

1 **Research paper**

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3 **Combined bioreduction and volatilization of Se^{VI} by**
4 ***Stenotrophomonas bentonitica*: formation of trigonal selenium**
5 **nanorods and methylated species**

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7 Miguel A. Ruiz-Fresneda^{1,*}, María V. Fernández-Cantos^{1,#}, Jaime Gómez-Bolívar¹,
8 Abdurrahman S. Eswayah², Philip H. E. Gardiner³, Maria Pinel-Cabello¹, Pier L. Solari⁴,
9 Mohamed L. Merroun¹

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12 ¹Department of Microbiology, University of Granada, Granada, Spain

13 ²Biotechnology Research Centre, Tripoli, Libya

14 ³Biomolecular Sciences Research Centre, Sheffield Hallam University, Sheffield, UK

15 ⁴MARS Beamline, Synchrotron SOLEIL, L'Orme des Merisiers, Saint-Aubin, Gif-sur-Yvette
16 Cedex, France

17

18 [#]Present address: Department of Molecular Genetics, Groningen Biomolecular Sciences and
19 Biotechnology Institute, University of Groningen, Nijenborgh 7, 9747AG Groningen, The
20 Netherlands.

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22 *Corresponding author: Miguel Angel Ruiz-Fresneda. Email: mafres@ugr.es

23

24 **Abstract**

25

26 Nowadays, metal pollution due to the huge release of toxic elements to the environment has
27 become one of the world's biggest problems. Bioremediation is a promising tool for reducing
28 the mobility and toxicity of these contaminants (e.g. selenium), being an efficient,
29 environmentally friendly, and inexpensive strategy. The present study describes the capacity of
30 *Stenotrophomonas bentonitica* to biotransform Se^{VI} through enzymatic reduction and
31 volatilization processes. HAADF-STEM analysis showed the bacterium to effectively reduce
32 Se^{VI} (200 mM) into intra- and extracellular crystalline Se^0 nanorods, made mainly of two
33 different Se allotropes: monoclinic (*m*-Se) and trigonal (*t*-Se). XAS analysis appears to indicate
34 a Se crystallization process based on the biotransformation of amorphous Se^0 into stable *t*-Se
35 nanorods. In addition, results from headspace analysis by gas chromatography-mass
36 spectrometry (GC-MS) revealed the formation of methylated volatile Se species such as DMSe
37 (dimethyl selenide), DMDS₂ (dimethyl diselenide), and DMSeS (dimethyl selenenyl sulphide).
38 The biotransformation pathways and tolerance are remarkably different from those reported
39 with this bacterium in the presence of Se^{IV} . The formation of crystalline Se^0 nanorods could
40 have positive environmental implications (e.g. bioremediation) through the production of Se of
41 lower toxicity and higher settleability with potential industrial applications.

42

43 **Keywords:** bacteria, selenate, reduction, bioremediation, nanorod, volatilization

1. Introduction

Exhaustive investigations in the past decade have shed light on the occurrence of a wide diversity and distribution of microorganisms recognized as capable of selenium (Se) oxyanion bio-transformations (Avendaño et al. 2016; Presentato et al. 2018; Ojeda et al. 2020). Oxidation, reduction, and volatilization have been reported as the main pathways involved in the biotransformation of the different Se oxidation states: selenate (Se^{VI}), selenite (Se^{IV}), elemental Se (Se^0), and selenide (Se^{II}). Selenate and selenite are known to be environmentally hazardous due to their high solubility, mobility, and bioavailability, while Se^0 is insoluble and less toxic. A few studies have reported the oxidation of reduced Se species by microorganisms (Dowdle and Oremland 1998; Losi and Frankenberger 1998; Luo et al. 2022; Nancharaiyah and Lens 2015). However, microbial oxidation of Se is not usually considered to be of major relevance for the environment because of the low rates at which these reactions occur (Eswayah et al. 2016). Se volatilization is a biotransformation process now considered as a promising mechanism for bioremediation purposes. Some selenium-resistant microorganisms are able to volatilize Se through biomethylation processes (Eswayah et al. 2017). Selenium methylated compounds such as DMSe or DMDSe reportedly present limited bioavailability, toxicity, and solubility in comparison to Se oxyanions (Doran 1982; Hasanuzzaman et al 2020; Ranjard et al. 2003). Finally, the reduction of oxidized and toxic forms of Se (Se^{VI} and Se^{IV}) to Se^0 nanoparticles (NPs) has been extensively studied (Martínez et al. 2020; Tugarova et al. 2020). Even though changes in the oxidation state—hence bioreduction—have been confirmed in many cases, the specific mechanisms, regulators, and biochemical pathways involved in this process have not yet been fully elucidated. To date, the number of Se^{IV} -reducing microbial isolates is considerably higher than Se^{VI} -reducing ones, which generally require a two-step process in which Se^{IV} act as an intermediate product. Therefore, both volatilization and reduction are potential mechanisms to be used in bioremediation of contaminated environments since they involve the removal of toxic Se species. However, there is still much to be investigated about bioremediation and its possible in-situ application, so any step forward in its research would be of crucial importance in the search for an effective strategy.

Of industrial and environmental interest is the type, form, and location of the Se^0 NPs produced after Se bioreduction. Indeed, a wide array of nanoparticle shapes (spheres, nanotubes, nanorods, etc.), structures (amorphous, trigonal, monoclinical, etc.), sizes, and cellular locations has been described. This suggests that there is no unique Se biotransformation mechanism. Elemental Se exists in nature in several allotropic forms including both amorphous (*a*-Se) and crystalline varieties (monoclinic (*m*-Se) and trigonal Se (*t*-Se)). Amorphous Se tends to change to the more stable crystalline Se by heating, use of chemical reagents, and other physico-

81 chemical methods (Zhang et al. 2012). However, some bacteria can transform *a*-Se to *m*-Se and
82 *t*-Se at room temperature and without the use of additional reactive (Komova et al. 2018; Ruiz-
83 Fresneda et al. 2020; Pinel-Cabello et al. 2021). Indeed, some authors propose that *a*-Se
84 nanospheres are released from the cell and transformed to Se crystal nanostructures of different
85 shapes (Wang et al. 2010; Ruiz-Fresneda et al. 2018; Ruiz-Fresneda et al. 2020). For example,
86 the bacterium *Stenotrophomonas bentonitica* have been recently identified to reduce Se^{IV} to Se⁰
87 nanospheres (*a*-Se), different crystalline nanostructures (*m*-Se and *t*-Se), and with the formation
88 of volatile methylated Se (Ruiz-Fresneda et al. 2020). They proposed a transformation
89 mechanism from *a*-Se to Se crystals including the intracellular synthesis of the *a*-Se
90 nanospheres and their subsequent release, aggregation, and transformation in the extracellular
91 space. Unfortunately, it remains unclear how these relatively huge nanospheres—in comparison
92 with the cell size—are formed, assembled, released, and transformed. It is likely that many
93 unknown proteins, enzymes, and transport complexes may be involved. Interestingly, results
94 from recent proteomic studies indicate the possible role of specific proteins in Se^{IV} reduction,
95 including RND (resistance-nodulation-division) transport systems, and glutathione reductase
96 (Pinel-Cabello et al. 2021), which has been found to be important in Se reduction in other
97 bacterial species (Ni et al. 2015; Martínez et al. 2020).

