



Crop residues in corn-wheat rotation in a semi-arid region increase CO₂ efflux under conventional tillage but not in a no-tillage system

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ABSTRACT

Appropriate management of crop residue plays a key role in mitigating greenhouse gas emissions. However, it has been inadequately implemented in general agricultural management practices. In a field investigation using static chambers, we evaluated the effects of crop residue at three different rates - 100 % (R₁₀₀), 50 % (R₅₀), and residue removal (R₀) - on carbon dioxide (CO₂) efflux. The field study was conducted in corn-wheat rotation under conventional (CT) and no-tillage (NT) systems in a semi-arid region. The main results showed that CO₂ efflux was positively correlated with higher soil temperature (0.43–0.79) and microbial biomass carbon (0.66–0.89). The crop residue treatments affected these traits. A strong positive relationship between CO₂ efflux and the crop residue (R² = 0.96, CT and R² = 0.9 for NT) was observed. In the CT system, significant increases were detected among residue rates on cumulative CO₂ efflux, where R₁₀₀ and R₅₀ resulted in 36 % and 25 % higher cumulative CO₂ efflux, respectively, than R₀. In contrast, there was no significant difference in cumulative CO₂ efflux among the crop residue retention (R₁₀₀ and R₅₀) and removal (R₀) treatments under the NT system. Our study revealed that crop residue retention led to increased CO₂ efflux under the CT system in semi-arid conditions during the first year of application, while under the NT system, CO₂ efflux was not impacted by crop residue. Our results indicate that there is considerable potential for improving soil management practices in the context of soil degradation, climate change, increasing crop productivity, and carbon (C) sequestration.

1. Introduction

The continuous increase in the concentration of greenhouse gases (GHGs) produced by human activities has become a major concern due to their direct impact on global warming and climate change (Oertel et al., 2016). The concentration of CO₂ in the atmosphere increased from 278 ppm in 1750 to 390.5 ppm in 2011 (Ciais et al., 2013), and then up to 418 ppm in 2021 (Tanhua et al., 2021). CO₂ emission is the primary mechanism of soil C loss and contributes to the elevated CO₂ concentration in the atmosphere (Parkin and Kaspar, 2003). Most of the CO₂

emitted from the soil is due to the decomposition of plant residue and roots, due to microbial metabolism and respiration (Sainju et al., 2008; Campbell et al., 2014).

It was estimated that 75–120 Pg C per year is emitted from soil to the atmosphere (Hibbard et al., 2005). Therefore, maintenance and build-up of soil C are essential in improving soil functions and offsetting atmospheric CO₂ concentration (Smit, 2004; Delgado-Baquerizo et al., 2017), which is a key challenge for humankind (Lal, 2004; Scharlemann et al., 2014; Rodrigo-Comino et al., 2020a). Crop residue management affects CO₂ emissions, primarily via altering C and nitrogen (N) cycling

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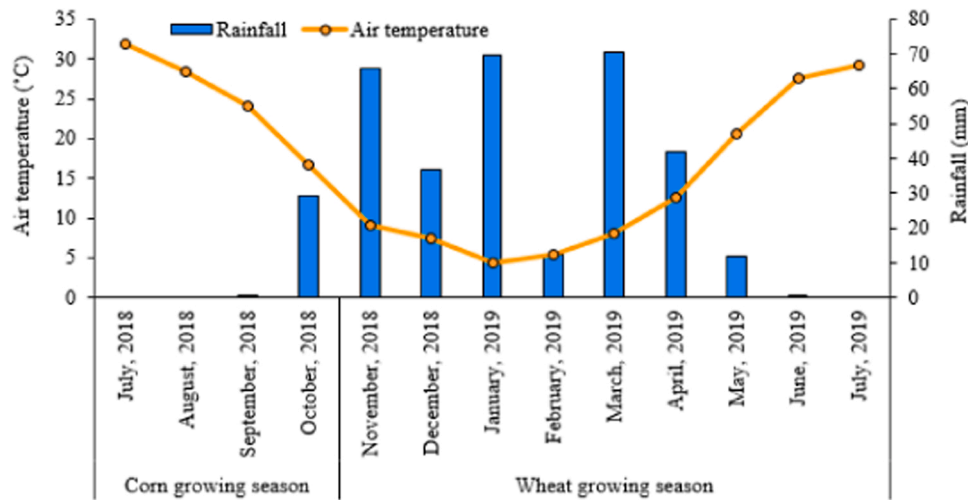


Fig. 1. Monthly precipitation and mean air temperature during the period of the experiment.

(Guzman et al., 2015; Nawaz et al., 2017). However, multiple factors regulate CO₂ emissions from the soil. Changes in soil temperature and moisture levels (Dong et al., 2017), organic matter (OM) inputs from decomposing residues- and their effect on faunal and microbial activities, all influence soil respiration (Toosi et al., 2012).

Due to the complexity of factors governing soil respiration following the addition of crop residues and their interactions, contrasting observations have been reported in the literature, i.e., increases (Badía et al., 2013; Sugasti and Pinzón, 2020) or decreases in soil CO₂ emissions following the application of crop residue in soil (Bai et al., 2017). Most cropping lands in Iran are semi-arid regions (i.e., high pH and CaCO₃, low OM, and suppressed biological status) where corn-wheat rotation is a common grain production cropping system. Despite its value as a source of OM, much of the postharvest crop residue is grazed or collected as a feed source. Residue removal and excessive tillage are the main cause of soil OM depletion in grain crops in the region (Mirzaei et al., 2021). Efforts have been made at the national level to encourage farmers to return crop residues to the soil. The effect of residue retention on CO₂ emission from the soil, typical of semi-arid conditions, has not been adequately studied.

Previous studies have shown that frequent drying-wetting cycles in arid and semi-arid conditions could lead to substantial C losses in soil from newly added OM (Borken and Matzner, 2009). We hypothesized that: i) the addition of crop residue to the soil results in higher CO₂ efflux compared to residue removal, independent of the tillage effect; ii) the cumulative CO₂ efflux would increase with an increasing amount of crop residue to the soil.

To test the above-mentioned hypotheses, we monitored CO₂ efflux in a field study, using crop residue at three rates – 100% (R₁₀₀), 50% (R₅₀), and residue removal (R₀) in corn-wheat rotation under conventional (CT) and no-tillage (NT) systems. As key regulators of soil respiration, microbial biomass and soil moisture and temperature were also monitored during the experiment to link their possible association with the soil CO₂ efflux.

2. Materials and methods

2.1. Site characteristics and experimental design

The study was conducted in 2018 at the Agriculture Research Station of the College of Agriculture and Natural Resources, University of Tehran, Karaj, Iran (35°48'32" N, 50°58'06" E, 1308 m a.s.l.). This area has semi-arid climate conditions with a mean annual temperature of 13.7 °C and precipitation of 245 mm (Fig. 1). Two fields under contrasting tillage management, conventional tillage (CT) and no-tillage (NT), were

Table 1

Soil properties of 0–10 and 10–20 cm soil depths in the fields in 2019.

Soil properties	Conventional tillage (CT)		No-tillage (NT)	
	0–10 cm	10–20 cm	0–10 cm	10–20 cm
pH	7.8 ± 0.1	7.7 ± 0.1	7.8 ± 0.0	7.6 ± 0.1
EC (ds m ⁻¹)	0.9 ± 0.0	0.7 ± 0.2	1 ± 0.0	0.7 ± 0.2
OC (mg/g)	0.9 ± 0.1	0.8 ± 0.2	1.2 ± 0.0	1.01 ± 0.1
TN (mg/g)	0.09 ± 0.01	0.07 ± 0.01	0.1 ± 0.0	0.08 ± 0.0
Avail K (mg kg ⁻¹)	167 ± 4	134 ± 3	279 ± 5.2	237.4 ± 6.0
Avail P (mg kg ⁻¹)	9 ± 0.2	9.5 ± 0.6	15 ± 1.5	15.3 ± 0.7
Bulk density (g cm ⁻³)	1.52	1.56	1.32	1.38
Total Porosity (%)	54 ± 2	52 ± 1.1	48 ± 1.9	45 ± 1
Sand (%)	57 ± 1	53 ± 1	28 ± 0.3	27 ± 0.7
Silt (%)	25 ± 1	28 ± 1	42 ± 0.2	42 ± 0.6
Clay (%)	18 ± 1	19 ± 2	30 ± 0.4	31 ± 0.7
Soil texture	Sandy Loam	Sandy Loam	Clay Loam	Clay Loam

EC: electrical conductivity, OC: organic carbon, TN: total nitrogen, Avail K: available potassium, Avail P: available phosphorus.

