

## **Arsenopyrite dissolution rates in O<sub>2</sub>-bearing solutions.**

Maria P. Asta<sup>1\*</sup>, Jordi Cama<sup>1</sup>, Carlos Ayora<sup>1</sup>, Patricia Acero<sup>2</sup> and Giovanni de Giudici<sup>3</sup>

<sup>1</sup> Institute of Environmental Assessment and Water Research (IDAEA), CSIC  
Jordi Girona, 18-26, Barcelona 08034, Spain

<sup>2</sup> Department of Earth Sciences, University of Zaragoza  
Pedro Cerbuna 12, Zaragoza 50009, Spain

<sup>3</sup> Department of Earth Science, University of Cagliari  
Via Trentino 51, Cagliari 09127, Italy

Phone +34 976761000 ext. 3159

Fax: +34 976 761106

E-mail: [mpasta@unizar.es](mailto:mpasta@unizar.es)

To be submitted to: *Chemical Geology*

\*Corresponding author; Present address; Department of Earth Sciences, University of Zaragoza, Pedro Cerbuna 12, Zaragoza 50009, Spain

## ABSTRACT

Arsenopyrite dissolution was studied by means of long-term, stirred and non-stirred flow-through experiments in the pH range of 1 to 9 at 25, 50 and 70 °C and at different input dissolved-O<sub>2</sub> concentrations (from 0.2 to 8.7 mg L<sup>-1</sup>).

At pH lower than 4, aqueous iron, which is mainly in the ferrous form, and arsenic are stoichiometrically released. Sulphur concentrations released were lower than stoichiometrically expected (S/As <1). X-ray Photoelectron Spectroscopy (XPS) and MicroRaman Spectroscopy surface analyses on reacted and unreacted samples showed an enrichment of the reacted arsenopyrite surface in sulphur and arsenic under acidic conditions.

In the light of these results, the steady-state dissolution rates were estimated by the release of arsenic at pH < 6 and were used to derive an empirical dissolution rate law expressed as:

$$R_{\text{arsenopyrite}} (\text{mol m}^{-2} \text{ s}^{-1})_{25^\circ \text{C}} = 10^{-7.41 \pm 0.47} \cdot a_{\text{O}_2}^{0.76 \pm 0.11} \cdot a_{\text{H}^+}^{-0.12 \pm 0.07}$$

where  $a_{\text{O}_2}$  and  $a_{\text{H}^+}$  are the activities of hydrogen ions and dissolved oxygen, respectively and their exponents were estimated from multiple linear regression of the dissolution rates. Temperature increase from 25 to 70 °C yields an apparent activation energy for the arsenopyrite oxidation by dissolved oxygen of 18.5±1.6 kJ mol<sup>-1</sup>.

At pH > 6, aqueous iron is mainly in the ferric form and is depleted as it precipitates as Fe-oxyhydroxide onto arsenopyrite surfaces, yielding Fe/As and Fe/S less than one; between pH 7 and 9, iron depletion is complete, and sulphur released is more abundant than arsenic released, which is precipitated as As-O phases, as confirmed by MicroRaman spectroscopy. At pH 6-9, iron-oxyhydroxide phases and arsenic oxide phases upon the arsenopyrite surface provide an effective layer that reduces diffusion of dissolved oxygen and arsenopyrite

25 dissolution. As coating on the arsenopyrite surface becomes the rate-limiting step, the  
26 Shrinking Core Model (SCM) allows quantification of the surface dissolution rate,  
27 especially from data obtained whereas the effect of coating was still negligible. The SCM  
28 also allowed us to calculate the effective coefficient for oxygen diffusion through the  
29 coating, which can vary from  $10^{-17}$  to  $1.5 \cdot 10^{-16} \text{ m}^2 \text{ s}^{-1}$ . The formation of such a coating  
30 produced a decrease in arsenic and sulphur release over time and a final surface passivation.

31 **Keywords:** arsenic, arsenopyrite, dissolution, kinetics, iron coating.

