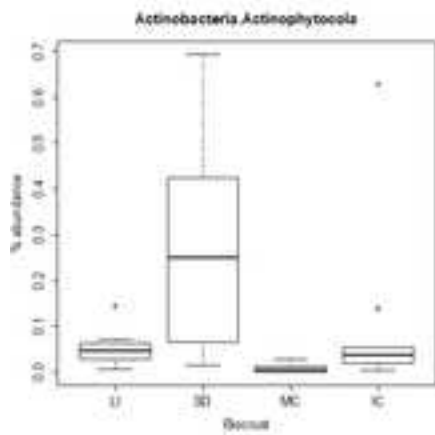
Phyla in biocrusts dominated by incipient colonization of *Cyanobacteria* (3%)

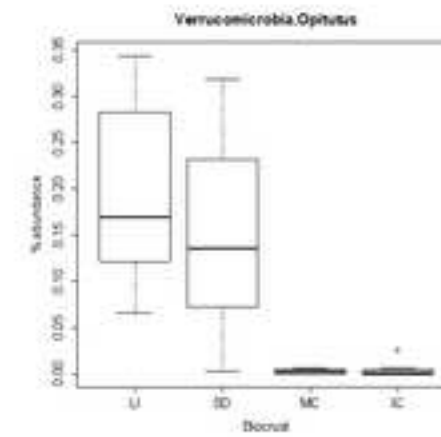
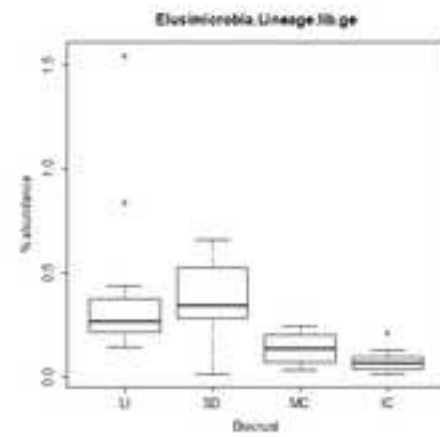
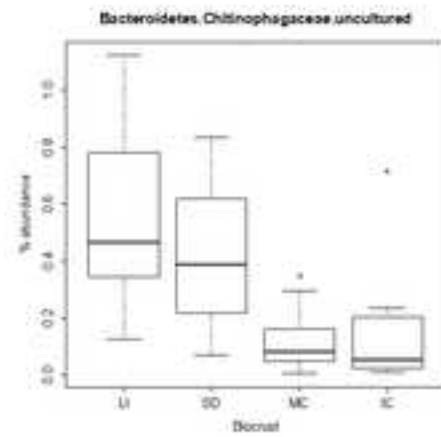
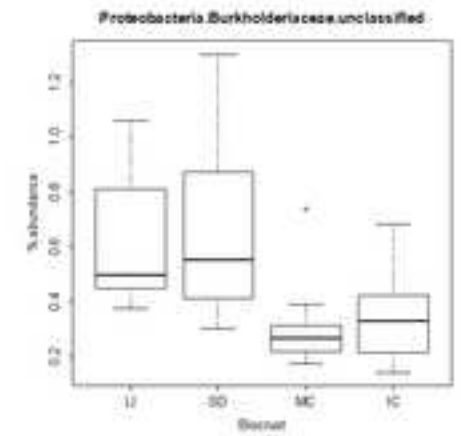
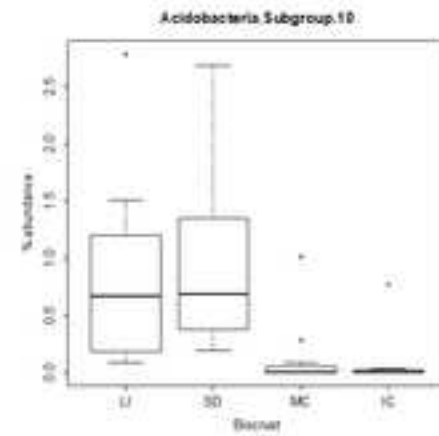
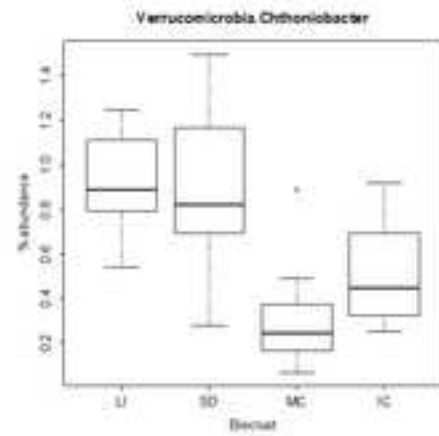
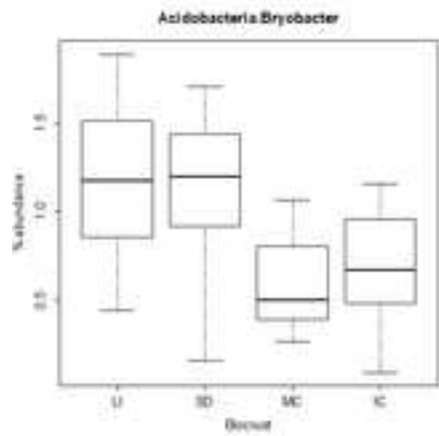