* Values are Mean ± Standard Deviation (n = 3).

selected. Both systems were under a wheat (*Triticum aestivum* L.) - corn (*Zea mays* L.) rotation for at least fifteen years before the study. Each field of 11 × 16 m was divided into nine 3 × 4 m sub-plots. The experimental design was a randomized complete block with three replicates per treatment. Table 1 summarizes the main physical and chemical soil properties. Ten soil samples were collected randomly using an auger sampler (8 cm diameter) from 0 to 10 cm and 10–20 cm depths in each field. Subsequently, soil samples were uniformly mixed to one composite sample per soil depth. Soil samples were air-dried, sieved (2 mm), and stored until analyzes were carried out. Soil texture was determined by the hydrometric method (Gee and Bauder, 1986). Soil organic carbon (SOC) using the Walkley and Black method (Walkley and Black, 1934), pH, and electrical conductivity (EC) were measured in saturated soil extracts (Richardes, 1954). Available phosphorus was determined using the NaHCO₃ method (Olsen and Sommers, 1982). The ammonium acetate method was used for determining available potassium (Knudsen et al., 1982). Total nitrogen (TN) was measured using the Kjeldahl method described by Bremner and Mulvaney (1982). Micro-nutrients were measured in DTPA extracts (Lindsay and Norvell, 1978) using atomic absorption spectrophotometry.

2.2. Treatment applications

The wheat residue was applied following wheat harvest (Jul 2018) in both NT and CT fields. The three levels of residue were 3.5 t ha⁻¹

Table 2
Elemental composition of wheat and corn residue.

Plant residue	N %	P	K	C	C/N
Wheat	*0.84 ± 0.12	0.09 ± 0.01	1.75 ± 0.07	55.4 ± 1.80	66 ± 2.56
Corn	0.92 ± 0.06	0.25 ± 0.02	1.08 ± 0.01	53.65 ± 1.30	58 ± 1.63

* Values are Mean ± standard deviation (n = 3).

(100 %, R₁₀₀), 1.75 t ha⁻¹ (50 %, R₅₀), and no residue (residue was completely removed, R₀). The selection of residue levels was based on crop yield, which was almost similar in both CT and NT systems. The plant residue was weighed according to the treatment and was then homogeneously and manually distributed over the surface of each plot. In the CT field, all the three residue levels were tilled and incorporated into the soil, while in the NT they were just left on the soil surface. For the residue removal treatment (R₀), under both CT and NT systems, the whole residue from the previous crop was removed from the plot, and the soil surface was left uncovered. After this, silage corn was planted by using a row crop planted at a dose of 35 kg ha⁻¹ with a distance of 15 cm within the row and 75 cm between the rows. In the NT field, seed placement was made using a planter with a single coulters to eliminate residue and loosen the soil before the standard planter unit. In the CT field, before seeding, the soil was plowed to a depth of 35 cm with a moldboard plow, disked, and leveled. Both NT and CT received equal

basal NPK fertilizers equal to 50 kg ha⁻¹ Urea, 70 kg ha⁻¹ Potassium sulfate, and 150 kg ha⁻¹ Superphosphate. Additional N was top-dressed at eight leaves (80 kg Urea ha⁻¹) and ten leaves (270 kg Urea ha⁻¹) stages. Plots were irrigated after cultivation and 7–10 days intervals thereafter using the sprinkler irrigation method. After harvesting the corn (Oct 2018), three levels of residue were applied equally to 1.8 (100 %, R₁₀₀), 0.9 ton ha⁻¹ (50 %, R₅₀), and no residue (0 %, R₀). Winter wheat was planted in Nov 2018, at a dose of 208 kg ha⁻¹ seed with a row distance of 13 cm by a mechanical drilling machine. Basal fertilization included the equivalents of 50 kg ha⁻¹ urea, 200 kg ha⁻¹ superphosphate, and 150 kg ha⁻¹ potassium sulfate. Additional N fertilizer was supplied during late tillering (110 kg Urea ha⁻¹), stem elongation (110 kg Urea ha⁻¹), and spiking (50 kg Urea ha⁻¹).

2.3. Plant residue analyses

A sub-sample of wheat and corn residues was dried and finely ground for chemical analyses (Table 2). Organic carbon (OC) was measured by wet oxidation (Walkley and Black, 1934), and the Kjeldahl method was used for determining total nitrogen (TN) content. For measurement of total phosphorus (P) and potassium (K), the samples were combusted at 500 °C, and P and K were determined using spectrophotometric and flame photometric methods, respectively.

2.4. Soil CO₂ efflux

Gas sampling was started in late July and early August 2018 in the

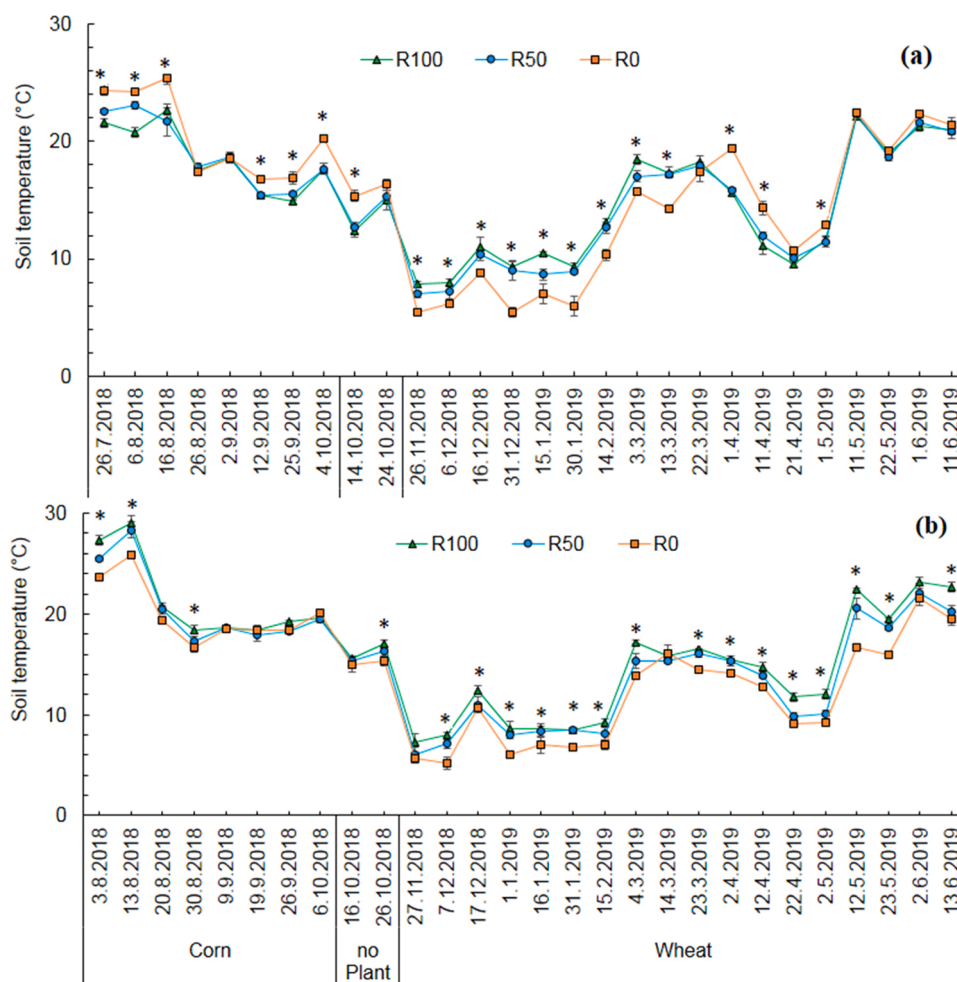


Fig. 2. The effect of residue levels on soil temperature at the gas sampling events under the no-tillage (NT) (a) and conventional tillage (CT) (b) systems. Bars represent standard error (n = 3). An asterisk (*) shows a significant difference between residue levels within a day at P < 0.05.

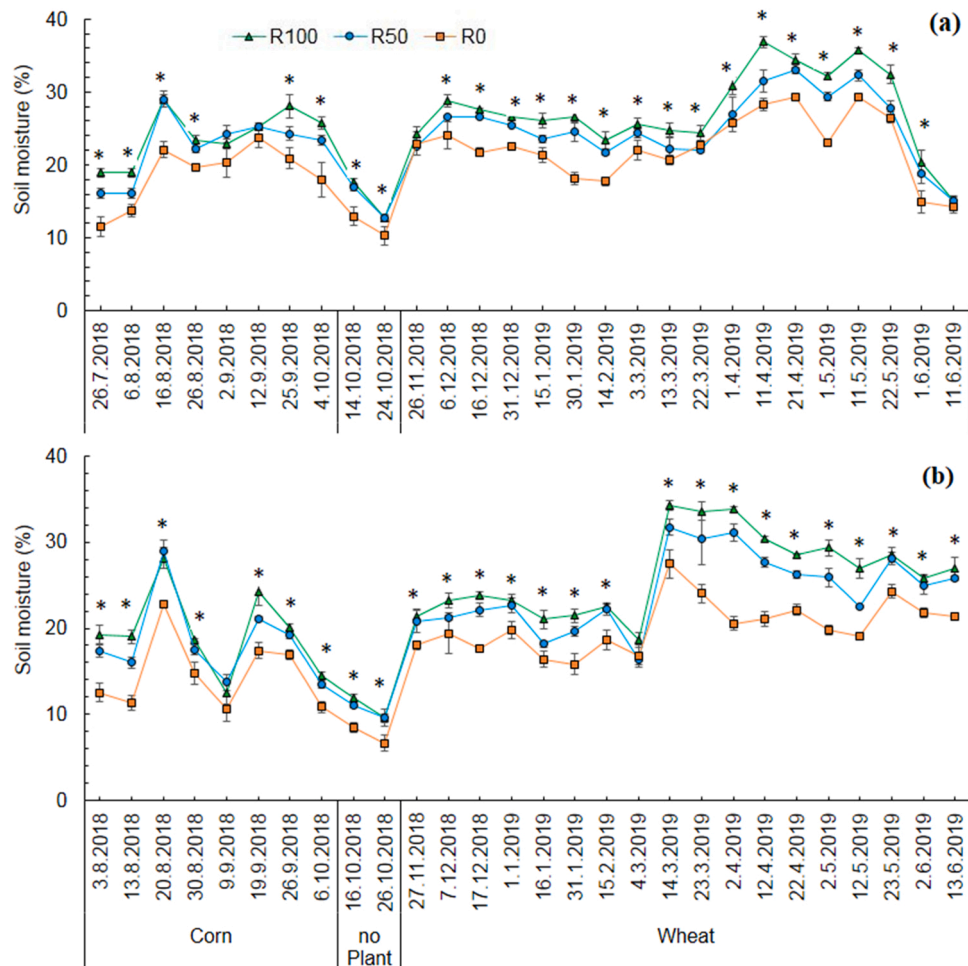


Fig. 3. The effect of residue levels on soil moisture at the gas sampling events under the no-tillage (NT) (a) and conventional tillage (CT) (b) systems. Bars represent standard error ($n = 3$). An asterisk (*) shows a significant difference between residue levels within a day at $P < 0.05$.