32

## 1 INTRODUCTION

33 Inorganic aqueous As release is an environmental and human health concern worldwide  
34 (Ferguson, 1990; Aposhian et al., 2004; Rosman et al., 2004; Bunnell et al., 2007).

35 The arsenic content in water (surficial or groundwater) has been attributed to two main  
36 processes: desorption of arsenic previously sorbed onto the structures of Fe or Mn oxy-  
37 hydroxides, and As release by oxidation of arsenopyrite and/or As-rich pyrites in mineralized  
38 areas (Casiot et al., 2003, 2005; Frau and Ardaù, 2003; Lee et al., 2005; Lee and Chon, 2006;  
39 Pfeifer et al., 2007). The former process usually occurs in aquifers where redox conditions  
40 have been changed, e.g., the arsenic contamination in the Bangladesh delta (Bhattacharya et  
41 al., 1997, 2001, 2002, 2006; Nickson et al., 1998; Routh et al., 2000; McArthur et al., 2001;  
42 Dowling et al., 2002; Smedley and Kinniburgh 2002; Anawar et al., 2003), whereas the latter  
43 process is mainly encountered in areas where As-bearing sulphides oxidize in acid mine  
44 drainage (AMD) or in near neutral to basic pH waters. In this study, we focus on the As  
45 release as arsenopyrite dissolution takes place. An important example of As mobilization is  
46 the Iberian Pyrite Belt (IPB) in the SW of Spain and S of Portugal, which is one of the most  
47 important massive sulphide provinces in the world. This area contains a large number of  
48 abandoned sulphide mines, open pits, galleries, tailings and sulphide-sludge ponds that  
49 generate creeks with acidic water with high levels of arsenic (Sánchez-Rodas et al., 2005;  
50 Sarmiento et al., 2005, 2007; Acero et al., 2006; Asta et al., 2009). Dissolved oxygen  
51 promotes dissolution of arsenopyrite (AsFeS) and arsenical pyrite (Fe(As,S)<sub>2</sub>) with the  
52 consequent As release into run-off water (Williams, 2001; Lazareva et al., 2002; Smedley and  
53 Kinniburgh, 2002; Casiot et al., 2003; Frau and Ardaù, 2003; Welch and Stollenwerk, 2003;  
54 Lee et al., 2005; Pfeifer et al., 2007). Other examples of arsenic contaminated groundwaters  
55 have been attributed to oxidation of arsenopyrite and As-bearing sulphides in non acidic  
56 waters at pH ranging from 7-9 in the Madrid Tertiary detrital aquifer (central Spain)

57 (Hernández-García and Custodio, 2004) or in groundwater with near neutral pH at Ester dome  
58 (Fairbanks, Alaska), where dissolved arsenic concentration appears to be controlled by  
59 oxidation of arsenopyrite in the near-surface environment (Verplanck et al., 2007). Smedley et  
60 al. (2007) have recently reported arsenic contamination in circumneutral-pH groundwaters in  
61 Proterozoic basement rocks in Burkina Faso. There are also other scenarios where higher pH  
62 (neutral or alkaline) is common. For example, acid generation may be artificially attenuated by  
63 adding alkaline substances to the AMD producing materials (Pérez-López et al., 2007), which  
64 results in acid neutralization. Similarly, hydrometallurgical techniques such as cyanidation  
65 have been conducted at high pH producing alkaline waters in contact with mine residues  
66 (Salzsauler et al., 2005).