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99 The importance of Se crystal formation lies in the potential number of applications derived from
100 it. For instance, crystalline Se has been described to be more settleable than *a*-Se (Lenz et al.
101 2009), which may be beneficial for the decontamination of Se polluted environments. Recent
102 experiments in bioreactors highlight bioremediation as potential tool for contaminated water
103 treatment (Dessi et al. 2016; Ojeda et al. 2020). Likewise, certain Se-reducing bacteria may play
104 an important role in the immobilization of Se, positively affecting the safety of deep geological
105 repositories (DGR) (Ruiz-Fresneda et al. 2018; Ruiz-Fresneda et al. 2019), the most accepted
106 option for final disposal of radioactive residues. In addition, the utility of *t*-Se in many industrial
107 and medical applications is well-known. *t*-Se is a photoconductor of broad spectral sensitivity,
108 making it very useful in solar cells, photocells, rectifiers, photographic exposure meters, and
109 xerography (Ibragimov et al. 2000; An et al. 2003; Zhu et al. 2019). From a medical standpoint,
110 Se nanoparticles hold potential as antitumoral and antibacterial agents (Kuršvietienė et al. 2020;
111 Filipović et al. 2021). It is worth mentioning that for many of the previous applications,
112 nanoparticles must be crystalline, smaller than 100 nm, and as flawless as possible (An et al.
113 2003). Although many studies are reporting the bioproduction of selenium nanoparticles
114 (SeNPs) through bacteria, archaea, plants, fungi, etc., there is still no optimized eco-friendly
115 methodology applied to the industrial production of NPs. For this reason, any contribution to the
116 field could be of great help in the search for an effective synthesis method.

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118 The study presented here describes the mechanisms involved in the reduction of Se^{VI} by *S.*
119 *bentonitica* as compared to those reported for Se^{IV} (Ruiz-Fresneda et al. 2018). This bacterium
120 reduced Se^{VI} to extracellular and intracellular Se⁰ nanorods (*m*-Se and *t*-Se) and methylated Se
121 compounds. The results evidenced a different and novel Se^{VI} reduction mechanism entailing the
122 formation of intracellular crystalline nanorods, never described before, with no observation of
123 *a*-Se nanospheres. This study thus provides new information to be considered in the
124 development of novel bioremediation strategies and tools. In addition, *S. bentonitica* is clearly
125 identified as a candidate for environmentally friendly methodologies in *t*-Se nanorod
126 fabrication.

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2. Materials and methods

2.1. Bacterial species and growing conditions under Se^{VI} stress

The bacterium *Stenotrophomonas bentonitica* used in the present study was isolated from Spanish bentonite clays (Almeria, Spain) and selected based on previous genomic and metal interaction studies (Sánchez-Castro et al. 2017). The bacterial cells were grown aerobically in Luria-Bertani (LB) broth medium (tryptone 10 g/l, NaCl 10 g/l, and yeast extract 5 g/l and, pH 7.0 ± 0.2) at 28 °C and 180 rpm. Specifically, the cells were inoculated to an initial optical density (OD) of 0.1 at a wavelength of 600 nm for all the experiments. Se^{VI} tolerance by *S. bentonitica* was studied by growing the cells aerobically in liquid LB (25 ml) added with 50, 100, and 200 mM Se^{VI} at 28 °C by shaking at 180 rpm. Growth was quantified at different incubation times (0, 8, 24, 48, and 72 h) by calculating the cell protein content following the method employed by Ruiz-Fresneda et al. (2018, 2019, 2020) based on Bradford's reagent.

2.2. X-ray diffraction (XRD)

XRD analysis was used to identify the crystalline phase of the Se reduction products. For this purpose, high volumes of *S. bentonitica* cultures (500 ml) treated with different initial Se^{VI} concentrations (50, 100, 150, and 200 mM) were harvested after 24 h (10,000 x g for 10 min). Dried powder samples were obtained as indicated by Ruiz-Fresneda et al. (2018) and measured with a Bruker D8 Advanced diffractometer linked to a LINXEYE detector.

2.3. X-ray absorption spectroscopy (XAS) measurements

XAS is a synchrotron-based analytical technique widely employed for local chemical structure determinations of different materials, and hence can be successfully used to determine the structure and oxidation state of Se allotropes produced by the cells of *S. bentonitica* (Lopez-Fernandez et al. 2020). For XAS experiments, *S. bentonitica* cells were grown in LB supplemented with 200 mM Se^{VI}. After 17, 24 and 48 h of incubation the samples were collected, washed, dried, and powdered as described above in section 2.2. Subsequently, the powder samples were pressed on Kapton tape as indicated by Ruiz-Fresneda et al. (2020). Se standards (sodium selenate-Na₂SeO₄ (Se^{VI}), sodium selenite-Na₂SeO₃ (Se^{IV}), Se⁰ foil (*t*-Se) and selenium sulphide-SeS₂ (Se^{-II})) were prepared with cellulose forming small disks following the procedures of Ruiz-Fresneda et al. 2020. The XAS data of the experimental samples and Se standards were collected in fluorescence and transmission mode, respectively.

166 Selenium K-edge X-ray absorption spectra were measured at the MARS beamline (SOLEIL
167 synchrotron facility in Paris, France), which is a bending magnet beamline for Multi Analyses
168 on Radioactive Samples. The experimental setup and technical parameters behind the
169 measurements were same as those followed in Ruiz-Fresneda et al. (2020); they are detailed in
170 the **Supplementary Material 1.1**.

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172 2.4. Electron Microscopy

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174 The crystallographic/structural analysis and cellular location of the Se reduction products were
175 analyzed by means of high-angle annular dark field scanning transmission electron microscopy
176 (HAADF-STEM) fitted with energy dispersive X-ray (EDX), selected-area electron diffraction
177 (SAED), and Fast Fourier Transform (FFT). The samples consisting of Se^{VI}-treated cells (150
178 and 200 mM Se^{VI}) were prepared as described in Merroun et al. (2005) after 24 and 48 h
179 incubating, and subsequently examined under a HAADF-STEM microscope FEI TITAN G2 80-
180 300 (University of Granada, Granada, Spain). The samples were further examined under FEG-
181 ESEM (field emission gun environmental scanning electron microscopy) on a FEG-SEM
182 Microscope FEI QEMSCAN 650F (University of Granada, Spain). The samples were prepared
183 using the critical point drying method as described in Ruiz-Fresneda et al. (2018).

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185 2.5. Gas Chromatography Mass Spectrometry (GC-MS).

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187 GC-MS combined with thermal desorption (250°C) system was employed for the qualitative
188 analysis of volatile Se-containing compounds released by *S. bentonitica* cells. To this end, the
189 cells were incubated with 2 and 100 mM Se^{VI} in special conical flasks (Quickfit™) capped with
190 Suba-Seals (Sigma™) rubber septa as described in Ruiz-Fresneda et al. (2020). Headspace
191 gases were sampled and analyzed after 144 h of incubation as described previously (Eswayah et
192 al. 2017). Se^{VI}-free cultures (biotic) and Se^{VI}-added media (abiotic) were employed as controls.
193 All measurements were performed in duplicate.