corn cropping system under the NT and CT systems respectively and ended in early June 2019 in the wheat cropping system under both systems. Gas sampling was performed at 7–10 d intervals generally, and every two weeks in winter (December–February), based on the GRACEnet Chamber-based Trace Gas Flux Measurement protocol (Parkin and Venterea, 2010). The static closed chamber method was used to measure the amount of CO_2 efflux from the soil. The method has been widely used for the measurement of soil gas efflux, including carbon dioxide, methane, and nitrous oxide (Oertel et al., 2012). Polyvinyl chloride (PVC) chambers (15 cm diameter, 12.5 cm height) with ports for gas sampling were gradually hammered into the soil surface to depths of 5 cm during the whole monitoring period. Chambers were only removed during cultivation events. To enable chamber operation, the crops were mowed when their height exceeded the chamber's height. Gas samples were collected at least 24 h after chamber installation to avoid disturbance effects. Gas sampling was performed from 9 to 10 am at 0, 30, and 60 min time points by inserting a needle attached to a 20 mL syringe in the sampling port and transferring in 12 mL pre-vacuumed vials sealed with butyl rubber septa (Labco Exetainer, UK). The concentration of CO_2 in the vials was measured using gas chromatography (Teif Gostar Faraz, TG 2552, Iran), equipped with a thermal conductivity detector (TCD).

The CO_2 efflux was calculated as the changes in linear concentration gradient over time and from the ratio between chamber volume and soil surface area (Liebig et al., 2010). The CO_2 efflux was converted to its carbon equivalent ($\text{CO}_2\text{-C}$) by multiplying it by the ratio of the molecular weight of carbon to that of carbon dioxide (12/44). Cumulative amounts

of CO_2 efflux were calculated using linearly interpolating data points and integrating the underlying area (Sainju et al., 2012; Wegner et al., 2018).

2.5. Ancillary measurements

To evaluate the relationship between soil conditions and CO_2 efflux, soil temperature, and moisture contents, as well as microbial biomass C, were determined at each gas sampling event. Soil temperature was measured with a thermometer at 10 cm depth next to the collars. Three soil samples (0–10 cm) were taken with a core sampler (100 cm^3 volume) from each plot that was then mixed into one composite sample. The soil moisture content of the samples was determined using oven-drying at 105 °C multiplied by soil bulk density. Soil bulk density was measured using stainless steel cylinders (100 cm^3 volume). Microbial biomass carbon (MBC) was determined by the chloroform fumigation extraction method (Jenkinson et al., 2004).

2.6. Statistical analyses

The data were analyzed using the Analysis of Repeated Measures procedure in general linear models (GLM) of SAS software version 9.4 (SAS Institute, Cary, NC, USA). Residue rate, tillage type, and sampling time were the fixed effects. Soil parameters (moisture, temperature, MBC) and CO_2 efflux were the repeated measure variables. Means for each residue rate, tillage type, and sampling time were compared by the Duncan method at the 0.05 probability level. Pearson linear correlation

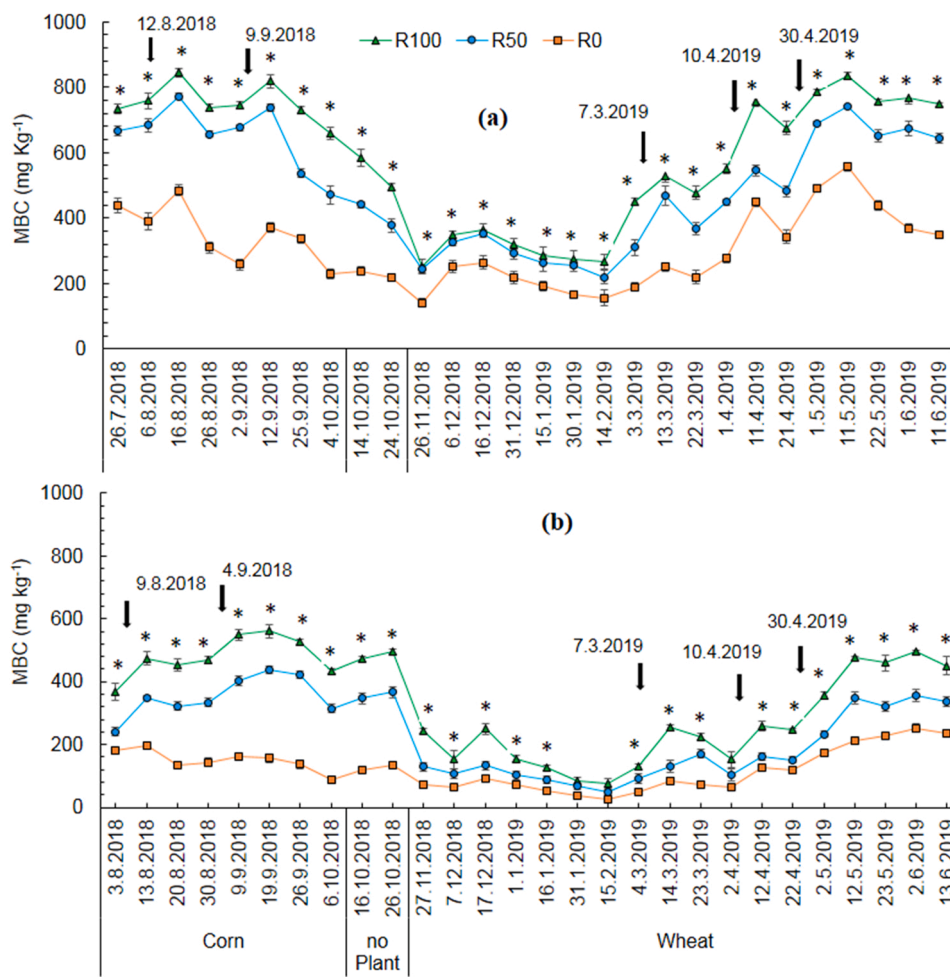


Fig. 4. The effect of residue levels on soil microbial biomass carbon (MBC) in no-tillage (NT) (a) and conventional tillage (CT) (b) systems at the gas sampling events. Bars represent standard error ($n = 3$). The bold black arrows indicate fertilization events. An asterisk (*) shows a significant difference between residue levels within a day at $P < 0.05$.

analysis was used to determine the relationships between soil parameters (moisture, temperature, MBC) and CO_2 efflux.

3. Results and discussion

3.1. Soil temperature and soil moisture

In the NT system, during the corn season (Jul–Oct), and also in the fallow period between corn and wheat seasons (14/10/2018 and 24/10/2018), soil temperature was higher in plots receiving residue at lower rates, where a significant increase ($P < 0.05$) in soil temperature was observed in the R_0 compared to R_{50} and R_{100} treatments (Fig. 2a). On the other hand, in most cases, no significant differences were observed between R_{100} and R_{50} . From the beginning of the wheat-growing season in autumn (26/11/2018) until the end of winter (22/3/2019), the opposite trend was noticed, where high residue rate led to increased soil temperature, and R_{100} and R_{50} resulted in a significant increase ($P < 0.05$) in soil temperature compared to R_0 . Here, in most cases, the differences between R_{100} and R_{50} were not significant. In addition, from the beginning of spring until the end of the experimental period, the soil temperature in R_0 treatment increased compared to R_{100} and R_{50} treatments. The results of this study are consistent with previous findings (Guzman et al., 2015). The increasing soil temperature in R_0 in summer and R_{100} and R_{50} in autumn and winter under the NT system can be attributed to the insulating effect of plant residue. The presence of plant residue on the soil surface in the NT system protects the soil

against severe thermal fluctuations and modulates surface radiation energy and thermal changes between soil and atmosphere (Horton et al., 1996). In the summer, plant residue on the soil surface prevents the absorption of more radiation, thus preventing the soil surface from warming. In autumn and winter, plant residue on the soil surface prevents heat exchange and, thus causes the soil to warm up compared to bare soil.