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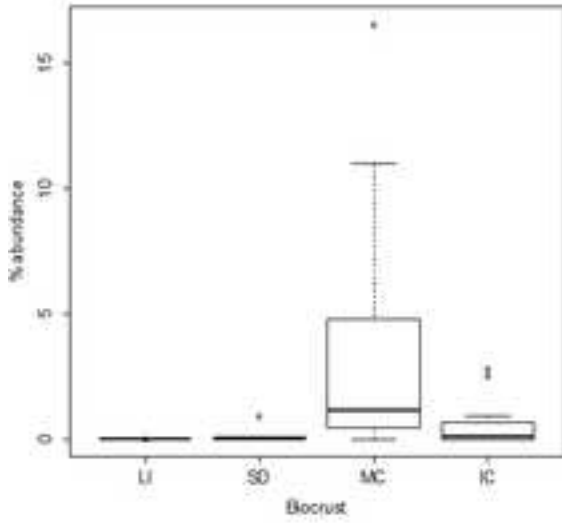




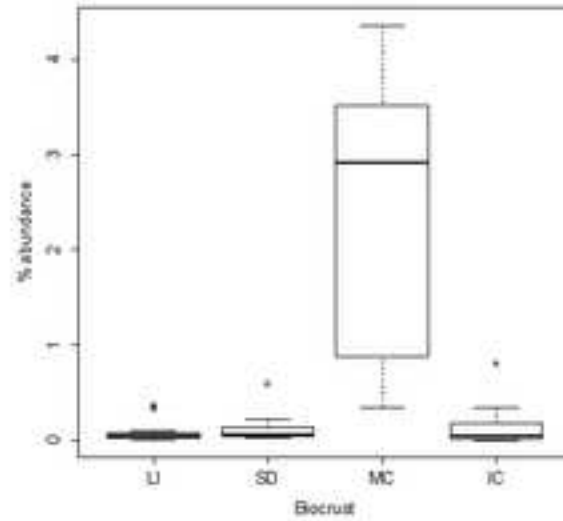




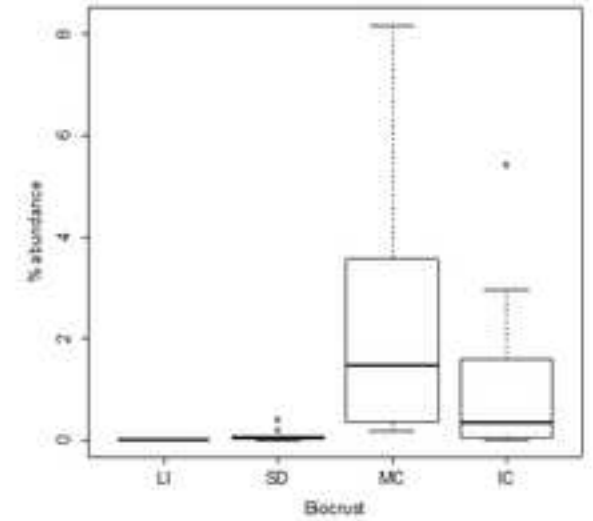
Cyanobacteria.Nostocales.uncultured.ge



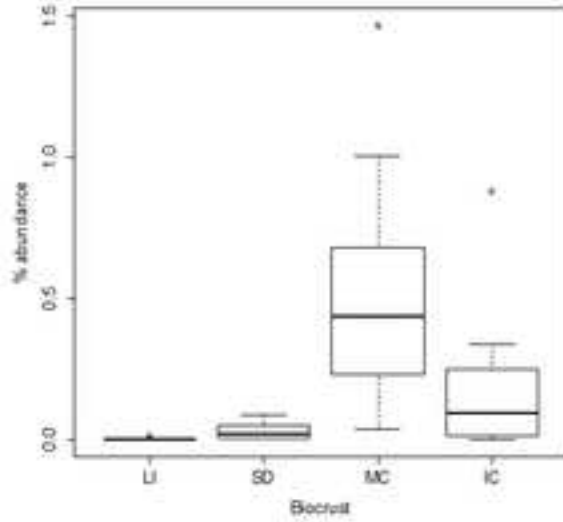
Cyanobacteria.Chroococciaceae.uncultured



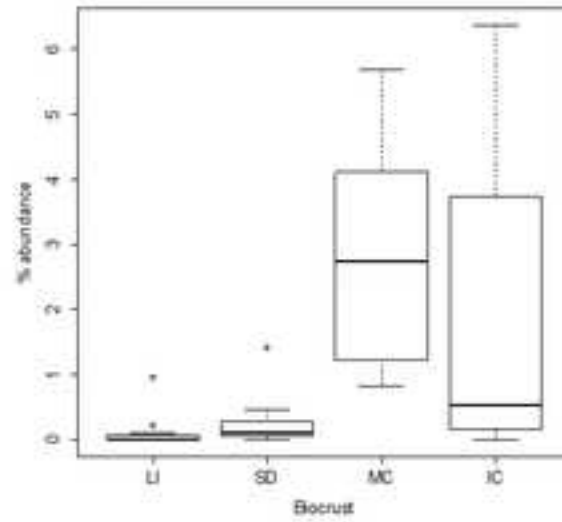
Cyanobacteria.Chroococciopsis.SAG.2023



Cyanobacteria.Coleofasciculaceae.unclassified



Cyanobacteria.Nostocales.unclassified



Acidobacteria.Blastocatella

