



doi:10.1093/aobpla/plaa037 Advance Access Publication July 25, 2020

## **STUDIES**

SPECIAL ISSUE: THE ECOLOGY AND GENETICS OF POPULATION DIFFERENTIATION IN PLANTS

# Temporal migration rates affect the genetic structure of populations in the biennial Erysimum mediohispanicum with reproductive asynchrony

A. Jesús Muñoz-Pajares<sup>1,2,\*,†</sup>, Mohamed Abdelaziz<sup>1,3,†</sup> and F. Xavier Picó<sup>4</sup>

<sup>1</sup>Departamento de Genética, Universidad de Granada, 18071 Granada, Spain, <sup>2</sup>Research Center in Biodiversity and Genetic Resources (CIBIO), Campus Agrário de Vairão, 4485-661 Vairão, Portugal, <sup>3</sup>Biological and Environmental Sciences, School of Natural Sciences, University of Stirling, Stirling FK9 4LA, UK, <sup>4</sup>Departamento de Ecología Integrativa, Estación Biológica de Doñana (EBD), Consejo Superior de Investigaciones Científicas (CSIC), 41092 Sevilla, Spain

\*Corresponding author's e-mail address: ajesusmp@ugr.es

†These authors contributed equally to this work.

Associate Editor: Antonio Castilla

# **Abstract**

Migration is a process with important implications for the genetic structure of populations. However, there is an aspect of migration seldom investigated in plants: migration between temporally isolated groups of individuals within the same geographic population. The genetic implications of temporal migration can be particularly relevant for semelparous organisms, which are those that reproduce only once in a lifetime after a certain period of growth. In this case, reproductive asynchrony in individuals of the same population generates demes of individuals differing in their developmental stage (non-reproductive and reproductive). These demes are connected by temporal migrants, that is, individuals that become annually asynchronous with respect to the rest of individuals of their same deme. Here, we investigated the extent of temporal migration and its effects on temporal genetic structure in the biennial plant Erysimum mediohispanicum. To this end, we conducted two independent complementary approaches. First, we empirically estimated temporal migration rates and temporal genetic structure in four populations of E. mediohispanicum during three consecutive years using nuclear microsatellites markers. Second, we developed a demographic genetic simulation model to assess genetic structure for different migration scenarios differing in temporal migration rates and their occurrence probabilities. We hypothesized that genetic structure decreased with increasing temporal migration rates due to the homogenizing effect of migration. Empirical and modelling results were consistent and indicated a U-shape relationship between genetic structure and temporal migration rates. Overall, they indicated the existence of temporal genetic structure and that such genetic structure indeed decreased with increasing temporal migration rates. However, genetic structure increased again at high temporal migration rates. The results shed light into the effects of reproductive asynchrony on important population genetic parameters. Our study contributes to unravel the complexity of some processes that may account for genetic diversity and genetic structure of natural populations.

**Keywords:** Biennials; *Erysimum mediohispanicum*; genetic diversity; genetic structure; nuclear microsatellites; simulation model; temporal migration rate.

# Introduction

Population is a concept with several definitions that still generates vivid discussions in biological sciences (Jonckers 1973; Waples and Gaggiotti 2006; Hey 2011). Intuitively, we can describe a population as a group of co-occurring, interbreeding individuals that pass their genetic features on to the next generation. A common characteristic of any population is that it inexorably exhibits variation in the number of individuals contributing genetically to the next generation (i.e. the effective population size). This attribute determines spatio-temporal variation in allele frequency and shapes genetic structure (Wright 1951, 1965; Kimura and Weiss 1964; Rosenberg et al. 2002). The most common usage of genetic structure data deals with the demographic and evolutionary history of populations across wide geographic areas, which allows the detection of major large-scale genetic patterns. In other words, the role that population dynamics, migration, adaptive variation and genetic drift plays in determining the extent of genetic structure across space (Loveless and Hamrick 1984; Slatkin 1985, 1987; Rosenberg et al. 2002; Manel et al. 2003; Storfer et al. 2007; Dionne et al. 2008).

The mentioned evolutionary and demographic processes structure populations spatially but also temporally. Recent studies on annuals comparing population genetic attributes in different points in time provided insight into how fast and how much plant populations can modify their genetic composition and genetic structure over several generations (Franks et al. 2007, 2016; Franks and Weis 2008; Nevo et al. 2012; Sultan et al. 2013; Bustos-Segura et al. 2014; Thomann et al. 2015; Welt et al. 2015; Horgan-Kobelski et al. 2016; Kuester et al. 2016; Frachon et al. 2017; Gómez et al. 2018). Such a temporal dimension in the process of genetic differentiation in plant populations can even acquire a higher degree of complexity when considering variation in key life-history traits, such as reproduction at the withinpopulation scale (Wells and Wells 1980; Loveless and Hamrick 1984). For example, a decline in the mate probability among individuals of the same population varying in reproductive time may lead to isolation by time. In this case, individuals showing early and late reproductive times would be the extremes of a reproductive time gradient, which should exhibit the largest genetic differences (Hendry and Day 2005). Individuals with intermediate reproductive time values within this gradient would increase gene flow between the above-mentioned extremes. However, a gradient in reproduction times is not the only mechanism to generate assortative mating. This is because seed banks may also generate similar consequences on genetic structure by incorporating to the mating arena individuals from previous generations with different reproductive times (Nunney 2002; Vitalis et al. 2004; Dolan et al. 2008; Königer et al. 2012; Falahati-Anbaran et al. 2011, 2014).

Temporal variation in reproductive time will surely be more important in organisms requiring >1 year between reproductive periods. Among them, biennial species are the simplest example. Such organisms are able to show asynchrony at larger time scales (e.g. between years) in addition to the within-year asynchrony described above for isolation by time. Consequences of this reproductive asynchrony will depend on how heritable reproduction times are (Hendry et al. 1999; Fox 2003; Weis and Kossler 2004). Thus, in the case of complete heritability, individuals inhabiting a given geographic location are subdivided into groups of mating (demes hereafter) that would never inbreed among them, favouring the assortative mating and thus stratifying populations (Fox 2003; Weis and Kossler 2004). In the case in which reproduction time shows a significant environmental component, some individuals from a given deme may behave, for example, as triennial instead of as biennial, contributing to gene flow between demes. In that sense, individuals belonging to one deme but reproducing with another would act as temporal migrants reducing deme differentiation (Fig. 1).

In this context of temporal genetic structure, individuals from different demes of the same population co-occur in the same physical site, and developmental variation in reproductionand probably microhabitat conditions (Denney et al. 2020) determines the amount of temporal migrants. Just as with geographic migration, demes within a population may structure over time if migration rates among demes remain low. In the case of semelparous organisms, those that reproduce only once in a lifetime after a certain period of growth, the effects of temporal variation in reproduction between demes can be especially dramatic because individuals will have no more chances to pass on their genes after their unique reproductive event. According to their life histories, we could expect reproductive individuals of biennial semelparous organisms to be detected either in odd or even years. However, reproductive individuals commonly occur every year in populations of most biennial semelparous species (Kelly 1985; De Jong and Klinkhamer 1988; Petrů 2005; Valverde et al. 2016). The evolutionary consequences of temporal asynchrony on population differentiation has widely been studied in animals, particularly in salmonids (Aspinwall 1974; Waples 1990a, b, 2002, 2006), but examples in plant species are rather scarce (Wells and Wells 1980; Valverde et al. 2016). So far, most of the attention paid to temporal asynchrony in plants focused on early and late flowering time, the role of seed banks and their effects on assortative mating, population structure and speciation (Hossaert-McKey et al. 1996; Heath et al. 2002; Chung et al. 2003; Fox 2003; Weis and Kossler 2004; Hendry and Day 2005; Kent et al. 2007; Martínez-Cruz et al. 2007; Lundemo et al. 2009; Riccioni et al. 2010; Falahati-Anbaran et al. 2011, 2014; Gomaa et al. 2011; Sloop et al. 2011; Königer et al. 2012).