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3. Results and discussion

3.1. Selenate reduction and tolerance assays for *S. bentonitica*

The reduction of Se^{VI} was confirmed by the formation of red precipitates characteristic of Se^0 in Se^{VI} -treated cultures after 24 h of incubation (**Figure 1A**). The non-formation of such precipitates in Se^{VI} -untreated cultures (**Figure 1B**) and Se^{VI} -treated media (biotic and abiotic controls) (**Figure 1C-D**) indicated that Se^{VI} reduction is a biological process mediated by the cells of *S. bentonitica*. Interestingly, the formation of the reddish colour and thus of Se^0 nanoparticles of interest in the Se^{VI} -treated cultures was only observed at high initial concentrations (100-200 mM), but not at lower initial concentrations (10 and 50 mM) (**Figure S1-Supplementary Material**). These results suggest that Se^{VI} would not exert a high toxic effect upon the cells of *S. bentonitica* at low metalloid concentrations. For this reason, 200 mM Se^{VI} was selected as initial concentration for most of the experiments. It is noteworthy that transcriptomic studies showed that the cells respond metabolically differently with Se^{VI} concentration (Pinel, 2021). Thus, genes encoding for efflux transporters and several oxidoreductases were clearly overexpressed at concentrations below 200 mM (e.g. 50 mM). In this case, where no red precipitation is present, the cells would be capable of reducing Se^{VI} to Se^{IV} , which is removed most probably by these transporters to the extracellular space. For all the above reasons, it is quite probable that an additional resistance mechanism apart from reduction —such as volatilization, which does not trigger a colour change in the cultures— is involved.

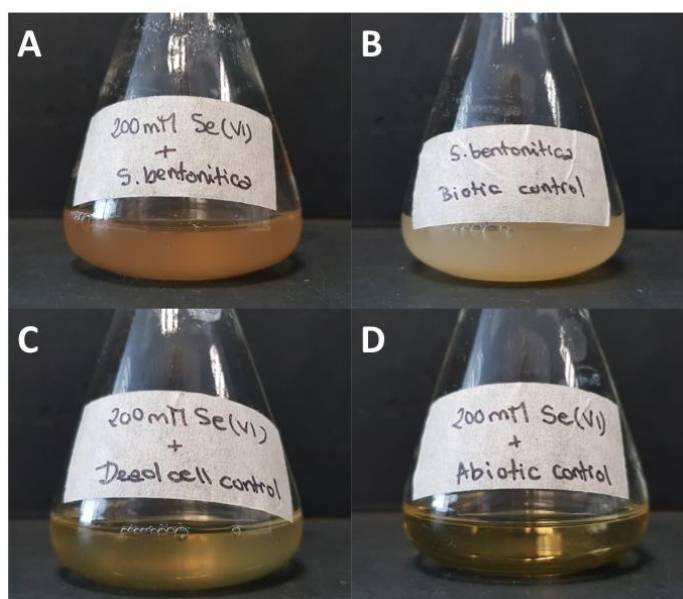
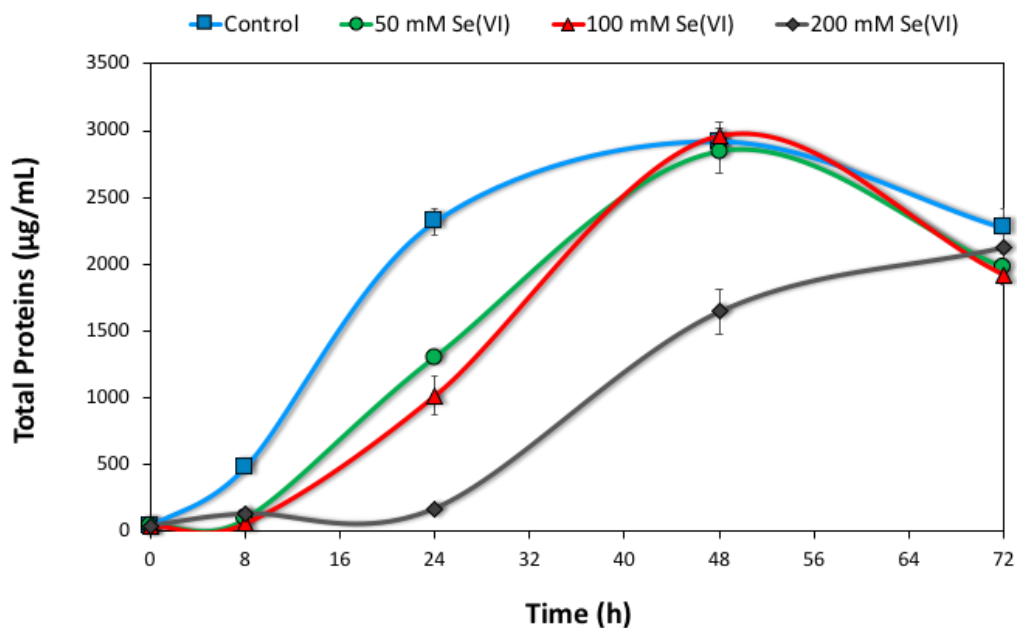


Figure 1. Cultures of *Stenotrophomonas bentonitica* in LB broth medium with (A) and without 200 mM Se^{VI} (B) after 24h incubation. Dead cell (C) and abiotic (D) controls were both supplemented with 200 mM Se^{VI} .

223 The toxicity of Se^{VI} was evaluated by determining the bacterial growth rate under different Se^{VI}
 224 concentration levels (0, 50, 100, and 200 mM). As can be seen in **Figure 2**, cell growth curve of
 225 the biotic control in the absence of selenate (0 mM Se^{VI}) showed a normal pattern of bacterial
 226 growth. Considerable differences in cell growth in the presence of 50 and 100 mM during the
 227 first 48 h confirmed the toxicity of Se^{VI}, with an observable small delay in the lag phase of
 228 growth of about 8 h. However, after 48 h growing, no differences were appreciated between the
 229 control and these Se^{VI} concentrations. When the cells were grown under 200 mM Se^{VI}, the
 230 growth rate is clearly affected. The lag and the exponential phases are prolonged until 24 and 48
 231 h, respectively. However, despite the evident toxic effect, the cells were able to reach growth
 232 levels compare with the control after 72 h, revealing the capacity of *S. bentonitica* to tolerate
 233 high concentration levels of Se^{VI}. All these observations suggest that the cells cope with the
 234 toxicity of Se^{VI} through several mechanisms depending on concentrations, including enzymatic
 235 reduction, export of Se^{VI} by oxidoreductases and transporters, and/or mitigation of oxidative
 236 stress by enzymes such as glutathione S-transferase (Pinel, 2021). The bacterium of study was
 237 previously characterized as efficiently reducing Se^{IV} at lower initial concentrations (0.1 to 2
 238 mM) than those employed here for Se^{VI} (Ruiz-Fresneda et al. 2020) —in their study, *S.*
 239 *bentonitica* growth rate was strongly affected by Se^{IV} at 2 mM. This finding reveals that *S.*
 240 *bentonitica* cells are more resistant to Se^{VI} than to Se^{IV}.



241
 242 **Figure 2.** Growth profile of *Stenotrophomonas bentonitica* at different concentrations of Se^{VI}: 0, 50, 100,
 243 and 200 mM Se^{VI}.