During crop rotation under the CT system, soil temperature increased with increasing residue rate at all sampling times, and, in most cases, these increases in R_{100} and R_{50} treatments were significant ($P < 0.05$) compared to R_0 . Still, in some cases, no significant difference was observed between R_{100} and R_{50} treatments (Fig. 2b). Results obtained in the CT system are congruent with Zhang et al. (2018) in their study in China. In the CT system, crop residue is incorporated into the soil instead of remaining on the soil surface, leaving a small amount of plant residue on the soil surface, which reduces the effect of crop residues and causes the soil to absorb more heat, and leads to increases in soil temperature (Dendooven et al., 2012).

In almost all sampling times, R_{100} and R_{50} treatments caused a significant increase ($P < 0.05$) in moisture compared to R_0 under both CT and NT systems (Fig. 3a,b). However, in several cases, no significant differences were observed between R_{100} and R_{50} . Increased soil moisture in R_{100} and R_{50} treatments compared to R_0 under both CT and NT systems could be explained by the fact that crop residue reduces moisture evaporation and increases water-holding capacity and soil water content. Previous studies have also suggested that crop residue increases

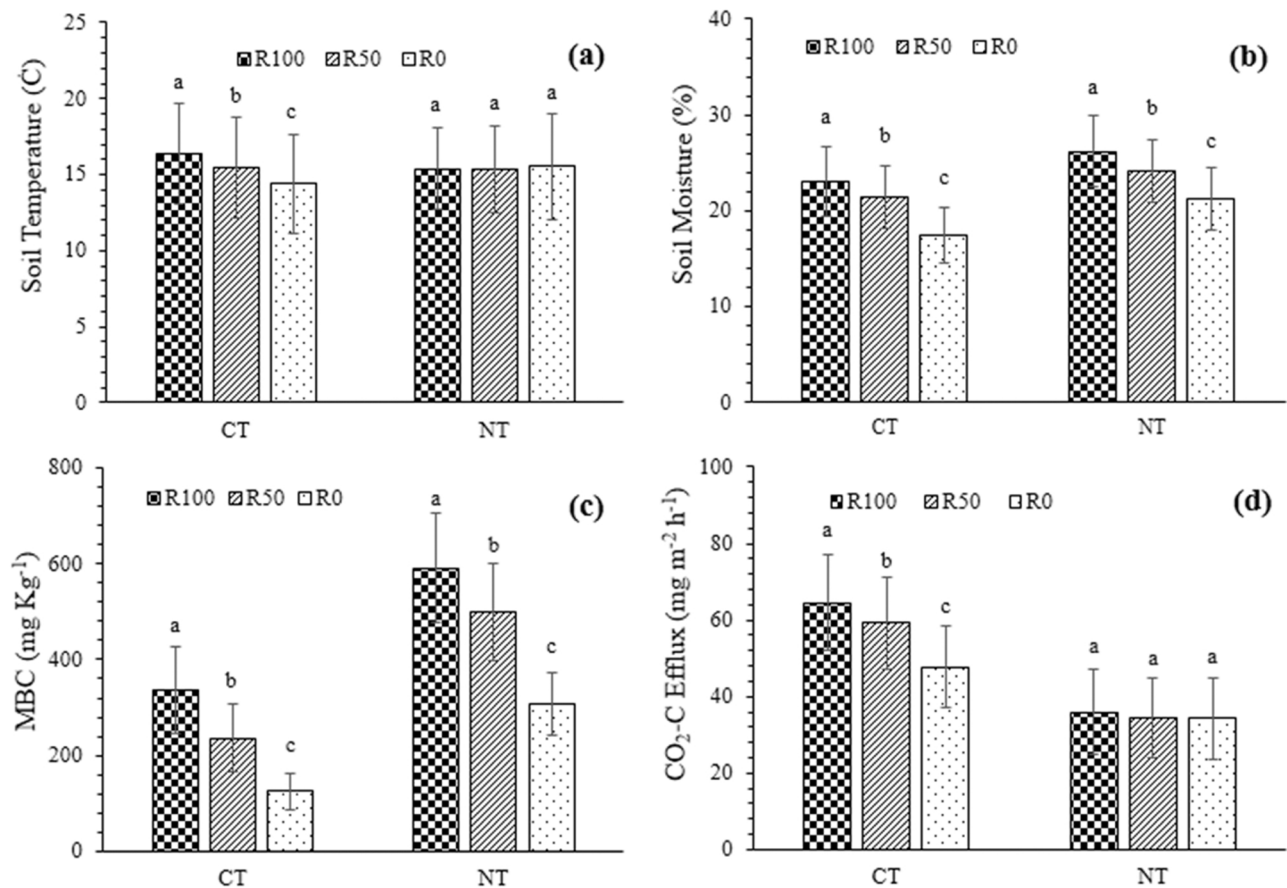


Fig. 5. The effect of crop residue rates on annual mean soil temperature (a) moisture (b), microbial biomass carbon (MBC) (c), and CO₂-C efflux (d) during corn-wheat rotation under conventional tillage (CT) and no-tillage (NT) systems. Means with the same letter are not significantly different. Bars represent standard error (n = 3).

soil water content (Sainju et al., 2012; Chalise et al., 2019) as well as other types of cover crops or mulches (Rodrigo-Comino et al., 2020b).

3.2. Soil microbial biomass carbon

There were significant differences for microbial biomass carbon (MBC) across all treatments of residue rates during the rotation cycle under both tillage systems (Fig. 4a,b). MBC was the highest in summer during the corn season; and declined during the cold season and again increased with rising soil temperature. In both CT and NT systems, MBC was positively correlated with the level of crop residue, i.e., the annual means of 335, 236, and 125 (mg kg⁻¹) in R₁₀₀, R₅₀, and R₀ under CT, and 590, 500, and 300 (mg kg⁻¹) for R₁₀₀, R₅₀, and R₀ under NT (Fig. 5c). The addition of N resulted in microbial growth, as measured by MBC, in both CT and NT, consistent with the literature (Stewart et al., 2018; Mgelwa et al., 2019). The soil microbial biomass pool is one of the biological properties that strongly regulates the dynamics of nutrients and C in soil (Toosi et al., 2012). It has also been shown that microbial biomass responds more rapidly to plant residue management compared to the total soil OM pool (Yang et al., 2012). Our findings are in line with previous studies, indicating that increasing the level of crop residues positively affects the MBC pool size (Chowdhury et al., 2015; Li et al., 2017). These could be due to the elevated MBC in soils covered by crop residue, namely enhanced soil moisture content, and also a supply of soluble C (and other nutrients) derived from the decomposing residue (Yang et al., 2012). The lower rate of MBC during the cold season could be attributed to the lower soil temperature during this period. A significant and positive correlation between soil temperature and MBC also confirms this (Table S1). Low soil temperature may limit soil microbial

activity and their population. Temperature and moisture are the most important factors influencing soil microbial biomass (Babur and Dindaroglu, 2020).

3.3. CO₂ efflux

CO₂ efflux for all residue rates followed a similar pattern for all residue levels in both NT and CT (Fig. 6a,b). Higher soil CO₂ efflux during summer and spring is attributed to higher soil temperature (Fig. 2a,b), and as a result, enhanced biological (microbial, faunal, and root) activity. Seasonal fluctuations in soil temperature and their influence on CO₂ (and other GHGs) efflux from the soil are well documented (Schaufler et al., 2010).

The CO₂-C efflux rate decreased with increasing residue retention at all sampling dates during the corn season in summer (26/7/2018–4/10/2018) under the NT system, where soil respiration was significantly higher ($P < 0.05$) under R₀ than R₁₀₀ and R₅₀ treatments (Fig. 6a). A higher intensity of soil respiration in uncovered soil has also been reported previously (Guzman et al., 2015). A key reason for the significant increase in soil CO₂ efflux in R₀ compared to R₁₀₀ and R₅₀ treatments during the corn season under the NT system appears to be greater fluctuations in soil temperature (Fig. 2a). A positive and significant correlation between soil temperature and CO₂ efflux also confirms this (Table 3). Soil temperature is the main factor influencing the emission of CO₂ from the soil (Dendooven et al., 2012). Yin et al. (2016) and Wegner et al. (2018) also reported a positive correlation between CO₂ emissions and soil temperature. In addition, another reason for the increasing CO₂ efflux in R₀ compared to R₁₀₀ and R₅₀ can be possibly attributed to the high C/N ratio (66) of wheat residues (Table 2) applied before the corn