Here, we quantified the consequences of temporal migration for genetic structure of plant populations over time. To this end, we conducted two independent complementary approaches using a monocarpic plant with a biennial life history. First, we collected data from natural populations of the biennial plant Erysimum mediohispanicum to estimate temporal migration rates and temporal genetic structure. We estimated temporal migration rates by monitoring the fate of 1-year-old vegetative rosettes from E. mediohispanicum populations over several years. To assess temporal genetic structure, we genotyped with polymorphic nuclear microsatellites (Muñoz-Pajares et al. 2011) vegetative and reproductive E. mediohispanicum individuals, representing the two different demes co-occurring at each population over years. Second, we developed a simulation model to explore the population dynamics and population genetics of a biennial plant and analyse the effects of varying temporal migration rates on population genetic differentiation between two demes. In particular, the aim of this work is to address the following questions: (i) Do demes within a geographic population show significant genetic differentiation? (ii) Does temporal migration rate reduce between-demes differentiation? Overall, the results highlight the intricacy of ecological and biological factors accounting for the genetic structure of plant populations.

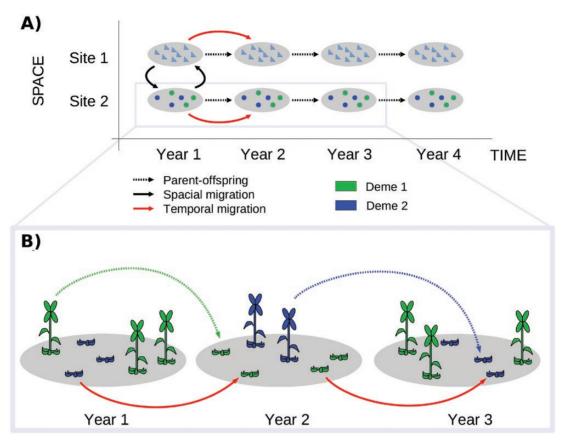


Figure 1. Graphical representation of the conceptual study system: a biennial organism's life cycle. (A) Depiction of two geographically separated populations over 4 years. The genetic properties of every site and year depend on the previous generation through parent-offspring inheritance (dashed lines), the spatial migration from one geographic site to another (solid black line) and the temporal migration from 1 year to another within the same geographic site (solid red line). (B) Details on temporal migration. Individuals in a single population flowering in odd and even years (green and blue, respectively) are defined as different demes. Reproductive plants of deme #1 in year #1 produce rosettes in year #2 (dashed line). Vegetative rosettes in year #1 have two options: they can become reproductive in year #2 (thus contributing to deme #2) or remain as vegetative until year #3 (thus representing a temporal migrant and contributing to deme #1, solid red line).

## **Methods**

## Empirical demographic and genetic data

In this study, we obtained empirical data from E. mediohispanicum (Brassicaceae), which is an endemic but common herb of the Iberian Peninsula exhibiting a biennial habit (Fig. 2). The plant can usually grow for 2-3 years as vegetative rosettes before dying after reproduction (Gómez 2005). Hence, E. mediohispanicum exhibits variation in age at first reproduction, which is the trait required to generate temporal migrants among demes as defined above. Erysimum mediohispanicum has a mixed mating system and a wide array of insects differing in morphology, size and behaviour (Gómez et al. 2007) pollinating the plant. Wind is the only known seed disperser and no seed bank has been reported for this species (Gómez 2007).

We collected data from four large E. mediohispanicum sites at Sierra Nevada National Park (SE Spain), separated by 0.33-1.45 km and coded as Em01, Em08, Em23 and Em25 (see Gómez et al. 2007 for further details of study sites). In June-July 2010, 2011 and 2012, we collected leaf tissue from 30-40 vegetative rosettes (entire above-ground rosettes) and 30-34 reproductive individuals (rosette and caulinar leaves) at each sampling site. In Em25 in 2012, we did not find vegetative rosettes and only reproductive individuals represented this sampling site for this particular year. Every year, samples were immediately stored in

cellophane bags, labelled and dried with Silica Gel until DNA extraction in the laboratory.

During field campaigns, we haphazardly tagged with numbered metal plates 100-120 vegetative rosettes at each sampling site per year to estimate the proportion of rosettes that completed the life cycle in 2 years. We estimated the proportion of rosettes that required >2 years to become reproductive individuals, representing an estimation of the temporal migration rate between demes. All tagged vegetative rosettes were of similar size (~40-50 mm diameter and 10-15 leaves). We purposely discarded too small or too large rosettes to avoid merging individuals from different demes. Based on previous field observations on E. mediohispanicum, we considered that all tagged rosettes of similar size probably belonged to the same deme germinated in summer/autumn the year before. In other words, all tagged vegetative rosettes were 1-year-old individuals. We surveyed all tagged rosettes 1 year later to record their status including survival and their developmental stage (vegetative or reproductive). Overall, we tagged and monitored 807 E. mediohispanicum vegetative rosettes.

In addition, we collected leaf material from 30 individuals per deme to quantify genetic diversity and differentiation. Because our sampling design includes 3 years, four populations and two demes per population, we collected tissue from 720 individuals. We extracted DNA from all samples using DNeasy Plant Mini

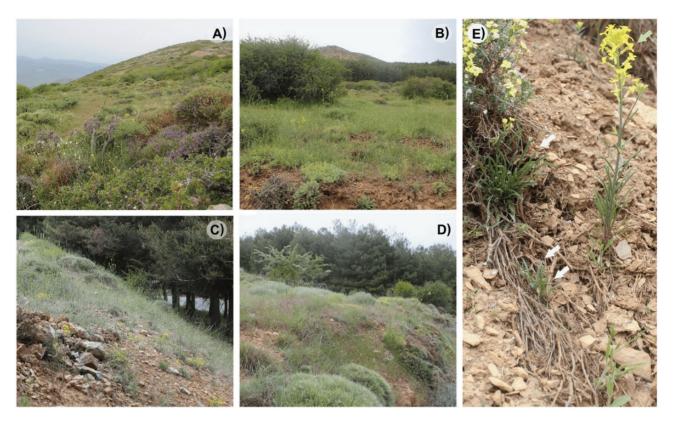


Figure 2. Panoramic view of the populations included in the present study. (A) Em25, 2064 m a.s.l.; (B) Em23, 1874 m a.s.l.; (C) Em01, 1750 m a.s.l.; (D) Em08, 1690 m a.s.l. Flowering plants of E. mediohispanicum can be observed in every picture. (E) Erysimum mediohispanicum flowering plants (even year, green arrow) co-occurring with vegetative plants (odd year, white arrows). Note the different rosette sizes showed by the vegetative plants in the picture.

Kit (Qiagen, Hilden, Germany) and stored them at -80 °C until genotyping with nine nuclear microsatellite markers (codes: C5, D2, D4, D10, D11, E4, E5, E6 and E8). These markers previously showed polymorphism in E. mediohispanicum and other genetic properties appropriate for population genetic studies (Muñoz-Pajares et al. 2011). Polymerase chain reactions (PCRs) were performed in 25  $\mu L$  of reaction mixture containing 2.5  $\mu L$  of Buffer (10×), 1  $\mu$ L of MgCl<sub>2</sub> (50 mM), 0.2  $\mu$ L of dNTPs (25 mM), 1  $\mu$ L of forward and reverse primers (50 ng  $\mu L^{-1}$ ), 5  $\mu L$  of DNA (2 ng  $\mu L^{-1}$ ), 0.15  $\mu L$  of Taq polymerase (5U  $\mu L^{-1}$ ) and 15.15  $\mu L$  of Milli-Q water. Forward primers were fluorescently tagged with 6-FAM™ (C5, E5 and E8), NED™ (D4, D10 and E6) and VIC® (D2, D11 and E4) labels (Applied Biosystems, Foster City, CA, USA). Polymerase chain reactions were performed in a Biometra Gradient Thermal Cycler (Biometra, Göttingen, Germany). Thermocycler conditions consisted of an initial 2-min denaturation at 94 °C followed by 35 cycles of denaturing for 1 min at 94 °C, annealing for 1 min at primer-specific annealing temperatures, extension for 2 min at 72 °C and a final extension step of 10 min at 72 °C. Annealing temperatures were 54 °C for D2 and E5, and 56 °C for the rest of microsatellites. The fluorescently tagged PCR products were sized using GeneScan™ LIZ 500 size standard (Applied Biosystems, Foster City, CA, USA) on an ABI 3130 Genetic Analyser (Applied Biosystems, Foster City, CA, USA). Alleles were called by the same person with Peak Scanner v.1.0 (Applied Biosystems, Foster City, CA, USA).