67 The effects of different factors on the oxidative dissolution of arsenopyrite, such as  
68 temperature, Fe(III) concentration, the presence of bacteria (*Acidithiobacillus ferrooxidans*),  
69 pH and particle size have been reported in the literature, either related to aqueous chemistry  
70 studies (Breed et al., 1997; Ruitenberg et al., 1999; McGuire et al., 2001a; Craw et al., 2003;  
71 Yu et al. 2004, 2007; Tallant and McKibben, 2005, Walker et al., 2006; McKibben et al.,  
72 2008) or to arsenopyrite surface spectroscopy research, mainly by X-ray Photoelectron  
73 Spectroscopy (XPS) (Buckley and Walker, 1988; Richardson and Vaughan, 1989; Nesbitt et  
74 al., 1995; Nesbitt and Muir, 1998; Hacquard et al., 1999; Mikhlin et al., 2006). This extensive  
75 literature is indicative of the important role of arsenopyrite oxidative dissolution in different  
76 geochemical environments, from AMD to metallurgical processes, and groundwater  
77 contamination.

78 Recently, two thorough reviews on arsenopyrite oxidative dissolution that take into  
79 account the aforementioned studies have been published (Corkhill and Vaughan, 2009;  
80 Lengke et al., 2009). These authors indicate that discrepancies in the kinetics of arsenopyrite  
81 dissolution still exist, being induced by different experimental parameters and limited datasets.  
82 It is pointed out that there is no consensus in fundamental processes that are involved in the

83 overall reaction, such as: (1) formation of S layers on arsenopyrite surface, (2) dissolution  
84 stoichiometry, (3) variability in dissolution rate, (4) dissolution rate-DO dependence, (5)  
85 anomalous rate-Arrhenius behaviour.

86 The aim of this paper is to describe the kinetics of arsenopyrite oxidative dissolution at  
87 different oxidizing conditions by assessing the effects of environmental factors, such as pH,  
88 dissolved oxygen, sulphate content, and temperature. The main innovation of this study is that  
89 oxidation kinetic data were obtained at a long-term steady state (hundreds of hours) and a  
90 wide range of pH values, dissolved oxygen concentrations and temperature, combining the  
91 analysis of the reacted solutions with an exhaustive examination of arsenopyrite surface by  
92 means of Scanning Electron Microscopy (SEM), X-ray Photoelectron Spectroscopy (XPS) and  
93 MicroRaman Spectroscopy. In addition, the effect of iron coatings formed at high pH on  
94 arsenopyrite surface and its implication on arsenic release has been contemplated using the  
95 Shrinking Core Model (Wen, 1968). Therefore, the results from this study should be useful to  
96 improve the applicability of the empirical rate laws for As-bearing sulphide dissolution.

97

## 98 2 MATERIALS AND METHODS

### 99 2.1 Characterization of arsenopyrite

100 The arsenopyrite samples used in this study were obtained from Martinet skarn  
101 mineralization (East Pyrenees range). Some arsenopyrite fragments were crushed in an agate  
102 mortar and sieved to a size fraction below 100 µm. The powdered sample was examined by X-  
103 ray diffraction (XRD) using Cu K $\alpha$  radiation over a 2 $\theta$  range from 0 to 60 degrees and using a  
104 scan speed of 0.025°/18s. The XRD patterns obtained showed that the sample consisted of  
105 arsenopyrite as the main phase and a minor amount of quartz (approx. 5%). BSE image of the  
106 raw sample by Cameca SX-50 electron microprobe showed that arsenopyrite (AsFeS) was the

107 dominant phase with minor amounts of quartz (SiO<sub>2</sub>) (<3%), pyrite (FeS<sub>2</sub>) and native Bi (<  
108 1%). The electron microprobe analysis (EMP) of the arsenopyrite atomic composition (%) was  
109 Fe 33.5±0.1, As 32.1±0.4 and S 34.4±0.4 based on 18 points, yielding an average chemical  
110 formula of Fe<sub>1.0</sub>As<sub>0.94</sub>S<sub>1.05</sub>. Geometric areas were estimated from particle size distributions,  
111 determined for powdered reacted and unreacted arsenopyrite samples by using a laser  
112 diffraction size analyser (Malver, mastersize E) after ultrasonic disaggregation in ethanol. The  
113 obtained geometric areas ranged from 0.04 to 0.1 m<sup>2</sup> g<sup>-1</sup>.