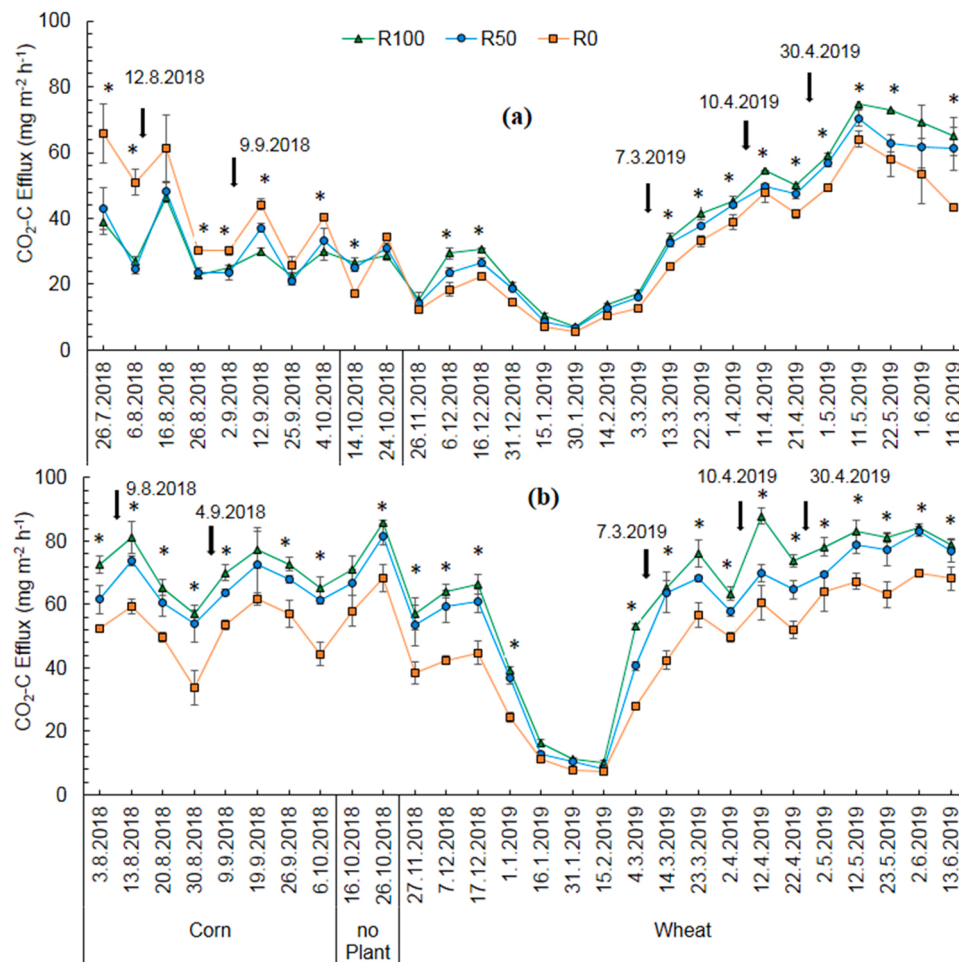


Fig. 6. Residue rate effect on daily soil CO₂-C efflux at the time of greenhouse gas measurement in corn-wheat rotation under no-tillage (NT) (a) and conventional tillage (CT) (b) systems. Bars represent standard error. The bold black arrows indicate fertilization events. An asterisk (*) shows a significant difference between residue levels within a day at $P < 0.05$.

growing season, which reduces the decomposition rate and release of CO₂. It has also been reported that the C/N ratio of an organic residue applied to the soil mostly determines the rate of its mineralization (Nguyen and Marschner, 2016).

When residue was added, there was no difference in soil respiration between R₁₀₀ and R₅₀. In contrast, from the start of the wheat season until the end of the experiment, CO₂ efflux increased with an increasing residue retention rate at all sampling events compared to R₀. However, the differences among the residue treatments were not significant from early to late winter in 2019, which could be related to lower soil temperature and limited microbial activity and root respiration during this period. In the spring of 2019, significant increases ($P < 0.05$) were recorded for R₁₀₀ and R₅₀ compared to R₀ on several sampling dates, especially after urea fertilization events. A significant increase in CO₂ efflux after fertilization activities could be attributed to improved plant growth, photosynthesis, and respiration (Holou, 2010), stimulating and increasing the mineralization of crop residue (Dendooven et al., 2012; Sainju et al., 2012). In the CT system, the CO₂-C efflux rate increased with increasing residue rate, and these increases were significant ($P < 0.05$) at almost all sampling times in R₁₀₀ and R₅₀ treatments compared to R₀, but the significant differences ($P < 0.05$) between R₁₀₀ and R₅₀ treatments were only noticed in a few sampling dates, mainly after urea fertilization events in spring 2019, and for the most of the dates the differences were not significant (Fig. 6b).

The significant increase in CO₂ efflux in the R₁₀₀ and R₅₀ compared to the R₀, which was observed in both CT and NT during the spring, can

be explained by a substantial input from decomposing residue that serves as a substrate for soil microbes (Toosi et al., 2017).

Several studies have reported that the incorporation of plant residue increases soil CO₂ emissions (Badía et al., 2013; Wang et al., 2019). Meanwhile, increased plant residue mineralization, soil organic carbon, microbial biomass carbon, and dissolved organic carbon accounted for increasing CO₂ emissions (Zhao et al., 2014; Ding et al., 2017; Yang et al., 2017). However, other studies reported reduced soil CO₂ emissions due to the retention of plant residue in the soil (Bai et al., 2017).

3.4. Cumulative CO₂ efflux

The annual cumulative efflux of CO₂-C did not differ significantly between residue rate treatments under the NT system. Still, the cumulative efflux of CO₂-C was higher ($p < 0.05$) in the CT system where R₁₀₀ (4.65 Mg CO₂-C ha⁻¹ y⁻¹) and R₅₀ (4.26 Mg CO₂-C ha⁻¹ y⁻¹) represented 36.5 % and 25 % greater cumulative CO₂-C efflux, respectively, than R₀ (3.4 Mg CO₂-C ha⁻¹ y⁻¹) (Fig. 7). Similarly, Drury et al. (2021) reported that corn residue minimally affected CO₂ efflux in soils under the NT. Still, in the CT system, full retention of corn residue increased soil respiration up to 47 %. Contrasting results on the effect of residue in soil respiration (e.g., Wegner et al., 2018) are due to the complex interaction of different factors such as residue amount and quality, as well as soil and environmental conditions.

The strong positive relationship between cumulative CO₂ efflux and the crop residue ($R^2 = 0.96$ for CT and $R^2 = 0.9$ for NT) suggests the

Table 3

Correlations of CO₂ with selected soil properties for different rates of plant residues under conventional tillage (CT) and no-tillage (NT) systems.

Tillage system	Residue rate	Soil parameters	CO ₂ emission	
			Pearson correlation (R)	P-value
CT	R100	Temperature	0.61**	0.001
		Moisture	0.13	0.49
		Microbial biomass carbon	0.68**	.000
	R50	Temperature	0.49**	0.001
		Moisture	0.11	0.46
		Microbial biomass carbon	0.69**	.000
	R0	Temperature	0.60**	0.001
		Moisture	0.00	0.99
		Microbial biomass carbon	0.78**	.000
NT	R100	Temperature	0.46*	0.01
		Moisture	0.17	0.36
		Microbial biomass carbon	0.66**	.000
	R50	Temperature	0.43**	0.004
		Moisture	0.12	0.44
		Microbial biomass carbon	0.70**	.000
	R0	Temperature	0.79**	.000
		Moisture	0.06	0.73
		Microbial biomass carbon	0.89**	.000

** . Correlation is significant at the 0.01 level.

* . Correlation is significant at the 0.05 level.

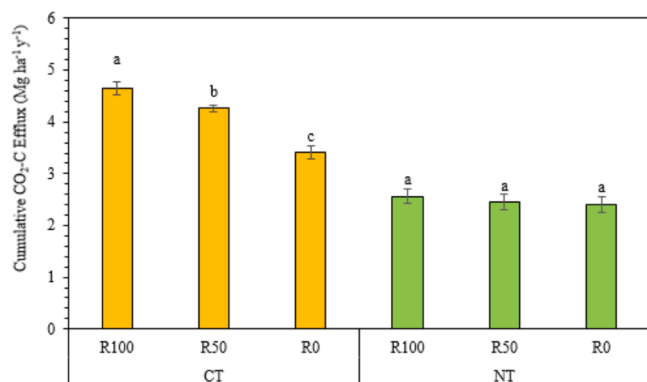


Fig. 7. Residue rate effect on annual cumulative CO₂-C (Mg ha⁻¹ y⁻¹) efflux under no-tillage (NT) and conventional tillage (CT) systems. Means with different letters in each tillage system indicate significant differences (P < 0.05). Bars represent standard error (n = 3).

contribution of the plant residue to soil respiration (Fig. 8).

In the CT system, when calculated, an additional 1.25 Mg C ha⁻¹ was respired from the R₁₀₀ than from the no residue soil, equivalent to 43 % of the total (2.9 Mg C ha⁻¹) of the added residue (Table 4). For the R₅₀, this was equivalent to 59 % of the added residue (1.44 Mg C ha⁻¹). Under the NT system, these values were 5.5 %, and about 3 % of total C added from crop residue in R₁₀₀ and R₅₀, respectively, compared to that respired from the no residue treatment. This indicates that under both CT and NT systems, there was C accumulation in the soil following the addition of crop residue. However, it has been well documented that in the long term, the vast majority of the added C from residue will respire to the atmosphere and only a small fraction of residual C will be incorporated into soil OM (Kuzyakov, 2006; Campos et al., 2011).