We found some E. mediohispanicum individuals with more than two alleles at few particular markers, probably because of partial genome duplication. This finding is in accordance with the genome size variation previously described for this species (Muñoz-Pajares et al. 2018). Specifically, the extent of

duplication was rather low in this set of individuals as only 3-19 individuals per marker (0.4-2.7 %) had three or four alleles per microsatellite. In these particular cases, we selected the two most common alleles within the population among the different alleles observed per marker and individual. By doing so, we kept missing values low and avoided overestimation of low frequency alleles in our data set.

We successfully genotyped 228, 239 and 220 E. mediohispanicum individuals in 2010, 2011 and 2012, respectively, totalling 687 individuals. All individuals included in this study amplified successfully in at least six of nine microsatellites. Overall, our data set included 3.4 % of missing values. We genotyped twice 33 and 21 individuals from 2010 and 2011, respectively, with all microsatellites to estimate the genotyping error per locus that was of 0.1 %. Repeated genotyping indicated that genotyping error was due to allelic dropout in heterozygous loci, as usual in microsatellite genotyping (Hoffman and Amos 2005; Pompanon et al. 2005).

Genetic data for each site, year and developmental stage (group of reproductive and vegetative individuals) were analysed with FSTAT v.2.9.3 (Goudet 1995) and GenAlEx v.6 (Peakall and Smouse 2006) to calculate the mean number of alleles per locus  $(n_a)$ , mean observed heterozygosity  $(H_0)$  and mean gene diversity (H<sub>s</sub>). To estimate the proportion of the total genetic variation explained by each of these levels we performed a hierarchical analysis of molecular variance as implemented in the 'hierfstat' package in R (Goudet 2013). The significance level of the hierarchical F-statistics was calculated from 1000 bootstrap permutations.

To evaluate the temporal consistency of the genetic structure found between demes within a population, we performed hierarchical analyses of molecular variance separately for different years. We estimated genetic differentiation between reproductive and vegetative groups of individuals for each site independently for 2010, 2011 and 2012 using Arlequin v.3.5.1.2 (Excoffier and Lischer 2010). All  $F_{\rm ST}$  statistics (Weir and Cockerham 1984) and their significance levels were calculated from 1000 permutations.

The genetic structure among all 687 E. mediohispanicum individuals was estimated using the Bayesian clustering method implemented in STRUCTURE v.2.3.4 (Pritchard et al. 2000; Falush et al. 2003). All individuals were genetically different from each other, so they were non-redundant multilocus genotypes. Each run consisted of 50 000 burn-in Markov Chain Monte Carlo (MCMC) iterations and 150 000 MCMC after-burning repetitions for parameter estimation. To determine the K number of ancestral populations and the ancestry membership proportions of each individual, we run the algorithm 15 times for each K value from 1 to 15 and apply the Evanno method (Evanno et al. 2005) as implemented in the Structure Harvester Server (http://taylor0.biology.ucla.edu/structureHarvester/). We also applied the same procedure for every population separately, ranging K value from 1 to 6 in this case.

Genetic distance between demes was estimated based on the Discriminant Analysis of Principal Components (DAPC; Jombart et al. 2010), available in the R package adegenet v.2.1.1 (Jombart 2008), after retaining the first 50 principal component dimensions and the first two discriminant dimensions. Euclidean distances between centroids of demes per population and year indicated the genetic distance between them and therefore represented another measure of genetic structure due to temporal migration.

#### The demographic genetic model

A time lag of 2 years between seed germination and sexual reproduction characterizes the lifespan of biennial plants. In populations of a biennial plant, both vegetative and reproductive individuals can occur simultaneously. Obviously, vegetative and reproductive individuals cannot interbreed because of their different developmental stages. Reproductive individuals come from seeds germinated 2 years back, whereas vegetative individuals come from seeds germinated just 1 year back. Hence, these populations encompass two distinct demes of reproductive individuals separated temporally by a single year (Fig. 1). Some individuals can show departures from this pattern, such as variation in age at first reproduction, representing migration events between demes. Ecological and genetic factors underlying reproductive asynchrony determine the occurrence of temporal migration events. Depending on these factors, migration events in natural populations may occur every year or be an extremely rare event. To explore if the evolutionary consequences of both scenarios are equivalent, we split temporal migration into two components. First, we defined temporal migration rate as the proportion of individuals from a given deme and year reproducing in the other deme 1 year later. Second, we define temporal migration event probability as the frequency at which a migration event occurs over time. Based on this, strict biennials have no temporal migration rates or no temporal migration event probabilities, whereas facultative biennials typically exhibit temporal migration rates and temporal migration event probabilities higher than zero.

We developed a basic demographic model to compute total deme size over time  $(N_T)$  taking into account the total number of reproductive individuals recruited into a deme, via sexual reproduction  $(N_R)$  or immigration  $(N_M)$ , and the number

of individuals emigrating  $(N_{\scriptscriptstyle E})$  from that deme. This basic demographic model has the following structure:

$$N_T = N_R + N_M - N_E \tag{1}$$

This equation was applied to each of the two demes d1 and d2. For each deme, reproductive events occurred every 2 years in a way that when one deme is in a vegetative stage the other is reproducing, and vice versa. Hence, the tth generation occurred at time y for deme d1 and at time z for deme d2, which was described as:

$$y = 2t - 1 \tag{2}$$

$$z = 2t \tag{3}$$

For each of the two demes d1 and d2, the number of individuals recruited by sexual reproduction,  $N_{\rm R}$ , in years y and z, respectively, was computed as follows:

$$N_R(y, d1) = \lambda(y, d1) \times N_T(y - 2, d1)$$
 (4)

$$N_R(z, d2) = \lambda(z, d2) \times N_T(z - 2, d2)$$
 (5)

where  $\lambda(y,\,d1)$  and  $\lambda(z,\,d2)$  are deme growth rates for deme d1 in year y and deme d2 in year z, respectively. Such  $\lambda$  values simulated random fluctuations of deme size sampling from a normal distribution with mean 1 and standard deviation 0.01. Although such distribution parameter values are arbitrary, they generated a stable population with slight fluctuations over time. We performed previous simulations with the same mean and different standard deviations to confirm the expected effect of increasing variation on increasing extinction rates (results not shown). Our model only explored the genetic effects of temporal isolation and migration between demes within a demographically stable population.

For each of the two demes d1 and d2, the number of emigrating individuals,  $N_{\rm E}$ , in years y and z, respectively, was computed as follows:

$$N_E(y, d1) = m_{12}(y) \times N_R(y, d1)$$
 (6)

$$N_E(z, d2) = m_{21}(z) \times N_R(z, d2)$$
 (7)

where  $m_{12}(y)$  and  $m_{21}(z)$  is the migration rate of vegetative individuals from deme d1 to deme d2 in year y, and from deme d2 to deme d1 in year z, respectively.

Finally, for each of the two demes d1 and d2, the number of recruited individuals via immigration,  $N_{\rm M}$ , in years y and z, respectively, was computed as follows:

$$N_{\rm M}(y,d1) = m_{21}(z-2) \times N_{\rm R}(z-2,d2)$$
 (8)

$$N_{\rm M}(z,d2) = m_{12}(y) \times N_{\rm R}(y,d1)$$
 (9)

Note that because we only have two demes, emigrating individuals from one deme in a given year corresponded to immigrating individuals to the other deme right in the next year. Given the definition of reproductive year for deme d1 and d2 (Equations 2 and 3), immigrants of d2 in year z were emigrants from d1 in year y (Equation 9). In contrast, immigrants of d1 in year y were emigrants from d2 in year d2 in year d2 (Equation 8) [see Supporting Information—Table S1].