244 Although Se is usually present in nature at relatively low concentrations, there is a growing
 245 increase in the release of this contaminant into the environment in the last years. For instance,
 246 Chang et al. (2019) reported concentrations up to 7007 mg/kg in seleniferous soil areas in
 247

248 central China, similar to those tested in the present study. In addition, several plant species such
249 as *Astragalus L.*, *Oonopsis (Nutt.) Greene*, *Oryzopsis Michx.*, *Xylorhiza Nutt.*, or *Mentzelia L.*
250 (Presser, 1999), can accumulate from 1000 to 10,000 mg/kg selenium (dry weight), being a
251 potential source of Se in the environment. There is no report in the literature with similar results
252 with bacterial strains tolerating and reducing such high concentrations. In this regard, *S.*
253 *bentonitica* is presented herein as a potential microorganism to be used for decontamination of
254 highly polluted environment.

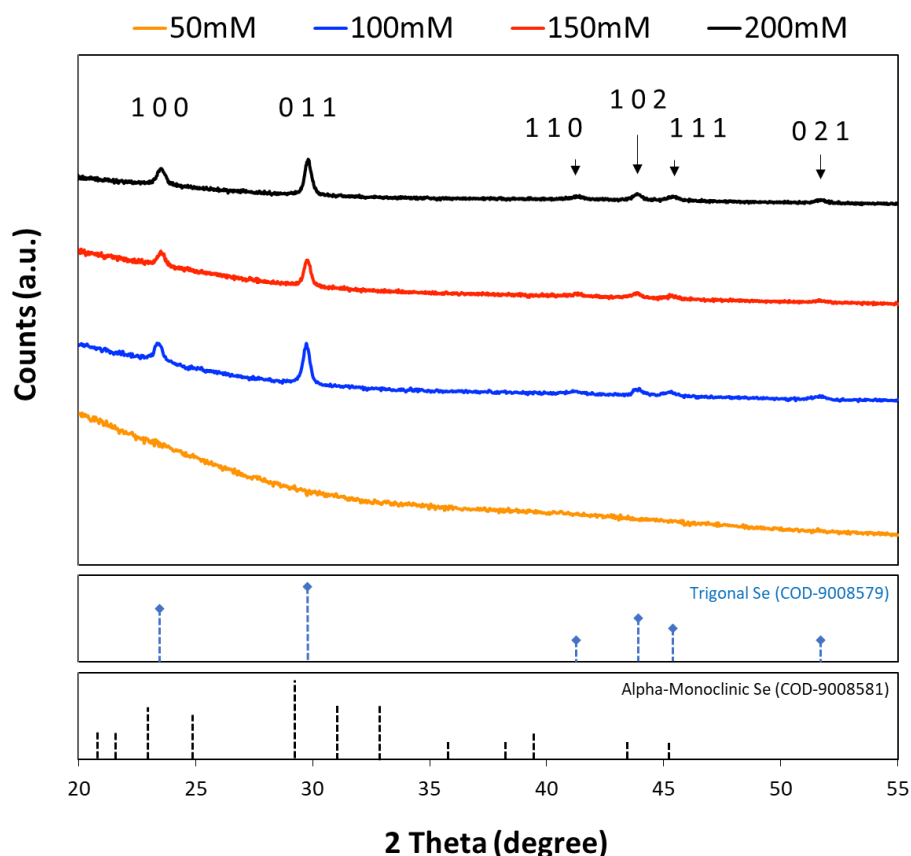
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256 3.2. XRD analysis

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258 The structure of the Se^{VI}-reduction products was characterized using the XRD technique. The
259 XRD patterns of the samples obtained after 24 h of incubation show the main peaks
260 characteristic of crystalline *t*-Se (COD-9008579) at 2θ values of 23.4°, 29.7°, 41.1°, 43.6° and
261 45.3°, and 51.7° (**Figure 3**), corresponding to the crystal planes (1 0 0), (0 1 1), (1 1 0), (1 0 2),
262 (1 1 1), and (0 2 1), respectively, for initial Se^{VI} concentrations of 100, 150, and 200 mM. For
263 50 mM Se^{VI}, no peaks were observed, most probably due to the lower toxicity exerted by this
264 element at this concentration and non-formation of Se crystals. Interestingly, for 100 and 150
265 mM concentrations, peaks belonging to the *t*-Se phase can be observed in the absence of a
266 colour change. This suggests that the reduction rate increased, but not enough to produce the
267 characteristic red precipitates. No differences in the diffraction patterns were observed when
268 varying the incubation time from 17 to 48h at the same concentration (200 mM Se^{VI}) (**Figure**
269 **S2-Supplementary Material**)

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271
 272 **Figure 3.** X-ray diffraction patterns of cultures of *S. bentonitica* supplemented with 50, 100, 150, and 200
 273 mM Se^{VI} after 24h of incubation. Peaks for Se with a *t*-Se were detected for cultures supplemented with
 274 100, 150, and 200 mM as indicated the corresponding crystal planes (black arrows). X-ray reflections of
 275 *t*-Se (COD-9008579) and m-Se (COD-9008581) obtained from Crystallography Open Database
 276 (<http://www.crystallography.net/cod/>) are shown at the bottom.

277 3.3. FEG-ESEM and HAADF-STEM analysis

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 280 A combination of FEG-ESEM and STEM/HAADF was used to determine the structure and
 281 physical properties of the Se^{VI} reduction products needed for the elucidation of biological Se^{VI}
 282 biotransformation by the cells of *S. bentonitica*. Three-dimensional (3D) images obtained by a
 283 FEG-ESEM system showed the presence of electron-dense nanorods mostly in the intracellular
 284 space of the cells, but also extracellularly, after 24 and 48 h of incubation with 150 (**Figure 4A-**
 285 **B**) and 200 mM Se^{VI} (**Figure 4C-E**). Energy Dispersive X-Ray Analysis (EDX) clearly showed
 286 these nanostructures to be composed of Se in addition to small peaks of sulfur (S) (**Figure 4F**).
 287 The presence of both Se and S suggest the participation of S-containing compounds as
 288 intermediates and biocatalysts in Se^{VI} reduction. A set of biochemical pathways implicating S-
 289 doped molecules such as glutathione (GSH), glutathione reductase (GR), glutathione peroxidase
 290 (GP), or thioredoxin reductase (TrxR) in formation of Se⁰ through a series of redox reactions
 291 have been widely reported before (Xu and Barton 2013; Eswayah et al. 2019). Interestingly,