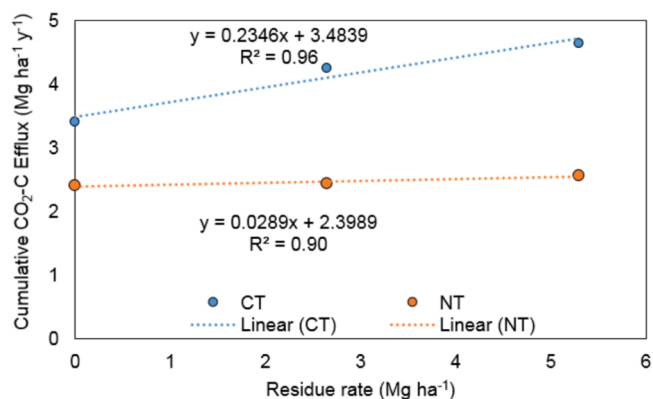


Fig. 8. Linear regression between residue rates and cumulative CO₂-C efflux per soil management system. CT: conventional tillage; NT: no-tillage. The points on each CT and NT line represent the annual cumulative CO₂-C efflux for different rates of residue (corn and wheat) during the whole corn-wheat rotation cycle.

Table 4

Crop residue carbon application rates (dry weight basis) in corn-wheat rotation under each conventional tillage (CT) and no-tillage (NT) system.

Crop residue rate (%)	Wheat residue (Mg C ha ⁻¹)	Corn residue (Mg C ha ⁻¹)	Total (Mg C ha ⁻¹)
100	*1.93	0.96	2.90
50	0.96	0.48	1.44
0	0	0	0

* Crop residue carbon application rates (dry weight basis) were based on crop yield and almost similar under both CT and NT systems.

Table 5

The effect of tillage on mean soil CO₂ emissions, soil temperature, soil moisture, microbial biomass carbon, and cumulative CO₂ emissions.

Tillage system	Mean CO ₂ emissions (mg m ⁻² h ⁻¹)	Temperature (°C)	Moisture (%)	Microbial biomass carbon (mg kg ⁻¹)	Cumulative CO ₂ emissions (Mg ha ⁻¹ y ⁻¹)
CT	*57.15 ± 12.36a	15.41 ± 3.30a	20.62 ± 3.52b	232.05 ± 85.80b	4.10 ± 0.55a
NT	34.86 ± 10.71b	15.40 ± 3.01a	23.89 ± 1.82a	465.96 ± 117.93a	2.47 ± 0.10b

Means with the same letter in each column are not significantly different.

^a Mean ± standard deviation, n = 3.

3.5. The effect of tillage and interaction effects of residue rate and tillage system on soil properties (temperature, moisture, MBC), and CO₂ efflux

The effect of tillage and the interaction effects of residue rate and tillage system on soil temperature was not significant (Table 5 and Fig. 9a). More significant increases (P < 0.05) in mean and cumulative CO₂ efflux were observed in the CT system than in the NT system (Table 5). Soil moisture and microbial biomass carbon were significantly higher in the NT system compared to the CT system.

In addition, significant (P < 0.05) interaction effects of residue rate and tillage system were observed for soil moisture, soil microbial biomass carbon, and mean and cumulative CO₂ efflux (Fig. 9b-e). NT₁₀₀ resulted in the highest values of soil moisture (26 %) and microbial biomass carbon (591 mg kg⁻¹) among the treatments, while the lowest value of soil moisture (17.5 %) and microbial biomass carbon (124.5 mg kg⁻¹) were obtained from CT₀ (Fig. 9b,c). The highest values of mean CO₂ efflux (64.5 mg m⁻² h⁻¹) and cumulative CO₂ efflux (4.65 Mg ha⁻¹ y⁻¹) were observed in CT₁₀₀ (Fig. 9d,e), whereas the

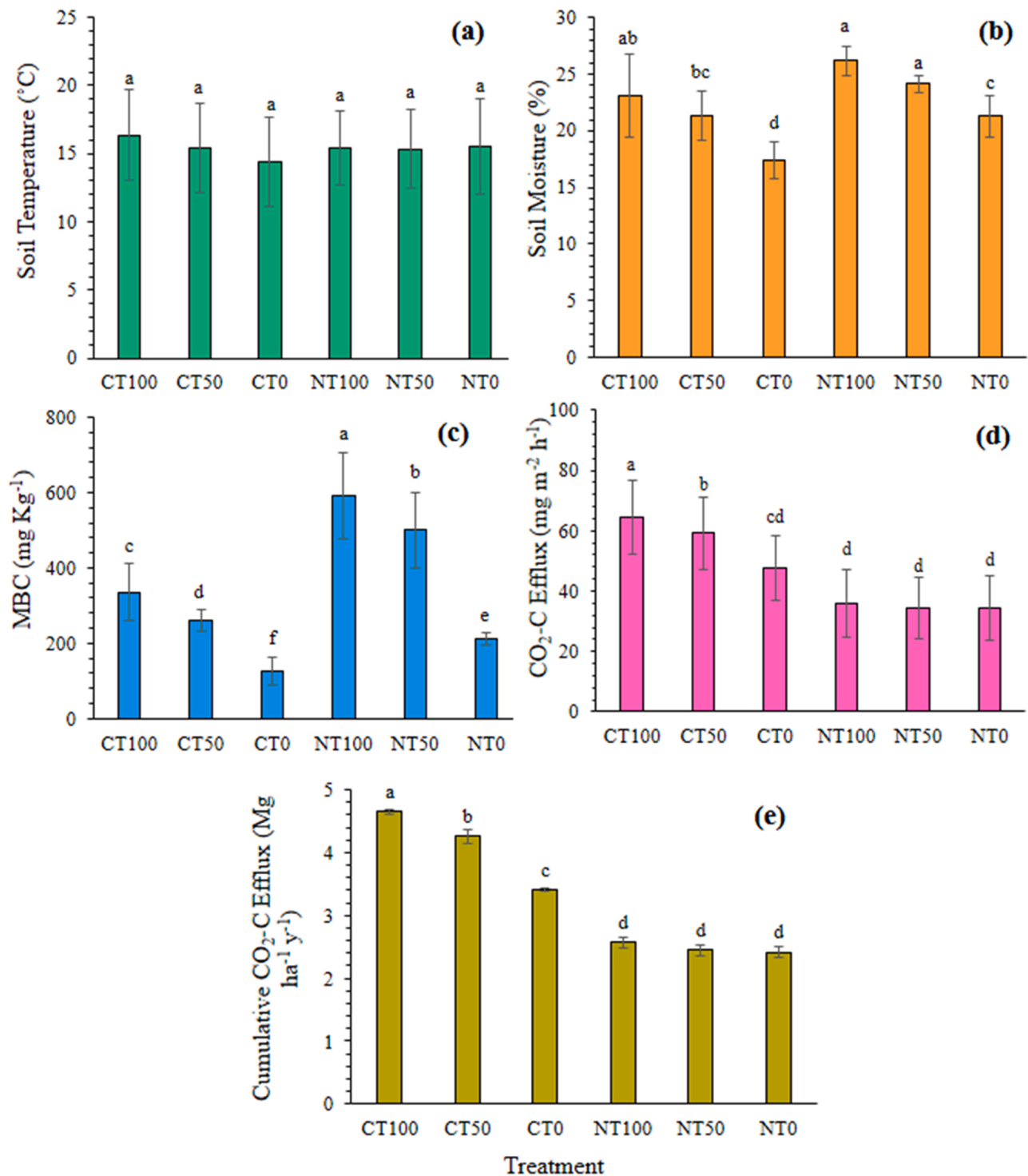


Fig. 9. The interaction effect of residue rate and tillage system on annual mean soil temperature (a), moisture (b), microbial biomass carbon (MBC) (c), CO₂-C efflux (d), and cumulative CO₂-C efflux (e). Means with the same letter are not significantly different. Bars represent standard error (n = 3).

lowest values of mean CO₂ emissions (34 mg m⁻² h⁻¹) and cumulative CO₂ efflux (2.4 Mg ha⁻¹ y⁻¹) were found in NT₀ (Fig. 9d,e). Furthermore, no significant differences were found for cumulative CO₂ efflux among NT₁₀₀, NT₅₀, and NT₀ (Fig. 9e). In the NT system, during the corn cropping season CO₂ efflux was higher in the R₀ compared to R₁₀₀ and R₅₀, while during the wheat season, a reverse trend was observed in which the rate of CO₂ efflux increased in R₁₀₀ and R₅₀ compared to R₀ (Fig. 6a). By considering the whole corn-wheat rotation, it is possible that these two different trends neutralized the effect of each other and

led to no significant differences among the residue treatments. Also, under conservation tillage systems including the NT system, crop residue is not incorporated into the soil and they are not sufficiently fragmented and decomposed, which consequently reduces the real potential of crop residue management on CO₂ efflux. In agreement with our findings, Drury et al. (2021) reported no significant effect of different crop residue removal rates on CO₂ emissions under the NT system. Also, Teixeira et al. (2013) found no large differences in CO₂ emissions from plots with or without crop residues under rotary tillage.