For each deme, the model ran for 100 generations. A detailed exploration of model performance at each generation indicated that initial conditions did not affect the model, and that 100 generations were enough to accurately estimate all genetic parameters of interest [see Supporting Information—Fig. S1]. Model simulations encompassed 10 different scenarios with constant migration rates ranging from 0 to 90 % of individuals migrating between demes. For each migration rate, the model

generated 11 additional scenarios simulating different migration event probabilities (from 0 to 1) over the 100 generations. The no-migration scenario represented the baseline simulation and showed the demographic behaviour of the simulated population. This simulation schedule was repeated 100 times totalling 11 000 runs (10 migration rates  $\times$  11 migration event probabilities  $\times$  100 replicates). The model started with 1000 initial individuals from each deme. Simulated individuals can only interbreed either when they are 2 years old or when they are 3 years old if they have become temporal migrants.

Each individual possessed an array of 100 unlinked diallelic loci that are at Hardy-Weinberg equilibrium at the start of the simulation. For each deme and generation, the model calculated allelic frequencies for each locus. Assuming panmixia, for each deme the model obtained genotypes for individuals of a given generation sampling from a binomial distribution based on allelic frequencies from the previous generation. At generation 100, the model generated an input file for the program Arlequin v.3.5.1.2 (Excoffier and Lischer 2010) including individual genotypes for each deme. We estimated gene diversity (H<sub>s</sub>) and genetic differentiation between demes and among individuals within demes with a hierarchical analysis of molecular variance (AMOVA).  $F_{ST}$  statistics (Weir and Cockerham 1984) and their significance were calculated from 1000 permutations. We computed mean  $F_{\rm ST}$  values for genetic differentiation between demes and the proportion of significant P-values for each F<sub>ST</sub> value using the 100 replicates for each simulation scenario including all migration rates and migration event probabilities. We used Euclidean distances between centroids of DAPC to estimate genetic distance between demes, as described above for empirical data.

The model was implemented in R (R Development Core Team 2011) and the script is available in the Supplementary Material (see below).

## **Results**

## **Empirical** estimates

The demographic survey conducted on four E. mediohispanicum sites between July 2010 and June 2012 allowed the estimation of temporal migration rates that varied from a low of 15 % to a high of 67 % (Table 1). On average, mortality rate of vegetative individuals was  $36.52 \pm 9.76$  % in 2010–11 and  $53.95 \pm 13.20$  % in 2011–12 (Table 1).

On average, the mean number of alleles per locus was 9.53  $\pm$  0.59 (range  $n_{\rm a}$  = 7.89–10.56 alleles per locus; Table 2), observed heterozygosity was 0.57  $\pm$  0.04 (range  $H_{\rm o}$  = 0.48–0.65; Table 2) and mean expected heterozygosity was 0.75  $\pm$  0.02 (range  $H_{\rm o}$  = 0.72–0.78; Table 2).

Hierarchical analysis of molecular variance indicated that sites were not significantly differentiated (Table 3). In contrast, year and developmental stages were significantly differentiated although accounting for a low amount of genetic variance (Table 3). Genetic differentiation was also more important between stages than among geographic locations in 2 out of the 3 years studied (Table 4). In both cases, almost all genetic variance was among individuals (Tables 3 and 4). Based on these results, it is not surprising that STRUCTURE was not able to find genetic structure in any set of *E. mediohispanicum* individuals tested in this study [see Supporting Information—Appendix S1, Fig. S2].

In contrast, genetic differences between demes estimated as distances between centroids obtained from DAPC did yield interpretable results when compared to field data. As hypothesized, genetic distances between demes within a population and year depended on the observed temporal migration rate for that particular population. However, the relationship between genetic differences between demes and temporal migration rates was not entirely as expected. We found a turning point in the relationship between the variables: decreasing genetic distances with low-moderate temporal migration rates (up to temporal migration rates of 40–50 %) and increasing genetic distances with high temporal migration rates above 50 % (Fig. 3A).

#### Model simulations

The strongest effect of temporal migration on the simulated populations was observed for large migration rates and intermediate migration event probabilities. These scenarios maximized the probability of deme extinction and recolonization, minimized stochastic growth rate values (Fig. 4), and affected genetic diversity by reducing deme mean values and increasing standard deviation values (Fig. 5). These patterns were probably due to the higher stochastic fluctuations observed at high migration rate which affects, in turns, to extinction/recolonization dynamics and diversity [see Supporting Information—Fig. S1, Appendix S1]. In contrast, the mean number of individuals per deme reached the minimum values for higher migration event probabilities (Fig. 4).

Table 1. Demographic field data for each transition and sampling site including the total number of tagged *E. mediohispanicum* individuals. We show re-sampled individuals the next year (re-sampled), the number of tagged vegetative rosettes that died during the study year (dead), the number of tagged vegetative rosettes that remained as vegetative (vegetative), the number of tagged vegetative rosettes that became reproductive (reproductive), the estimated mortality rate and the estimated temporal migration rate (the ratio between vegetative and re-sampled individuals).

Transition	Site code	Re-sampled	Dead	Vegetative	Reproductive	Mortality (%)	Temporal migrants (%)
2010–11	Em01	103	42	33	28	40.78	54.10
	Em08	37	16	9	12	43.24	42.86
	Em23	104	23	12	69	22.12	14.81
	Em25	95	37	11	47	38.95	18.97
2011–12	Em01	33	20	5	8	60.61	38.46
	Em08	56	35	14	7	62.50	66.67
	Em23	80	31	10	39	38.75	20.41
	Em25	1	-	-	1	_	-

Table 2. Genetic diversity data including vegetative and reproductive groups of E. mediohispanicum individuals (total N = 687) for each year and sampling site. Each developmental state for a given year and sampling site represents a different deme. The genetic parameters include the mean number of alleles per marker ( $n_a$ ), observed heterozygosity ( $H_o$ ) and expected mean heterozygosity ( $H_g$ ). Standard deviations (±SDs) for each parameter are also given.

Parameter		2010		2011		2012		
	Site code	Vegetative	Reproductive	Vegetative	Reproductive	Vegetative	Reproductive	
n <sub>a</sub>	Em01	8.67 ± 4.87	10.56 ± 5.60	9.33 ± 4.78	9.89 ± 5.53	8.56 ± 4.07	9.33 ± 4.33	
a	Em08	9.22 ± 3.77	$9.89 \pm 4.54$	9.67 ± 5.07	$9.22 \pm 4.52$	$7.89 \pm 3.59$	$9.00 \pm 3.91$	
	Em23	$9.56 \pm 4.13$	9.56 ± 4.59	$9.44 \pm 4.16$	$9.56 \pm 4.33$	$10.00 \pm 4.87$	$10.44 \pm 5.27$	
	Em25	9.11 ± 4.08	9.11 ± 4.29	9.67 ± 4.30	9.22 ± 4.63	-	$9.89 \pm 3.95$	
$H_{o}$	Em01	0.51 ± 0.18	0.61 ± 0.18	0.62 ± 0.15	0.65 ± 0.14	0.55 ± 0.15	0.59 ± 0.13	
	Em08	$0.57 \pm 0.24$	0.57 ± 0.24	$0.57 \pm 0.11$	$0.57 \pm 0.18$	$0.55 \pm 0.17$	$0.52 \pm 0.15$	
	Em23	$0.57 \pm 0.23$	$0.54 \pm 0.21$	$0.59 \pm 0.17$	$0.59 \pm 0.20$	$0.53 \pm 0.16$	$0.48 \pm 0.14$	
	Em25	$0.61 \pm 0.08$	0.57 ± 0.07	$0.59 \pm 0.06$	$0.60 \pm 0.06$	-	$0.58 \pm 0.13$	
$H_s$	Em01	0.75 ± 0.14	0.78 ± 0.18	0.73 ± 0.14	0.76 ± 0.13	0.72 ± 0.15	0.75 ± 0.16	
	Em08	$0.76 \pm 0.14$	$0.74 \pm 0.13$	$0.74 \pm 0.18$	$0.72 \pm 0.18$	$0.73 \pm 0.13$	$0.73 \pm 0.16$	
	Em23	$0.76 \pm 0.12$	$0.76 \pm 0.14$	$0.74 \pm 0.16$	$0.75 \pm 0.14$	$0.76 \pm 0.12$	$0.78 \pm 0.13$	
	Em25	$0.76 \pm 0.13$	$0.77 \pm 0.12$	$0.72 \pm 0.15$	$0.74 \pm 0.15$	_	$0.77 \pm 0.11$	

Table 3. Hierarchical analysis of molecular variance components and hierarchical F-statistics over all loci, including four E. mediohispanicum sites ('Total') sampled during 3 years ('Year'), each showing two groups of plants that differ in developmental stage (vegetative and reproductive, 'Stage') representing two demes. For each level (including individuals within stage) we show the estimated F-statistic and the confidence interval (CI) after 1000 bootstrap replicates. F-statistics are significant if CI does not include zero (represented in bold).