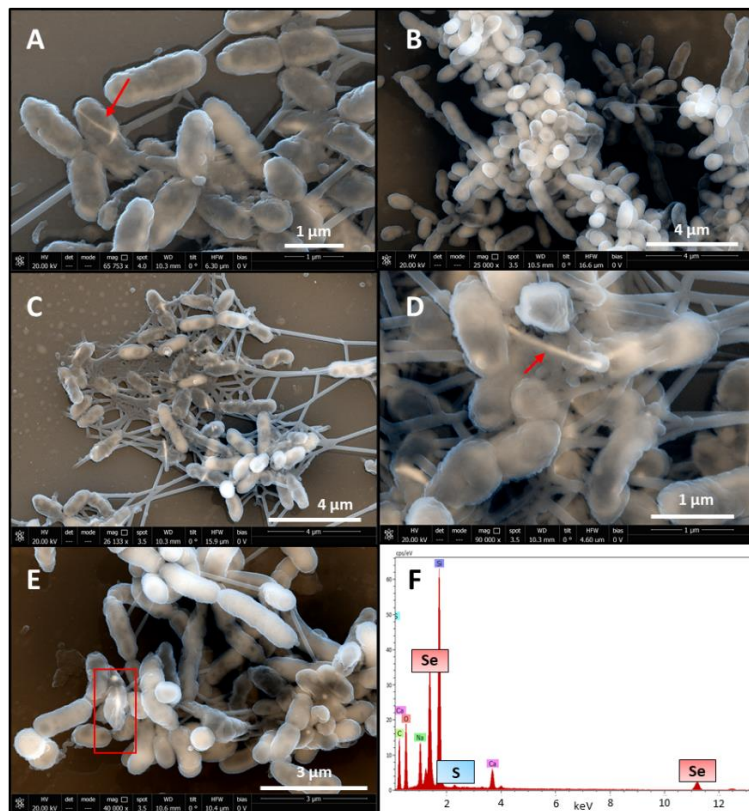
292 organic Se molecules such as selenocysteine amino acids are known to be essential in the
293 catalytic activity of TrxR or GP forming a redox active disulfide bond in the active site (Hawkes
294 and Alkan 2010). Both enzymes GR and TrxR could be involve in reduction to Se⁰ in *S.*
295 *bentonitica* since the genes encoding them are present in its genome (accession number:
296 MKCZ00000000).

297

298 The formation of Se nanoparticles having diverse morphologies is widely described (Presentato
299 et al. 2018; Fischer et al. 2020). Yet to the best of our knowledge, it is the first time that tubular-
300 shaped Se nanoparticles produced from a Se^{VI} reduction process could be observed
301 intracellularly. Interestingly, these relatively big and long nanorods can be observed
302 intracellularly without signs of cell lysis. The strain of study *S. bentonitica* has been highlighted
303 in recent years for its ability to reduce Se^{IV} to Se⁰ forming amorphous Se nanospheres and
304 trigonal nanorods (Ruiz-Fresneda et al. 2018; Ruiz-Fresneda et al. 2020). Based on such
305 findings, the authors invoked a biotransformation mechanism whereby intracellular amorphous
306 Se nanospheres are released to the extracellular space during the first 24-72 h, followed by
307 transformation to different Se allotropes (monoclinic and trigonal) after 144 h. Other authors
308 suggest a similar transformation pathway in different bacterial species: *Bacillus subtilis* and
309 *Zoogloea ramigera* (Wang et al. 2010; Srivastava and Mukhopadhyay 2013). In the present
310 study, however, Se nanorod formation was observed at 24 h and mainly located intracellularly,
311 evidencing that different transformation and interaction processes may occur in *S. bentonitica*
312 when the initial Se source is Se^{VI} instead of Se^{IV} (Ruiz-Fresneda et al. 2018). Our findings
313 suggested the reduction and Se nanorod synthesis occurred intracellularly and are then released
314 somehow. Cytoplasmic enzymes from several microorganisms (*Burkholderia fungorum* 95,
315 *Burkholderia fungorum* DBT1, or *Bacillus mycoides* SeITE01) have been reported to be
316 involved in Se^{IV} reduction (Khoei et al. 2017; Lampis et al. 2014). Lampis et al. (2014)
317 proposed thioredoxin reductase (TrxRed) systems and thiol-containing molecules with redox
318 activity in Se⁰ nanoparticle formation of *B. mycoides* SeITE01. Different export mechanisms
319 have been hypothesized by several authors causing SeNPs leak outside the cells. Some of them
320 suggested the SeNPs are released after cell death and lysis (Lampis et al. 2014). A few
321 extracellular nanorods were produced in our case, but no signs of bacterial lysis were observed
322 after 24h. Other export system proposed were based on secretion of the NPs through outer
323 membrane vesiculation and encapsulation (Kulp et al. 2010). For example, the gram-negative
324 bacterium *Thauera selenatis* use the protein Se factor A (SefA) for SeNPs binding, stabilization,
325 and secretion to the extracellular space (Debieux et al 2011). Further upcoming investigations
326 will help to elucidate the synthesis pathway for *S. bentonitica*.

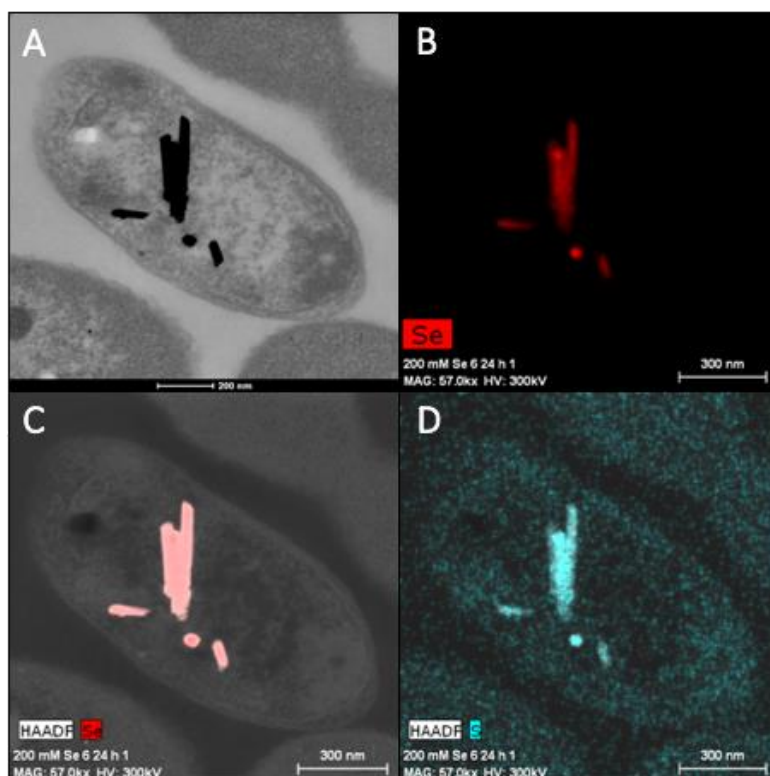
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328 From an industrial point of view, the faster production of crystalline SeNPs presented herein
329 points out a most promising crystalline NPs synthesis process compared to others. Wang et al.
330 (2010) reported a more rapid formation of crystalline Se nanorods (12 h) by *B. subtilis*, but with
331 the help of reactive agents such as ethanol. Similar Se tubular structures were produced by
332 anaerobic granular sludge after 18 h at 55°C (Jain et al. 2017). Interestingly, *S. bentonitica* can
333 form *t*-Se nanorods without the use of chemicals and room temperature, providing a more
334 ecologic and less expensive synthesis method. In terms of quantitative data, *S. bentonitica* is
335 highlighted herein as promising SeNPs producer as observed in the micrographs of thin sections
336 with up to 6 nanorods in a single cell (**Figure 5**).
337



338
339 **Figure 4.** FEG-ESEM of cultures of *S. bentonitica* supplemented with 150 mM of Se^{VI} after 24 (A) and
340 48h (B), and with 200 mM of Se^{VI} after 24 (C and D) and 48h (E). EDX analysis of a single nanorod
341 (highlighted area in panel in E) confirmed the presence of Se (F).