There are several reasons for higher CO₂ efflux in the CT system than in the NT system. Tillage intensifies the oxidation of soil organic carbon and leads to the release of large amounts of carbon dioxide (La Scala Jr et al., 2006). Also, the incorporation of crop residue into the soil under conventional tillage practices accelerates the decomposition of crop residues (Lal, 2010). Furthermore, tillage operations break down soil aggregates and cause faster decomposition of soil OM (Mitchell et al., 2019). Reduction of tillage intensity using modern conservation tillage methods, including the no-tillage method, has positive effects on reducing soil organic matter mineralization and reducing CO₂ efflux (Snyder et al., 2009). Higher soil moisture and microbial biomass carbon in the NT system with full retention of crop residue (NT₁₀₀) could be related to the fact that NT accompanied by higher residue retention increases organic inputs into the soil and reduces soil disturbance, which consequently favors greater microbial biomass carbon and enzyme activities (Choudhary et al., 2018). Additionally, greater infiltration, reduced evaporation, and soil protection from rainfall impact can account for the highest soil moisture content under NT₁₀₀ (Šarauskis et al., 2009).

4. Conclusions

In this study, the effect of crop residue at three rates - 100 %, 50 %, and no residue (0 %) - on soil CO₂ efflux under conventional and no-tillage systems was investigated. The results showed that soil CO₂ efflux was responsive to the application of residue mainly by alterations of soil temperature and microbial biomass carbon. Higher cumulative CO₂ efflux coincided with the addition of a higher crop residue rate under the CT system when R₁₀₀ and R₅₀ significantly increased the cumulative CO₂ efflux compared to R₀. Under the NT system, residue treatments had minimal impact on cumulative CO₂ efflux, and no significant differences were observed among residue treatments. Also, CO₂ efflux was significantly higher under the CT system compared to the NT system. Among the treatments, conventional tillage with the full incorporation of crop residue (CT₁₀₀) led to the highest amount of CO₂ efflux, while the lowest amount of CO₂ efflux was observed in no-tillage without residue (NT₀), but no significant differences were found between NT₁₀₀, NT₅₀, and NT₀.

In conclusion, the findings of this study show that in this region or similar semi-arid environments, the agricultural conservation practice of no-tillage with full retention of crop residue (100 %) is a potential tool for mitigating CO₂ emissions and, consequently, global warming. This research can contribute to soil conservation on a practical level, offering valuable information which can improve agricultural management, and hence the efficiency of food production. The achievement of these positive effects depends, however, on the incorporation of the results in strategies of territorial planning.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supporting information

Supplementary data associated with this article can be found in the online version at [doi:10.1016/j.pedobi.2022.150819](https://doi.org/10.1016/j.pedobi.2022.150819).

References

- Badía, D., Martí, C., Aguirre, A.J., 2013. Straw management effects on CO₂ efflux and C storage in different Mediterranean agricultural soils. *Sci. Total Environ.* 464, 233–239.
- Bai, Y., Hai, J., Gang, Y., Jing, H., Dong, F., Liu, M., Nie, X., Aamp, N., 2017. Effect of the straw returning pattern on soil respiration and enzyme activities. *Chin. J. Appl. Environ. Biol.* 23, 28–32.
- Babur, E., Dindaroglu, T., 2020. Seasonal changes of soil organic carbon and microbial biomass carbon in different forest ecosystems. *Environ. Factors Affect. Hum. Health* 1, 1–21.
- Borken, W., Matzner, E., 2009. Reappraisal of drying and wetting effects on C and N mineralization and fluxes in soils. *Glob. Change Biol.* 15 (4), 808–824.
- Bremner, J., Mulvaney, C., 1982. Nitrogen-total. In: Page, al (ed.), *Methods of Soil Analysis. Part 2: Chemical and Microbiological Properties*. Am. Soc. Agron., Madison, WI, USA, pp. 595–624.
- Campbell, B., Chen, L., Dygert, C., Dick, W., 2014. Tillage and crop rotation impacts on greenhouse gas fluxes from soil at two long-term agronomic experimental sites in Ohio. *J. Soil Water Conserv.* 69 (6), 543–552.
- Campos, B.H.C.D., Amado, T.J.C., Tornquist, C.G., Nicoloso, R.D.S., Fiorin, J.E., 2011. Long-term C-CO₂ emissions and carbon crop residue mineralization in an oxisol under different tillage and crop rotation systems. *Rev. Bras. Ciênc. Solo* 35 (3), 819–832.
- Chalise, K.S., Singh, S., Wegner, B.R., Kumar, S., Pérez-Gutiérrez, J.D., Osborne, S.L., Nleya, T., Guzman, J., Rohila, J.S., 2019. Cover crops and returning residue impact on soil organic carbon, bulk density, penetration resistance, water retention, infiltration, and soybean yield. *Agron. J.* 111 (1), 99–108.
- Choudhary, M., Datta, A., Jat, H.S., Yadav, A.K., Gathala, M.K., Sapkota, T.B., Das, A.K., Sharma, P.C., Jat, M.L., Singh, R., et al., 2018. Changes in soil biology under conservation agriculture based sustainable intensification of cereal systems in Indo-Gangetic Plains. *Geoderma* 313, 193–204.
- Chowdhury, S., Farrell, M., Butler, G., Bolan, N., 2015. Assessing the effect of crop residue removal on soil organic carbon storage and microbial activity in a no-till cropping system. *Soil Use Manag.* 31 (4), 440–460.
- Ciais, P.C., Sabine, G., Bala, L., Bopp, V., Brovkin, J., Canadell, A., Chhabra, R., Defries, J., Galloway, M., Heimann, C., et al., 2013. Carbon and other biogeochemical cycles. In: Stocker, T.F., Qin, D., Plattner, G., Tignor, M., Allen, S.K., Boschung, J., Nauels, A., Xia, Y., Bex, V., Midgley, P.M. (Eds.), *Climate Change 2013: The Physical Science Basis*. Cambridge Univ. Press, Cambridge, New York, pp. 465–570.
- Delgado-Baquerizo, M., Eldridge, D., Maestre, F., Karunaratne, S., Trivedi, P., Reich, P., Singh, B., 2017. Climate legacies drive global soil carbon stocks in terrestrial ecosystems. *Sci. Adv.* 3, 1–7.
- Dendooven, L., Patiño-Zúñiga, L., Verhulst, N., Luna-Guido, M., Marsch, R., Govaerts, B., 2012. Global warming potential of agricultural systems with contrasting tillage and residue management in the central highlands of Mexico. *Agric. Ecosyst. Environ.* 142, 40–48.
- Ding, R., Wang, W., Zhang, Q., 2017. Effect of straw mulching on soil respiration and its' temperature sensitivity under different crop rotation systems. *Chin. J. Eco-Agric.* 25 (8), 1106–1118.
- Dong, W., Liu, E., Wang, J., Yan, C., Li, J., Zhang, Y., 2017. Impact of tillage management on the short-and long-term soil carbon dioxide emissions in the dryland of Loess Plateau in China. *Geoderma* 307, 38–44.
- Drury, C.F., Woodley, A.L., Reynolds, W.D., Yang, X.M., Phillips, L.A., Rehmann, L., Calder, W., 2021. Impacts of corn stover removal on carbon dioxide and nitrous oxide emissions. *Soil Sci. Soc. Am. J.* 85 (5), 1334–1348.
- Gee, G.W., Bauder, J.W., 1986. Particle-size analysis. In: Klute, A. (Ed.), *Methods of Soil Analysis: Physical and Mineralogical Methods. Part 1, 2nd ed.* Soil Science Society of America, Madison, Wisconsin, United States of America, pp. 383–411.
- Guzman, J., Al-Kaisi, M., Parkin, T., 2015. Greenhouse gas emissions dynamics as influenced by corn residue removal in continuous corn system. *Soil Sci. Soc. Am. J.* 79 (2), 612–625.
- Hibbard, K., Law, B., Reichstein, M., 2005. An analysis of soil respiration across northern hemisphere temperate ecosystems. *Biogeochemistry* 73, 29–70.
- Holou, R.A.Y. 2010. Nutrient, Water, and Soil Type Management of Biofuel Feedstock Production by Corn, Sorghum, and Switchgrass (Doctoral dissertation). University of Missouri–Columbia.
- Horton, R., Bristow, K.L., Kluitenberg, G.J., Sauer, T.J., 1996. Crop residue effects on surface radiation and energy balance. *Theor. Appl. Clim.* 54 (1), 27–37.
- Jenkinson, D.S., Brookes, P.C., Powlson, D.S., 2004. Measuring soil microbial biomass. *Soil Biol. Biochem.* 36 (1), 5–7.
- Knudsen, D., Peterson, G., Pratt, P., 1982. *Methods of Soil Analysis. Chemical and Microbiological Properties*. Ame. Soc. of Agron, and Soil. Sci. of Ame, Inc. Madison, Wisconsin, USA, pp. 229–30.
- Kuzayakov, Y., 2006. Sources of CO₂ efflux from soil and review of partitioning methods. *Soil Biol. Biochem.* 38, 425–448.
- La Scala Jr, N., Bolonhezi, D., Pereira, G.T., 2006. Short-term soil CO₂ emission after conventional and reduced tillage of a no-till sugar cane area in southern Brazil. *Soil Res.* 91 (1–2), 244–248.
- Lal, R., 2010. Managing soils and ecosystems for mitigating anthropogenic carbon emissions and advancing global food security. *BioScience* 6 (9), 708–721.
- Lal, R., 2004. Soil carbon sequestration impacts on global climate change and food security. *Science* 304, 1623–1627.
- Li, S., Chen, J., Shi, J., Tian, X., Li, X., Li, Y., Zhao, H., 2017. Impact of straw return on soil carbon indices, enzyme activity, and grain production. *Soil Sci. Soc. Am. J.* 81 (6), 1474–1484.