Levels	F-statistic	CI (2.5–97.5 %)
Population/Total	0.0008	-0.0018-0.0031
Year/Total	0.0086	0.0047-0.0125
Stage/Total	0.0114	0.0077-0.0148
Individual/Total	0.2131	0.1178-0.3366

Temporal genetic differentiation between demes estimated using  $F_{\rm ST}$  values was low (around 0.05) but significant in all the scenarios with no migration and the proportion of significant P-values decreased as migration rate and migration event probability increased (Fig. 6). The largest  $F_{\rm ST}$  value (0.10) in our simulation was obtained for higher migration rates and low migration event probabilities.

Interestingly, model simulations mimicked the results obtained with empirical data. In particular, genetic distances between simulated demes also showed a U-shaped relationship between genetic differences between demes and temporal migration rates, with decreasing genetic distances with increasing temporal migration rates, and increasing genetic distances with higher temporal migration rate (Fig. 3B).

## **Discussion**

Theoretically, populations of biennial organisms should be composed of individuals with homogenous developmental stages, either vegetative or reproductive at a given year. However, vegetative and reproductive individuals co-occur in natural populations of biennials (Gross 1981; Kelly 1985; De Jong and Klinkhamer 1988; Suzuki et al. 2003; Petrů 2005;

Valverde et al. 2016). This means that, whatever the reasons, temporal asynchrony of at least 1 year in the completion of the life cycle must have occurred between groups of individuals. Annual asynchrony in reproduction among individuals within populations, defining demes in our biennial plant E. mediohispanicum, leads to temporal genetic structure of populations. Examples of the consequences of such temporal genetic structure in biennial plant populations are scarce in the literature (Wells and Wells 1980; Valverde et al. 2016). Despite interesting questions are still open (such as whether the propensity to flower in year 2 or year 3 is heritable and whether flowering asynchrony of individuals is a heritable trait and whether temporal migrant and non-migrant individuals differ in size and timing of flowering), our study provided double evidence, empirical and simulated, for the existence of temporal structure in populations in the biennial E. mediohispanicum.

We expected genetic differences between demes within E. mediohispanicum populations to diminish with increasing temporal migration rates between them, because of the genetic homogenization caused by migration (Slatkin 1985, 1987; Charlesworth et al. 2003; Holderegger et al. 2010). The combination of empirical estimates of temporal migration rates from natural populations with estimates of genetic structure obtained with DAPC showed that our hypothesis was partially correct, but with an important particularity. As expected, the relationship between temporal migration rates and genetic differences between demes within populations was negative as long as temporal migration rates were below 50 %. However, when temporal migration rates were above 50 %, genetic differences between demes went up again, eventually depicting a U-shape relationship between temporal migration rates and genetic differences between demes within populations (Fig. 3A). Interestingly, the simulations performed also yielded the same U-shape relationship between the same two parameters (Fig. 3B), providing additional support to our finding and revealing how a simple model is able to capture the genetic effects of (temporal) migration on populations. As far as we know, this is the first work reporting this U-shape dependency between genetic differentiation and temporal migration. Although the negative relationship of the U-shape function between temporal

Table 4. Hierarchical analysis of molecular variance including four *E. mediohispanicum* sites. Each site is composed of two groups of plants that differ in developmental stage (vegetative and reproductive) representing two demes. Degrees of freedom (df), percentage of variation of each hierarchical level and their respective fixation indexes with their P-values are given. (\*) Population Em25 was excluded for 2012 estimation due to the lack of vegetative individuals.

	2010				2011				2012 (*)			
Site code	df	Variation (%)	Fixation index	P-value	df	Variation (%)	Fixation index	P-value	df	Variation (%)	Fixation index	P-value
Among sites	3	0.21	0.0021	0.1896	3	0.42	0.0042	0.06256	2	0.96	0.00959	<0.0001
Between stages (sites)	4	0.89	0.0089	0.042	4	0.77	0.0078	0.02727	3	0.31	0.00312	0.72434
Within stages Total	440 447	98.89	0.011	<0.0001	450 457	98.8	0.012	<0.0001	348 353	98.73	0.01269	0.00391

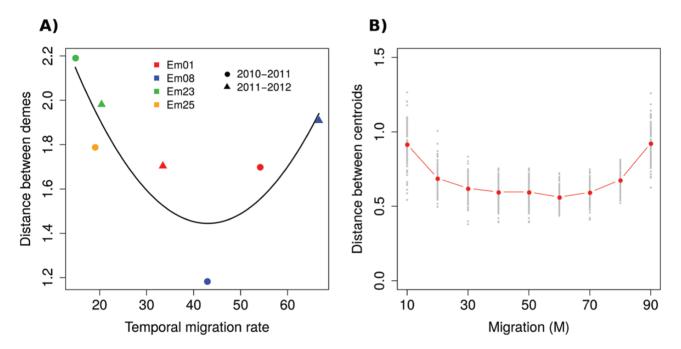


Figure 3. (A) Genetic distance between demes within a population and year estimated from natural E. mediohispanicum populations as a function of the observed temporal migration rates. (B) The same relationship but estimated using data simulated by our model.

migration rates and genetic differences among demes within populations is easy to understand, the positive relationship between the same traits emerging above 50 % of temporal migration rates deserves further attention (but see Supporting Information—Appendix S2).

Temporal migration might represent an advantage for a biennial because populations are made of temporally distributed subpopulations co-occurring in the same location. Two viable subpopulations exchanging genes over time should positively contribute to at least the maintenance of genetic variation of the entire system (Gillespie 1975; Star and Spencer 2013). Our results supported this expectation in two different ways. First, field data showed that reproductive and vegetative groups of individuals from different populations exhibited high gene diversity values with little differences between them (Table 2). Second, simulations clearly indicated that mean gene diversity between demes fluctuated little around a high value between low and medium temporal migration rates (Fig. 3A). In contrast, temporal migration could have some adverse consequences for

gene diversity at high migration rates because intense migration in biennials represents an important loss of individuals in the donor population. In fact, migrants are established vegetative rosettes, which are the survivors of massive mortality events commonly recorded among seedlings and juveniles of annual and short-lived plants (Quintana-Ascencio et al. 2003; Menges and Quintana-Ascencio 2004; Picó and Retana 2008; Montesinos et al. 2009; Picó 2012). Thus, sharp declines in population size due to temporal migrants eventually have detrimental effects on population performance and viability. Our simulations depicted this pattern showing lower growth rates and higher extinction and recolonization dynamics (Fig. 4), which in turn led to decreasing gene diversity at high migration rate scenarios (Fig. 5A). We believe that such dramatic changes in deme population size due to high temporal migration rates, which resemble the effects of demographic stochasticity on population fate (Menges 1991; Fischer and Matthies 1998; Quintana-Ascencio et al. 1998; Matthies et al. 2004; Menges and Quintana-Ascencio 2004), accounted for the change of trend between