342
343 Ultrathin sections observed with an HAADF-STEM system allowed for further structural
344 characterization of the Se nanostructures produced by *S. bentonitica*. By means of this technique
345 we again observed that the predominantly Se nanostructure was found intracellularly in the form
346 of nanorods (**Figure 5A**). Element-distribution maps indicated the presence of both Se and S in
347 the nanorods (**Figure 5B-D**), strongly supporting the possible relevance of thiol-containing
348 proteins in the reduction pathway.



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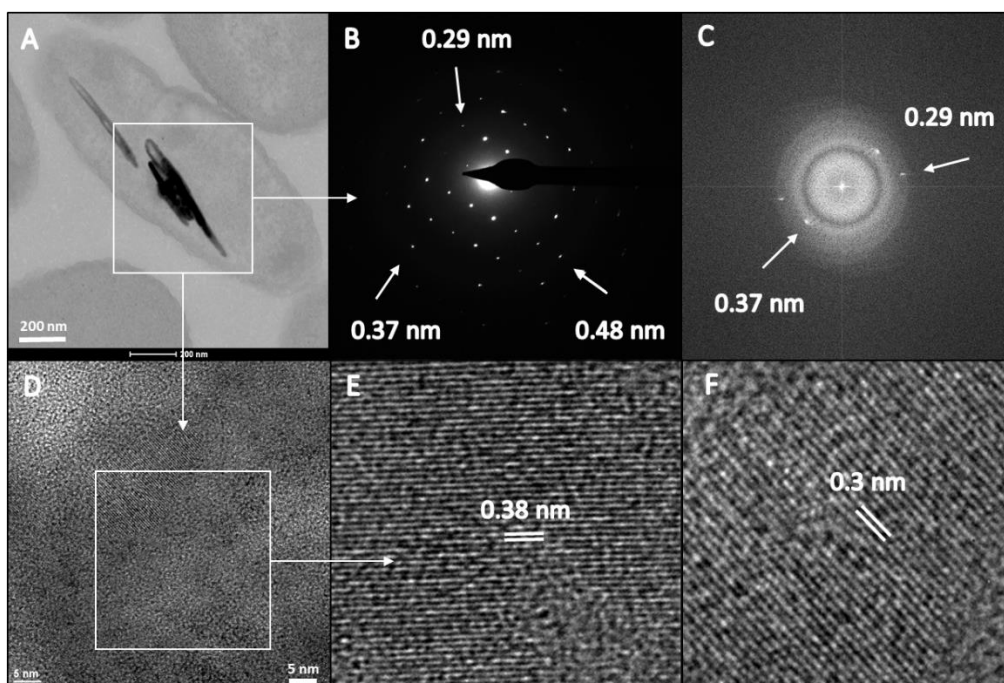
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Figure 5. HAADF-STEM micrographs of a thin section showing Se nanostructures produced by *S. bentonitica* after 24 h of incubation with Se^{VI} (A). EDX element-distribution maps showing their Se and S elemental composition (B-D).

A combination of tools of the HAADF-STEM system—including SAED (Selected Area Electron Diffraction), FFT (Fast-Fourier Transform), and STEM high resolution (HRTEM)—attested to the crystalline nature of the Se nanostructures. SAED and FFT patterns from a selected Se nanorod revealed signs of crystallization (**Figure 6A-C**); specifically, three different d-spacings of 0.29, 0.37, and 0.48 nm could correspond to crystal planes of monoclinic-Se (*m*-Se) (Pinel-Cabello et al. 2021) (**Figure 6B**). The space lattices 0.29 and 0.37 nm are characteristic as well in trigonal Se (*t*-Se) according to the American Mineralogist Crystal Structure Database (<http://rruff.geo.arizona.edu>) [accessed December, 2021]. Yet the detection of the 0.48 nm d-spacing confirmed the monoclinic structure of the nanorods. Similar results were obtained recently for Se nanorods produced by *S. bentonitica* contacted with Se^{IV} (Pinel-Cabello et al. 2021). HRTEM analyses, along with the FFT from a nanorod selected area, agree with the SAED pattern, revealing the presence of 0.29-0.3 and 0.38 nm lattice spacings (**Figure 6C-F**). As mentioned above, these spacings could correspond to different crystal planes of both *m*-Se and *t*-Se. More specifically, the d-spacing of 0.3 nm could correspond to planes (1 0 1 or 0 1 1) of *t*-Se and many different planes of *m*-Se. The d-spacing of 0.37 nm may correspond to the plane (1 0 0) of *t*-Se, and (0 2 2) or (4 0 0) of *m*-Se. Therefore, the existence of *t*-Se should not be discarded since common d-spacings (0.3 and 0.38 nm) between *t*-Se and *m*-Se were

371 observed. Even the co-existence of mixed crystal phases in one same nanocrystal is possible in
 372 our Se nanostructures, as reported before for Sn nanoparticles (Haq et al. 2019). In fact, *t*-Se
 373 was previously detected in our samples with XRD (see section 3.2.). Naturally, XRD is a more
 374 representative and precise technique, as the bulk sample is measured, whereas under HAADF-
 375 STEM some selected crystals were analyzed. Ultimately, these results could indicate the
 376 biotransformation of *m*-Se (as a transitional allotropic form) to *t*-Se, as was reported for Se^{IV}
 377 (Ruiz-Fresneda et al. 2018). Still, the results clearly underline the importance of a
 378 multidisciplinary approach, combining microscopic and spectroscopic techniques, to
 379 exhaustively characterize nanoparticle structures.

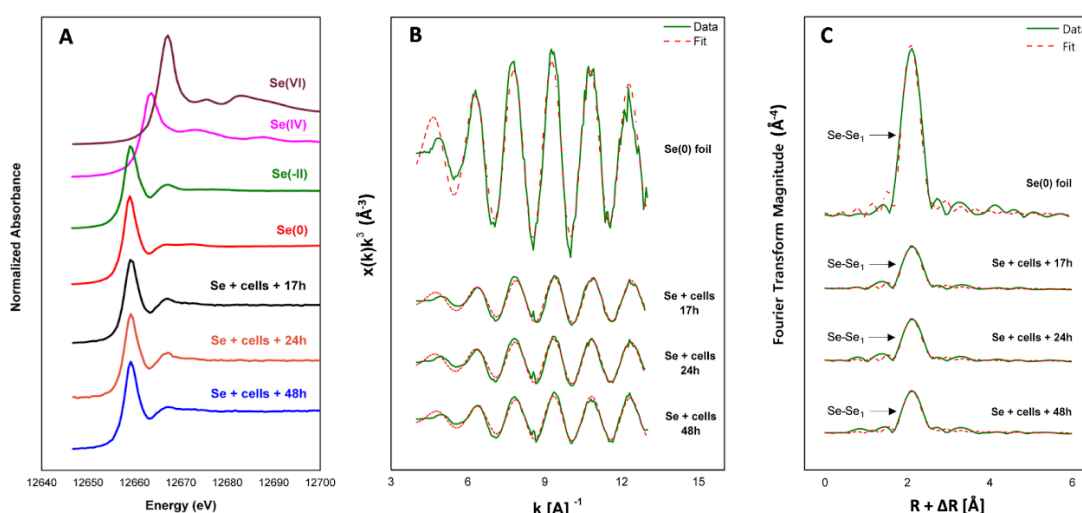


380
 381 **Figure 6.** HAADF-STEM micrograph of Se nanostructures produced by *S. bentonitica* incubated with
 382 200 mM Se^{VI} for 24 h (A). SAED pattern (B), FFT pattern (C), and HRTEM image (D) derived from a
 383 single nanostructure (highlighted area in panel A). Magnified HRTEM images (E and F) corresponding to
 384 highlighted area in D.