- Liebig, M.A., Tanaka, D.L., Gross, J.R., 2010. Fallow effects on soil carbon and greenhouse gas flux in central North Dakota. *Soil Sci. Soc. Am. J.* 74 (2), 348–364.
- Lindsay, W.L., Norvell, W.A., 1978. Development of a DTPA soil test for zinc, iron, manganese, and copper. *Soil. Sci. Soc. Am. J.* 42 (3), 421–428.
- Mgelwa, A.S., Hu, Y.L., Xu, W.B., Ge, Z.Q., Yu, T.W., 2019. Soil carbon and nitrogen availability are key determinants of soil microbial biomass and respiration in forests along urbanized rivers of southern China. *Urban For. Urban Green.* 43, 126341.
- Mirzaei, M., Gorji Anari, M., Razavy-Toosi, E., Asadi, H., Moghiseh, E., Saronjic, N., Rodrigo-Comino, J., 2021. Preliminary effects of crop residue management on soil quality and crop production under different soil management regimes in corn-wheat rotation systems. *Agronomy* 11 (2), 302.
- Mitchell, J.P., Reicosky, D.C., Kueneman, E.A., Fisher, J., Beck, D., 2019. Conservation agriculture systems. *CAB Rev.* 14 (001), 1–25.
- Nawaz, A., Lal, R., Shrestha, R.K., Farooq, M., 2017. Mulching affects soil properties and greenhouse gas emissions under long-term no-till and plough-till systems in alfisol of Central Ohio. *Land Degrad. Dev.* 28 (2), 673–681.
- Nguyen, T.T., Marschner, P., 2016. Soil respiration, microbial biomass and nutrient availability in soil after repeated addition of low and high C/N plant residues. *Biol. Fertil. Soils* 52, 165–176.
- Oertel, C., Herklotz, K., Matschullat, J., Zimmermann, F., 2012. Nitric oxide emissions from soils: a case study with temperate soils from Saxony, Germany. *Environ. Earth Sci.* 66 (8), 2343–2341.
- Oertel, C., Matschullat, J., Zurba, K., Zimmermann, F., Erasmi, S., 2016. Greenhouse gas emissions from soils—a review. *Chem. Erde-Geochem.* 76 (3), 327–352.
- Olsen, S.R., Sommers, L.E., 1982. Phosphorus. In: Page, A.L. (Ed.), *Methods of Soil Analysis, Chemical and Microbiological Properties. Part 2.* American Society of Agronomy, Soil Science Society of America, Madison, Wisconsin, pp. 403–430.
- Parkin, T.B., Venterea, R.T., 2010. USDA-ARS GRACEnet project protocols, chapter 3. Chamber-based trace gas flux measurements. Sampling protocols. Beltsville, MD, pp. 1–39.
- Parkin, T.B., Kaspar, T.C., 2003. Temperature controls on diurnal carbon dioxide flux: implications for estimating soil carbon loss. *Soil Sci. Soc. Am. J.* 67, 1763–1772.
- Richardes, L.A., 1954. *Diagnosis and Improvement of Saline and Alkali Soils.* United States Salinity Laboratory Staff. Agriculture Handbook 60. United States Department of Agriculture. 160p.
- Rodrigo-Comino, J., López-Vicente, M., Kumar, V., Rodríguez-Seijo, A., Valkó, O., Rojas, C., Pourghasemi, H.R., Salvati, L., Bakr, N., Vaudour, E., et al., 2020a. Soil science challenges in a new era: a transdisciplinary overview of relevant topics. *Air Soil Water Res.* 13, 1178622120977491.
- Rodrigo-Comino, J., Terol, E., Mora, G., Gimenez-Morera, A., Cerdà, A., 2020b. Vicia sativa Roth. can reduce soil and water losses in recently planted vineyards (*Vitis vinifera* L.). *Earth Syst. Environ.* 4 (4), 827–842.
- Sainju, U.M., Stevens, W.B., Caesar-TonThat, T., Liebig, M.A., 2012. Soil greenhouse gas emissions affected by irrigation, tillage, crop rotation, and nitrogen fertilization. *J. Environ. Qual.* 41 (6), 1774–1786.
- Sainju, U.M., Jabro, J.D., Stevens, W.B., 2008. Soil carbon dioxide emission and carbon sequestration as influenced by irrigation, tillage, cropping system, and nitrogen fertilization. *J. Environ. Qual.* 37, 98–106.
- Sarauskiš, E., Romaneckas, K., Buragienė, S., 2009. Impact of conventional and sustainable soil tillage and sowing technologies on physical-mechanical soil properties. *Environ. Res. Engine Manag.* 49 (3), 36–43.
- Scharlemann, J.P.W., Tanner, E.V.J., Hiederer, R., Kapos, V., 2014. Global soil carbon: understanding and managing the largest terrestrial carbon pool. *Carbon Manag.* 5, 81–91.
- Schauler, G., Kitzle, B., Schindlbacher, A., Skiba, U., Sutton, M.A., Zechmeister-Boltenstern, S., 2010. Greenhouse gas emissions from European soils under different land use: effects of soil moisture and temperature. *Eu. J. Soil Sci.* 61 (5), 683–696.
- Smit, P., 2004. Soils as carbon sinks: the global context. *Soil Use Manag.* 20, 212–218.
- Snyder, C.S., Bruulsema, T.W., Jensen, T.L., Fixen, P.E., 2009. Review of greenhouse gas emissions from crop production systems and fertilizer management effects. *Agric. Ecosyst. Environ.* 133 (3–4), 247–266.
- Stewart, C.E., Roosaendaal, D.L., Manter, D.K., Delgado, J.A., Del Grosso, S., 2018. Interactions of stover and nitrogen management on soil microbial community and labile carbon under irrigated no-till corn. *Soil Sci. Soc. Am. J.* 82 (2), 323–331.
- Sugasti, L., Pinzón, R., 2020. First approach of abiotic drivers of soil CO₂ efflux in Barro Colorado Island, Panama. *Air Soil Water Res.* 13, 1178622120960096.
- Tanhua, T., Lauvset, S.K., Lange, N., et al., 2021. A vision for FAIR ocean data products. *Commun. Earth Environ.* 2 (1), 1–4.
- Teixeira, L.G., Corradi, M.M., Fukuda, A., Panosso, A.R., Reicosky, D., Lopes, A., 2013. Soil and crop residue CO₂-C emission under tillage systems in sugarcane-producing areas of southern Brazil. *Sci. Agric.* 70 (5), 327–335.
- Toosi, E.R., Kravchenko, A.N., Guber, A.K., Rivers, M.L., 2017. Pore characteristics regulate priming and fate of carbon from plant residue. *Soil Biol. Biochem.* 113, 219–230.
- Toosi, E.R., Doane, T.A., Horwath, W.R., 2012. Abiotic solubilization of soil organic matter, a less-seen aspect of dissolved organic matter production. *Soil Biol. Biochem.* 40, 12–21.
- Walkley, A., Black, I.A., 1934. An examination of the Degtjareff method for determining soil organic matter, and a proposed modification of the chromic acid titration method. *Soil Sci.* 37 (1), 29–38.
- Wang, W., Akhta, K., Ren, G., Yang, G., Feng, Y., Yuan, L., 2019. Impact of straw management on seasonal soil carbon dioxide emissions, soil water content, and temperature in a semi-arid region of China. *Sci. Total Environ.* 642, 471–482.
- Wegner, B.R., Chalise, K.S., Singh, S., Lai, L., Abagandura, G.O., Kumar, S., Osborne, S.L., Lehman, R.M., Jagadamma, S., 2018. Response of soil surface greenhouse gas fluxes to crop residue removal and cover crops under a corn-soybean rotation. *J. Environ. Qual.* 47 (4) (1146–1144).
- Yang, X., Meng, J., Lan, Y., Chen, W., Yang, T., Yuan, J., Liu, S., Han, J., 2017. Effects of maize stover and its biochar on soil CO₂ emissions and labile organic carbon fractions in northeast China. *Agric. Ecosyst. Environ.* 240, 24–31.
- Yang, X.Y., Ren, W.D., Sun, B.H., Zhang, S.L., 2012. Effects of contrasting soil management regimes on total and labile soil organic carbon fractions in a loess soil in China. *Geoderma* 177–178.
- Yin, S., Zhang, X., Pumpanen, J., Shen, G., Xiong, F., Liu, C., 2016. Seasonal variation in soil greenhouse gas emissions at three age-stages of dawn redwood (*Metasequoia glyptostroboides*) stands in an Alluvial Island, Eastern China. *Forests* 7 (11), 256.
- Zhang, S., Wang, Y., Shen, Q., 2018. Influence of straw amendment on soil physicochemical properties and crop yield on a consecutive Mollisol slope in Northeastern China. *Water* 10 (4), 449.
- Zhao, Y., Xue, Z., Guo, H., Mu, X., Li, C., 2014. Effects of tillage and crop residue management on soil respiration and its mechanism. *Trans. Chin. Soc. Agric. Eng.* 30, 144–164.