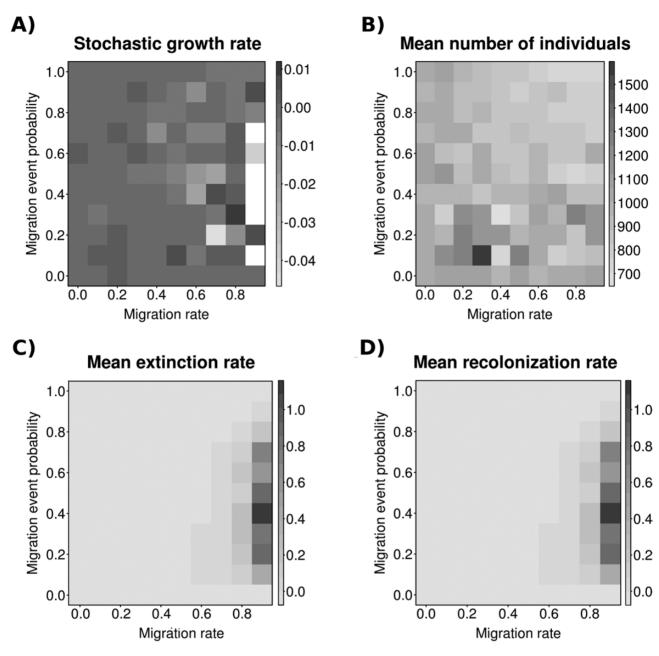


Figure 4. Demographic behaviour of the model across all migration rate and migration event probability scenarios: (A) stochastic deme growth rate, (B) mean number of individuals per deme, (C) mean extinction rates per deme and (D) mean recolonization rates per deme. For the sake of clarity, the results are based on data from one of the two demes because the complementary deme exhibited the same pattern (results not shown). Migration rates vary from 0 to 0.9 and migration event probabilities from 0 to 1.0 with 0.1 intervals. Results were obtained averaging 100 simulations per scenario.

temporal migration rates and genetic differences among demes within populations at high migration rates.

The four *E. mediohispanicum* populations were not genetically structured. In addition, the analysis of the distribution of genetic diversity within and among populations, albeit significant, indicated that the genetic differentiation among the four *E. mediohispanicum* populations was very low. We believe that the highly diverse pollinator community visiting *E. mediohispanicum* plants (Gómez et al. 2007), which represents high effective gene flow rates among populations, accounts for the lack of genetic structure across the entire system. Furthermore, it is well known that reproductive *E. mediohispanicum* in rainy years becomes ubiquitous across the whole area including the four study sites

(authors' personal observation). Hence, gene flow and genetic homogenization among sites can rapidly increase in that particular area, erasing population structure at the population level. Interestingly, our analyses only detected interpretable signals of genetic structure between demes within populations when using DAPC rather than the Bayesian clustering method STRUCTURE, probably because the former has the power to unravel complex population subdivisions (Jombart et al. 2010) due to the mathematical properties of the Discriminant Analysis (which maximizes the differences between groups while minimizes variation within groups).

In summary, our empirical and modeling approaches showed consistent patterns between genetic differences

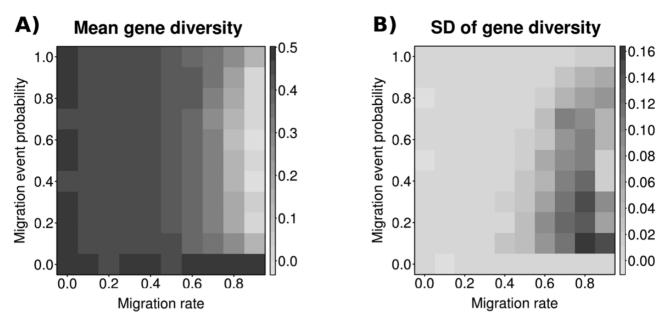


Figure 5. Genetic behaviour of the model across all migration rate and migration event probability scenarios: (A) mean gene diversity between demes ( $H_g$ ), and (B) standard deviation of mean  $H_g$ . Migration rates vary from 0 to 0.9 and migration event probabilities from 0 to 1.0 with 0.1 intervals. Results were obtained averaging 100 simulations per scenario.

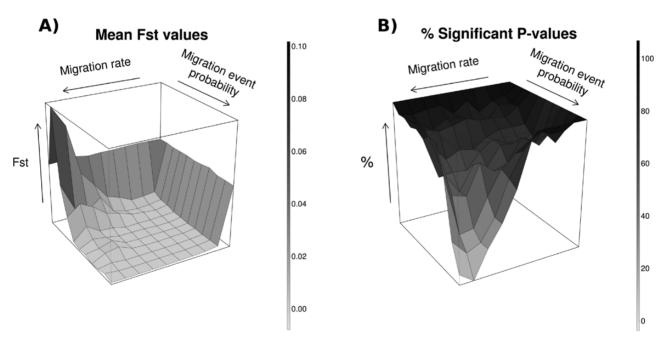


Figure 6. Genetic structure between demes across all migration rate and migration event probability scenarios given by (A)  $F_{ST}$  values and (B) their respective percentage of significant P-values. Migration rates vary from 0 to 0.9 and migration event probabilities from 0 to 1.0 with 0.1 intervals. Results were obtained averaging 100 simulations per scenario.

between demes within populations and temporal migration rates in *E. mediohispanicum*. We believe that temporal genetic structure could emerge with more intensity in systems with lower effective gene flow. Our study shed light on the complexity of population genetic features of some biological systems, such as biennial plants, in which the extent of temporal migration rates between demes within populations determines temporal genetic structure through migration or demographic stochasticity.

# **Supporting Information**

The following additional information is available in the online version of this article—

Table S1. Schematic representation of the model procedure. Appendix S1. Model simulations.

Appendix S2. Interpretation of the U-shape relationship.

Figure S1. Representative simulations pinpointing

population dynamics for central (0.5) and extreme values of

migration rate (0.0 and 0.9) and migration event probabilities

Figure S2. Summary of STRUCTURE analysis results.

# **Data Availability**

Data available in the Supplementary Material.

# Sources of Funding

Funding was provided by projects CGL2009-07487/BOS and CGL2016-77720-P (AEI/FEDER, UE) to F.X.P., by the Impact Fellow programme from the University of Stirling to M.A. and by the Portuguese Foundation for Science and Technology (SFRH/ BPD/111015/2015) to A.J.M.-P.

# Contributions by the Authors

A.J.M.-P., M.A. and F.X.P. planned and designed the research. A.J.M.-P., M.A. and F.X.P. conducted fieldwork. A.J.M.-P. developed the model. F.X.P. coordinated molecular work. A.J.M.-P., M.A. and F.X.P. analysed the data and wrote the manuscript.

### **Conflict of Interest**

None declared.

# Acknowledgements

We are grateful to Armando Caballero, Juan Pedro Martínez Camacho, Mario Vallejo-Marin, Mohammed Bakkali, Robin S. Waples, Xavier Thibert-Plante and Antonio Castilla for their comments and discussions on a previous draft of the manuscript. Esperanza Manzano, Leticia Ayllón and Rocío Gómez assisted in the laboratory. The EVOFLOR discussion group stimulated the development of this study. We thank the Sierra Nevada National Park headquarters for the permits and the support during our samplings in the field. We thank Bioportal at the University of Oslo and Residencia de Estudiantes de la Universidad de Zaragoza in Jaca for logistic support. We also thank the staff of the laboratory of molecular ecology (LEM) of the EBD-CSIC for assistance.

#### **Literature Cited**

- Aspinwall N. 1974. Genetic analysis of North American populations of the pink salmon, Oncorhynchus gorbuscha, possible evidence for the neutral mutation-random drift hypothesis. Evolution 28:295-305.
- Bustos-Segura C, Fornoni J, Núñez-Farfán J. 2014. Evolutionary changes in plant tolerance against herbivory through a resurrection experiment. Journal of Evolutionary Biology 27:488-496.
- Charlesworth B, Charlesworth D, Barton NH. 2003. The effects of genetic and geographic structure on neutral variation. Annual Review of Ecology, Evolution and Systematics 34:99-125.
- Chung MY, Nason JD, Epperson BK, Chung MG. 2003. Temporal aspects of the fine-scale genetic structure in a population of Cinnamomum insularimontanum (Lauraceae). Heredity 90:98-106.
- De Jong TJ, Klinkhamer PGL. 1988. Population ecology of the biennials Cirsium vulgare and Cynoglossum officinale in a coastal and sand-dune area. Journal of Ecology 76:366-382.
- Denney DA, Jameel MI, Bemmels JB, Rochford ME, Anderson JT. 2020. Small spaces, big impacts: contributions of micro-environmental variation to population persistence under climate change. AoB PLANTS 12:plaa005; doi:10.1093/aobpla/plaa005.