385
 386 **3.4. XAS analysis**

387
 388 X-ray absorption spectroscopy (XAS) spectra of the Se^{VI}-reduction products obtained after 17,
 389 24, and 48 h of incubation with *S. bentonitica* (200 mM Se^{VI} as initial concentration) provided
 390 more detailed structural information concerning the local coordination environment, as well as
 391 the oxidation state of the Se in the studied samples. The X-ray absorption near-edge structure
 392 (XANES) region clearly showed that the local coordination of Se is dominated by Se⁰ at all
 393 incubation times, as indicated by the maximum peak positions obtained at 12659.3 eV, 12659.3
 394 eV, and 12659.2 eV after 17, 24, and 48h incubating (**Figure 7A**) (**Figure S3 and S4-**

395 **Supplementary Material**). Generally, most of the bioreduced SeNPs in solid form consisted of
 396 Se in the zero-valent oxidation state (Zhang et al. 2012; Vogel et al. 2018; Ruiz-Fresneda et al.
 397 2020).
 398 The extended X-ray absorption fine structure (EXAFS) spectra of the Se^{VI}-reduction products
 399 and Se foil (as Se⁰ reference compound), along with their corresponding Fourier transforms
 400 (FT) and fit parameters of the obtained spectra, are respectively presented in **Figure 7B-C**,
 401 **Figure S5-Supplementary Material**, and **Table 1**. FT peak distances are reported in units of Å
 402 and are expressed as R + ΔR. The fit of the EXAFS spectra of the 3 experimental samples
 403 indicated the presence of one Se-Se coordination shell at a bond distance of about 2.35-2.37 ±
 404 0.02Å (**Table 1**). In previous XAS analyses obtained for the Se nanostructures produced by *S.*
 405 *bentonitica* with Se^{IV} this parameter ranged between 2.33 and 2.34 Å and were attributed to
 406 amorphous Se in view of the literature (Eswayah et al. 2017; Vogel et al. 2018; Ruiz-Fresneda
 407 et al. 2020). Slight increases of Se-Se bond distance values could indicate a crystallization
 408 process, while slight decreases in this frame point to an amorphization. Such is the case of the
 409 experiments by Zhao et al. (2004) and Breynaert et al. (2008), who found an amorphization
 410 process of crystalline Se upon a decrease from 2.37 to 2.35 Å. The increase of the Se-Se bond
 411 distances with incubation time found in our samples accordingly suggests a heightened
 412 structural order and therefore crystallization of Se⁰ over time. In view of the literature discussed
 413 above, the bond distances of 2.35 and 2.36 Å found after 17 and 24 h could correspond to a
 414 mixture of amorphous and trigonal Se, while the value of 2.37 Å found for 48 h samples might
 415 mark a Se crystalline phase mainly as trigonal Se (Zhao et al. 2004; Breynaert et al. 2008).
 416 However, our results are not conclusive about a dependent time crystallization process.



417
 418 **Figure 7.** XANES (A), EXAFS (B), and their corresponding FT spectra (C) of Se reference compounds [Se^{VI}
 419 (Na₂SeO₄), Se^{IV} (Na₂SeO₃), Se⁰ (Se foil), and Se^{-II} (SeS₂)] and *S. bentonitica* samples incubated with 200 mM Se^{VI} at
 420 different incubation times (17, 24, and 48 h).
 421

422

Table 1. EXAFS structural parameters of the Se foil and the Se^{VI}-reduction products.

Sample	Shell	N ^a	R[Å] ^b	σ ² [Å ²] ^c	ΔE[eV]
Se foil	Se-Se ₁	3.3 ± 0.2	2.37	0.0043	-5.6
Se ^{VI} - Cells-17h	Se-Se ₁	0.7 ± 0.1	2.35	0.0031	-3.1
Se ^{VI} - Cells-24h	Se-Se ₁	0.7 ± 0.1	2.36	0.003	-3.8
Se ^{VI} - Cells-48h	Se-Se ₁	0.6 ± 0.1	2.37	0.0028	-1.2

423

^a Errors in coordination numbers are ±25% and standard deviations as estimated by EXAFSPAK

424

^b Errors in distance are ±0.02 Å

425

^c Debye-Waller factor

426

427

3.5. Biological production of volatile Se compounds

428

429

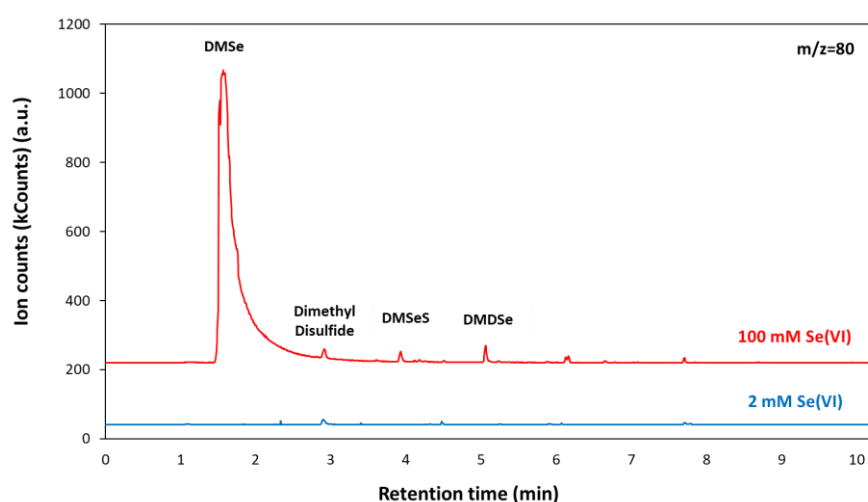
Headspace analysis using GC-MS of extracted volatiles from 100 mM Se^{VI}-treated cultures indicated the formation of dimethyl selenide (DMSe), dimethyl diselenide (DMDSe) and dimethyl selenenyl sulphide (DMSeS) by the cells after 144 h incubating (**Figure 8**). In contrast, no volatile Se-containing species were detected in biotic and abiotic controls (**Figure S6-Supplementary Material**). The Se^{VI} initial concentration of 100 mM was selected for this analysis to increase the amount of volatiles that are produced. At this concentration, trigonal Se⁰ red precipitates derived from reduction processes were not as observable as for 200 mM, suggesting that other interaction mechanisms such as volatilization may have happened. The formation of Se volatile compounds presented herein revealed that volatilization is involved as a biotransformation mechanism in *S. bentonitica*, in addition to reduction. It is worth noting that no Se volatile species were produced by the cells when the Se^{VI} initial concentration was 2 mM (**Figure 8**). This suggests, as discussed in previous sections, that a very high Se^{VI} initial concentration is needed to let biotransformation occur, whether for volatilization or reduction process.