- Dionne M, Caron F, Dodson JJ, Bernatchez L. 2008. Landscape genetics and hierarchical genetic structure in Atlantic salmon: the interaction of gene flow and local adaptation. Molecular Ecology 17:2382-2396.
- Dolan RW, Quintana-Ascencio PF, Menges ES. 2008. Genetic change following fire in populations of a seed-banking perennial plant. Oecologia 158:355-360.
- Evanno G, Regnaut S, Goudet J. 2005. Detecting the number of clusters of individuals using the software STRUCTURE: a simulation study. Molecular Ecology 14:2611-2620.
- Excoffier L, Lischer HE. 2010. Arlequin suite ver 3.5: a new series of programs to perform population genetics analyses under Linux and Windows. Molecular Ecology Resources 10:564–567.
- Falahati-Anbaran M, Lundemo S, Ågren J, Stenøien HK. 2011. Genetic consequences of seed banks in the perennial herb Arabidopsis lyrata subsp. petraea (Brassicaceae). American Journal of Botany 98:1475-1485.
- Falahati-Anbaran M, Lundemo S, Stenøien HK. 2014. Seed dispersal in time can counteract the effect of gene flow between natural populations of Arabidopsis thaliana. The New Phytologist 202:1043-1054.
- Falush D, Stephens M, Pritchard JK. 2003. Inference of population structure using multilocus genotype data: linked loci and correlated allele frequencies. Genetics 164:1567-1587.
- Fischer M, Matthies D. 1998. The effect of population size on performance in the rare plant Gentianella germanica. Journal of Ecology 86:195-204.
- Fox GA. 2003. Assortative mating and plant phenology: evolutionary and practical consequences. Evolutionary Ecology Research 5:1-18.
- Frachon L, Libourel C, Villoutreix R, Carrère S, Glorieux C, Huard-Chauveau C, Navascués M, Gay L, Vitalis R, Baron E, Amsellem L, Bouchez O, Vidal M, Le Corre V, Roby D, Bergelson J, Roux F. 2017. Intermediate degrees of synergistic pleiotropy drive adaptive evolution in ecological time. Nature Ecology & Evolution 1:1551-1561.
- Franks SJ, Kane NC, O'Hara NB, Tittes S, Rest JS. 2016. Rapid genome-wide evolution in Brassica rapa populations following drought revealed by sequencing of ancestral and descendant gene pools. Molecular Ecology **25**:3622-3631.
- Franks SJ, Sim S, Weis AE. 2007. Rapid evolution of flowering time by an annual plant in response to a climate fluctuation. Proceedings of the National Academy of Sciences of the United States of America 104:1278-1282.
- Franks SJ, Weis AE. 2008. A change in climate causes rapid evolution of multiple life-history traits and their interactions in an annual plant. Journal of Evolutionary Biology 21:1321-1334.
- Gillespie JH. 1975. The role of migration in the genetic structure of populations in temporarily and spatially varying environments I. Conditions for polymorphism. The American Naturalist 109:127–136.
- Gomaa NH, Montesinos-Navarro A, Alonso-Blanco C, Picó FX. 2011. Temporal variation in genetic diversity and effective population size of Mediterranean and subalpine Arabidopsis thaliana populations. Molecular Ecology 20:3540-3554.
- Gómez JM. 2005. Long-term effects of ungulates on performance, abundance, and spatial distribution of two montane herbs. Ecological Monographs **75**:231–258.
- Gómez JM. 2007. Dispersal-mediated selection on plant height in an autochorously dispersed herb. Plant Systematics and Evolution 268:119-130.
- Gómez JM, Bosch J, Perfectti F, Fernández J, Abdelaziz M. 2007. Pollinator diversity affects plant reproduction and recruitment: the tradeoffs of generalization. Oecologia 153:597-605.
- Gómez R, Méndez-Vigo B, Marcer A, Alonso-Blanco C, Picó FX. 2018. Quantifying temporal change in plant population attributes: insights from a resurrection approach. AoB PLANTS 10:ply063; doi:10.1093/ aobpla/plv063.
- Goudet J. 1995. Fstat version 1.2: a computer program to calculate F-statistics. Journal of Heredity 86:485-486.
- Goudet J. 2013. hierfstat: estimation and tests of hierarchical F-statistics. R package version 0.04-10. http://CRAN.R-project.org/package=hierfstat
- Gross KL. 1981. Predictions of fate from rosette size in four "biennial" plant species: Verbascum thapsus, Oenothera biennis, Daucus carota, and Tragopogon dubius. Oecologia 48:209-213.
- Heath DD, Busch C, Kelly J, Atagi DY. 2002. Temporal change in genetic structure and effective population size in steelhead trout (Oncorhynchus mykiss). Molecular Ecology 11:197-214.

- Hendry AP, Berg OK, Quinn TP. 1999. Condition dependence and adaptation-by-time: breeding date, life history, and energy allocation within a population of salmon. Oikos 85:499-514.
- Hendry AP, Day T. 2005. Population structure attributable to reproductive time: isolation by time and adaptation by time. Molecular Ecology 14:901-916
- Hey J. 2011. Regarding the confusion between the population concept and Mayr's "population thinking". The Quarterly Review of Biology 86:253-264.
- Hoffman JI, Amos W. 2005. Microsatellite genotyping errors: detection approaches, common sources and consequences for paternal exclusion. Molecular Ecology 14:599-612.
- Holderegger R, Buehler D, Gugerli F, Manel S. 2010. Landscape genetics of plants. Trends in Plant Science 15:675-683.
- Horgan-Kobelski T, Matesanz S, Sultan SE. 2016. Limits to future adaptation in the invasive plant Polygonum cespitosum: expression of functional and fitness traits at elevated CO2. The Journal of Heredity 107:42-50.
- Hossaert-McKey M, Valero M, Magda D, Jarry M, Cuguen J, Vernet P. 1996. The evolving genetic history of a population of Lathyrus sylvestris: evidence from temporal and spatial genetic structure. Evolution **50**·1808–1821
- Jombart T. 2008. adegenet: a R package for the multivariate analysis of genetic markers, Bioinformatics 24:1403-1405.
- Jombart T, Devillard S, Balloux F. 2010. Discriminant analysis of principal components: a new method for the analysis of genetically structured populations. BMC Genetics 11:94.
- Jonckers LH. 1973. The concept of population in biology. A logical analysis with a suggestion to diminish the confusion of thought. Acta Biotheoretica 22:78-108.
- Kelly D. 1985. On strict and facultative biennials. Oecologia 67:292-294.
- Kent RJ, Mharakurwa S, Norris DE. 2007. Spatial and temporal genetic structure of Anopheles arabiensis in Southern Zambia over consecutive wet and drought years. The American Journal of Tropical Medicine and Hygiene 77:316-323.
- Kimura M, Weiss GH. 1964. The stepping stone model of population structure and the decrease of genetic correlation with distance. Genetics 49:561-576.
- Königer J, Rebernig CA, Brabec J, Kiehl K, Greimler J. 2012. Spatial and temporal determinants of genetic structure in Gentianella bohemica. Ecology and Evolution 2:636-648.
- Kuester A, Wilson A, Chang SM, Baucom RS. 2016. A resurrection experiment finds evidence of both reduced genetic diversity and potential adaptive evolution in the agricultural weed Ipomoea purpurea. Molecular Ecology 25:4508-4520.
- Loveless MD, Hamrick JL. 1984. Ecological determinants of genetic structure of plant populations. Annual Review of Ecology, Evolution and Systematics 15:65-95.
- Lundemo S, Falahati-Anbaran M, Stenøien HK. 2009. Seed banks cause elevated generation times and effective population sizes of Arabidopsis thaliana in northern Europe. Molecular Ecology 18:2798-2811.
- Manel S, Schwartz MK, Luikart G, Taberlet P. 2003. Landscape genetics: combining landscape ecology and population genetics. Trends in Ecology and Evolution 18:189-197
- Martínez-Cruz B, Godoy JA, Negro JJ. 2007. Population fragmentation leads to spatial and temporal genetic structure in the endangered Spanish imperial eagle. Molecular Ecology 16:477-486.
- Matthies D, Bräuer I, Maibom W, Tscharntke T. 2004. Population size and the risk of local extinction: empirical evidence from rare plants. Oikos 105:481-488
- Menges ES. 1991. Seed germination percentage increases with population size in a fragmented prairie species. Conservation Biology 5:158–164.
- Menges ES, Quintana-Ascencio PF. 2004. Population viability with fire in Eryngium cuneifolium: deciphering a decade of demographic data. Ecological Monographs 74:79-99.
- Montesinos A, Tonsor SJ, Alonso-Blanco C, Picó FX. 2009. Demographic and genetic patterns of variation among populations of Arabidopsis thaliana from contrasting native environments. PLoS One 4:e7213.
- Muñoz-Pajares AJ, Herrador MB, Abdelaziz M, Picó FX, Sharbel TF, Gómez IM. Perfectti F. 2011. Characterization of microsatellite loci in Erysimum mediohispanicum (Brassicaceae) and cross-amplification in related species. American Journal of Botany 98:e287-e289

- Muñoz-Pajares AJ, Perfectti F, Loureiro J, Abdelaziz M, Biella P, Castro M, Castro S, Gómez JM. 2018. Niche differences may explain the geographic distribution of cytotypes in Erysimum mediohispanicum. Plant Biology 20:139-147.
- Nevo E. Fu YB. Pavlicek T. Khalifa S. Tavasi M. Beiles A. 2012. Evolution of wild cereals during 28 years of global warming in Israel. Proceedings of the National Academy of Sciences of the United States of America 109:3412-3415.
- Nunney L. 2002. The effective size of annual plant populations: the interaction of a seed bank with fluctuating population size in maintaining genetic variation. The American Naturalist 160:195-204.
- Peakall R, Smouse PE. 2006. GENALEX 6: genetic analysis in Excel. Population genetic software for teaching and research. Molecular Ecology Notes 6:288-295.
- Petrů M. 2005. Year-to-year oscillations in demography of the strictly biennial Pedicularis sylvatica and effects of experimental disturbances. Plant Ecology 181:289-298.
- Picó FX. 2012. Demographic fate of Arabidopsis thaliana cohorts of autumnand spring-germinated plants along an altitudinal gradient. Journal of Ecology 100:1009-1018.
- Picó FX, Retana J. 2008. Age-specific, density-dependent and environmentbased mortality of a short-lived perennial herb. Plant Biology 10:374-381.
- Pompanon F, Bonin A, Bellemain E, Taberlet P. 2005. Genotyping errors: causes, consequences and solutions. Nature Reviews Genetics 6:847-859.
- Pritchard JK, Stephens M, Donnelly P. 2000. Inference of population structure using multilocus genotype data. Genetics 155:945-959.
- Quintana-Ascencio PF, Dolan RW, Menges ES. 1998. Hypericum cumulicola demography in unoccupied and occupied Florida scrub patches with different time-since-fire, Journal of Ecology 86:640-651.
- Quintana-Ascencio PF, Menges ES, Weekly CW. 2003. A fire-explicit population viability analysis of Hypericum cumulicola in Florida rosemary scrub. Conservation Biology 17:433-449.
- R Development Core Team. 2011. R: a language and environment for statistical computing, Vienna, Austria: R Foundation for Statistical Computing,
- Riccioni G, Landi M, Ferrara G, Milano I, Cariani A, Zane L, Sella M, Barbujani G, Tinti F. 2010. Spatio-temporal population structuring and genetic diversity retention in depleted Atlantic Blue fin tuna of the Mediterranean Sea. Proceedings of the Natural Academy of Sciences of the United States of America 107:2102-2107.
- Rosenberg NA, Pritchard JK, Weber JL, Cann HM, Kidd KK, Zhivotovsky LA, Feldman MW. 2002. Genetic structure of human populations. Science **298**:2381-2385
- Slatkin M. 1985. Gene flow in natural populations. Annual Review of Ecology, Evolution and Systematics 16:393-430.
- Slatkin M. 1987. Gene flow and the geographic structure of natural populations. Science 236:787-792.
- Sloop CM, Ayres DR, Strong DR. 2011. Spatial and temporal genetic structure in a hybrid cordgrass invasion. Heredity 106:547-556.
- Star B, Spencer HG. 2013. Effects of genetic drift and gene flow on the selective maintenance of genetic variation. Genetics 194:235-244.
- Storfer A, Murphy MA, Evans JS, Goldberg CS, Robinson S, Spear SF, Dezzani R, Delmelle E, Vierling L, Waits LP. 2007. Putting the "landscape" in landscape genetics. Heredity 98:128-142.
- Sultan SE, Horgan-Kobelski T, Nichols LM, Riggs CE, Waples RK. 2013. A resurrection study reveals rapid adaptive evolution within populations of an invasive plant. Evolutionary Applications 6:266-278.
- Suzuki RO, Kudoh H, Kachi N. 2003. Spatial and temporal variations in mortality of the biennial plant, Lysimachia rubida: effects of intraspecific competition and environmental heterogeneity. Journal of Ecology 91:114-125
- Thomann M, Imbert E, Engstrand RC, Cheptou PO. 2015. Contemporary evolution of plant reproductive strategies under global change is revealed by stored seeds. Journal of Evolutionary Biology 28:766-778.
- Valverde J, Gómez JM, García C, Sharbel TF, Jiménez MN, Perfectti F. 2016. Inter-annual maintenance of the fine-scale genetic structure in a biennial plant. Scientific Reports 6:37712.
- Vitalis R, Glémin S, Olivieri I. 2004. When genes go to sleep: the population genetic consequences of seed dormancy and monocarpic perenniality. The American Naturalist 163:295-311.
- Waples RS. 1990a. Conservation genetics of Pacific salmon. II. Effective population size and the rate of loss of genetic variability. Journal of Heredity 81:267-276.

- Waples RS. 1990b. Conservation genetics of Pacific salmon. III. Estimating effective population size. Journal of Heredity 81:277-289.
- Waples RS. 2002. Effective size of fluctuating salmon populations. Genetics **161**:783-791.
- Waples RS. 2006. Seed banks, salmon, and sleeping genes: effective population size in semelparous, age-structured species with fluctuating abundance. The American Naturalist 167:118-135.
- Waples RS, Gaggiotti O. 2006. What is a population? An empirical evaluation of some genetic methods for identifying the number of gene pools and their degree of connectivity. Molecular Ecology
- Weir BS, Cockerham CC. 1984. Estimating F-statistics for the analysis of population structure. Evolution 38:1358-1370.
- Weis AE, Kossler TM. 2004. Genetic variation in flowering time induces phenological assortative mating: quantitative genetic methods applied to Brassica rapa. American Journal of Botany 91:825-836.
- Wells H, Wells PH. 1980. Are geographic populations equivalent to genetic populations in biennial species? A study using Verbascum virgatum (Scrophulariaceae). Genetic Research Cambridge 36:17-28.
- Welt RS, Litt A, Franks SJ. 2015. Analysis of population genetic structure and gene flow in an annual plant before and after a rapid evolutionary response to drought. AoB PLANTS 7:plv026; doi:10.1093/aobpla/plv026.
- Wright S. 1951. The genetical structure of populations. Annals of Eugenics
- Wright S. 1965. The interpretation of population structure by F-statistics with special regard to systems of mating. Evolution 19:395-420.