443

444

Interestingly, this bacterium does not produce DMSe when amended with Se^{IV} instead of Se^{VI} (Ruiz-Fresneda et al. 2020), thereby suggesting a different biotransformation pathway depending on the oxidation state of Se, even in the same microorganism. Thus, the versatile role of *S. bentonitica* in Se volatilization depends on the physico-chemical conditions. Not only is the Se oxidation state important, but also the type of SeNPs produced can influence Se volatilization. Otsuka and Yamashita (2020) showed that the structural nature of SeNPs affects the Se volatilization rates in *P. stutzeri* NT-I, a bacterial strain producing higher amounts of DMDSe when amorphous SeNPs were used as substrate as opposed to crystalline SeNPs. The differences between the nanoparticles formed by *S. bentonitica* (nanospheres and nanorods from Se^{IV}; only nanorods with Se^{VI}) might therefore condition the volatile compounds produced.

453



455

456 **Figure 8.** GC-MS chromatograms of the headspace gas of *S. bentonitica* cultures supplemented with Se^{VI}
 457 (2 and 100 mM) and after 144 h of incubation. All GC-MS chromatograms were obtained by selecting the
 458 80 m/z ion specific for selenium.

459

460 No evident biochemical pathway is expounded in the literature to clearly explain how Se
 461 biovolatilization occurs. Most studies indicate that Se volatilization relies on reduction and
 462 methylation reactions when the initial form is a Se oxyanion. Accordingly, Se^{VI} or Se^{IV} are
 463 firstly reduced to Se⁰ or Se^{-II}, and subsequently methylated to DMSe (Eswayah et al. 2016).
 464 Methylation appears to be the most crucial or differential step, since a great number of reactions
 465 and intermediates —MeSeH, dimethyl disulfide (DMDS), dimethyl selenone, and Se
 466 aminoacids (SeMet and SeCys) have been suggested to be involved (Chasteen and Bentley
 467 2003; Winkel et al. 2015). DMDS may be involved in the formation of DMSeS through its
 468 reaction with DMDSe (Chasteen 1993). The presence of these three compounds (DMDS,
 469 DMDSe, and DMSeS) in our samples comes to support this hypothesis as a possible mechanism
 470 active in *S. bentonitica* and the role of S-containing enzymes in the reduction process. Current
 471 evidence would suggest that the biochemical pathways responsible for Se volatilization remains
 472 to be elucidated, and it is likely that a variety of mechanisms could be involved depending on
 473 the given organism and the form of Se. From an environmental perspective, Se biomethylation
 474 is held to constitute a detoxification mechanism, as it enables the removal and transformation of
 475 toxic inorganic Se precursors toward less toxic methylated forms (Wilber 1980; Ranjard et al.
 476 2003). This fact in itself points to a positive role of *S. bentonitica* in the bioremediation of Se
 477 contaminated sites. In the context of deep disposal of radioactive waste, however, Se production
 478 by *S. bentonitica* could compromise the integrity of such repositories by causing gas
 479 overpressure.

480

481 According to the results obtained both enzymatic reduction and volatilization are involved in
482 Se^{VI} interactions with *S. bentonitica*. When Se^{VI} initial concentration is below 100 mM Se^{VI} is
483 probably reduced to Se^{IV} by S-containing oxidoreductases (e.g. GR or TrxR) and then removed
484 through efflux transporter to the extracellular space since no red precipitates were observable.
485 This hypothesis agrees with preliminary transcriptomics studies (to be published). However, at
486 concentrations over 100 mM the cells are capable of reducing Se^{VI} to Se^0 and $\text{Se}^{-\text{II}}$ volatile
487 compounds due to these more stressful conditions. The experimental data suggested the
488 reduction to Se^0 nanorods occurs intracellularly, before being released probably through vesicle
489 secretion. The observation of organic matter surrounding some nanorods could be a sign of this
490 (Figure 4D).

491

492 3.6. Environmental and industrial significance of the different Se allotropes produced

493

494 To sum up, the combination of spectroscopic and microscopic techniques showed the great
495 potential of *S. bentonitica* to reduce Se^{VI} to Se^0 as crystalline nanorods with different allotropes
496 (monoclinic and trigonal Se). The presence of *t*-Se crystalline phase in the biogenic nanorods
497 was demonstrated by XRD. Lattice d-spacing values corresponding to *m*-Se could be observed
498 by high resolution TEM analysis. These findings suggest a crystallization process from *m*-Se to
499 the most thermodynamically stable phase *t*-Se. The XANES region of the XAS spectra
500 confirmed the presence of the zero valent oxidation state of the Se reduction products, while the
501 EXAFS region of the spectra showed the possible presence of different Se allotropes
502 (amorphous, monoclinic, and trigonal). In addition, these Se nanorods could be stabilized by
503 their interactions with organic matter as it was demonstrated for Se nanostructures derived from
504 reduction of Se^{IV} by this bacterial strain (Ruiz-Fresneda et al. 2020). However, Se-C
505 coordination shell does not seem to be visible in the EXAFS spectra of the studied samples.
506 Therefore, further studies in the characterization of organic matter coating the Se nanorods
507 should be performed using spectroscopic techniques like Infrared spectroscopy.

508

509 Ultimately, the results presented here point to *S. bentonitica* as an interesting bacterial model for
510 bioremediation systems and for its potential as a new, green, and faster (only 24 h) way to
511 produce crystalline SeNPs of substantial interest for industrial and medical applications.

512

513

4. Conclusions

514

515 The present work describes, for the first time, the ability of the bacterium *S. bentonitica* to
516 reduce Se^{VI} to Se^0 nanorods and Se volatile compounds. The results clearly demonstrate a higher
517 tolerance capacity of this bacterial species against Se^{VI} when compared to Se^{IV} experiments

518 performed in previous research. Differences in the structure, location, and formation speed of
519 the Se reduction products when Se^{VI} is the initial source suggest a different biotransformation
520 pathway than the one proposed for Se^{IV}. Still, more research is needed to fully understand the
521 specific biotransformation processes. An exhaustive combination of spectroscopic and
522 microscopic techniques applied here revealed important structural and chemical data, including
523 the oxidation state and the crystalline phases of bio-reduced Se. The zero-valent oxidation state,
524 tubular shape, and crystallinity of the Se nanostructures produced would uphold *S. bentonitica*
525 as a potential bioremediation candidate in contaminated environments. The formation of volatile
526 methylated species furthermore points to a positive role in the removal of toxic Se from polluted
527 soils and waters, in terms of atmospheric bioremediation. Finally, *S. bentonitica* is put forth as a
528 candidate for novel, environmentally friendly, and relatively quick Se nanorod fabrication.

529

530 **CRedit authorship contribution statement**

531

532 **Miguel A. Ruiz-Fresneda:** Conceptualization, Methodology, Validation, Formal Analysis,
533 Investigation, Writing-Original Draft, Writing-Review & Editing, Visualization. **María V.**
534 **Fernández-Cantos:** Conceptualization, Methodology, Validation, Formal Analysis,
535 Investigation, Writing-Review & Editing. **Jaime Gómez-Bolívar:** Methodology, Validation.
536 **Abdurrahman S. Eswayah:** Methodology, Validation. **Philip H. E. Gardiner:**
537 Conceptualization, Resources, Writing-Review & Editing, Supervision. **Maria Pinel-Cabello:**
538 Formal analysis, Writing-Review & Editing. **Pier L. Solari:** Methodology, Formal Analysis.
539 **Mohamed L. Merroun:** Conceptualization, Methodology, Formal Analysis, Resources,
540 Writing-Original Draft, Writing-Review & Editing, Supervision, Project Administration,
541 Funding Acquisition.

542

543 **Declaration of Competing Interest**

544

545 Non-declared

546

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548

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553

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