114 BET surface areas in the same types of samples were measured using 5-point N<sub>2</sub>  
115 adsorption isotherms with a Micromeritics ASAP 2000 surface area analyzer and the obtained  
116 values were used for normalizing the experiments at pH<6. The final BET-determined  
117 specific surface area in those experiments varied from 0.3 to 0.7 m<sup>2</sup> g<sup>-1</sup>. These values are  
118 generally five to ten times larger than the corresponding estimated geometric areas, which is  
119 very a common finding in kinetic studies and has also been described in earlier works (White  
120 and Peterson, 1990; White and Brantley, 2003 and references therein). Unreacted and reacted  
121 powders were examined by Scanning Electron Microscopy (SEM) using a JEOL JSM-840  
122 microscope and a field-emission scanning microscope Hitachi H-4100FE.

123 XPS surface examination of the initial and reacted powdered samples mounted on  
124 carbon conductive tabs was carried out with a Physical Electronics (PHI) 5500 spectrometer  
125 using a monochromatic X-ray source (with an Al K $\alpha$  line of 1486.6 eV energy and 350 W)  
126 placed perpendicular to the analyzer axis and calibrated using the 3d5/2 line of Ag with a  
127 width of 0.8 eV and a binding energy of 368.3 eV. All these measurements were made in an  
128 ultra high vacuum (UHV) chamber (pressure between 6.6·10<sup>-11</sup> and 6.6·10<sup>-12</sup> atm). The  
129 analyser pass energy was 23 eV. An electron flood gun at low energies (below 25 eV) was  
130 used for charge compensation. Comparison of the relative positions of the different peaks in  
131 all the studied spectra indicated that charge shifting could be considered uniform. Spectra are  
132 shown as raw data corrected by adjusting the C1s peak (corresponding to adventitious carbon,

133 to a binding energy of 284.6 eV) because of the charge of the sample. Given the lack of  
134 sample cooling while acquiring the measurements, loss of elemental sulphur could occur.  
135 Therefore, the presence or absence of elemental sulphur is discussed below. Atomic  
136 concentrations of arsenic, iron and sulphur were determined from the XPS peak areas divided  
137 by atomic sensitivity factors following the Shirley background subtraction. A deconvolution  
138 of the spectra into different components was carried out. Each spectrum was fitted by means  
139 of an iterative least-squares procedure with Gaussian bands. The proportion of each surface  
140 species was then determined as a function of the areas covered by each band. However, a  
141 systematic quantification of the different iron species present in the samples is not presented  
142 here because of the low signal-to-noise ratio in most of the XPS spectra for these peaks. Only  
143 the approximate position of the observed sulphur species will be described below.

144 Surfaces of raw arsenopyrite powder and some powders retrieved at the end of the runs  
145 at different pH were examined by MicroRaman spectroscopy, revealing traces of quartz and  
146 pyrite at the surface. MicroRaman measurements were carried out in back scattering geometry  
147 by using the polarized 514.5 nm line of an Argon-ion laser. Raman scattering measurements  
148 were performed in air at room temperature with a triple spectrometer Jobin-Yvonne Dilor  
149 integrated system with a spectral resolution of about 1 cm<sup>-1</sup>. The power density at the sample  
150 was set between 5 and 10 mW mm<sup>2</sup>. Acquisition time was between 30 and 120 s depending of  
151 the quality of the spectra that were recorded in the Stokes region by a 1200 grooves/mm  
152 grating monochromator and CCD detector system. A confocal microscope Olympus B-201  
153 was used, with an objective 100× with 0.90 numerical aperture. The spatial resolution was less  
154 than 1 μm. In order to verify the homogeneity of the samples and the reproducibility of the  
155 reported data, all the measurements were repeated at different random points of the samples.

## 156 2.2 Solutions and analysis

157 All input solutions were prepared by mixing the respective analytical reagents and Millipore  
158 MQ water (18.2 MΩ · cm). The analytical-grade reagents in the acidic solutions (pH 1 and 3)  
159 were HCl and H<sub>2</sub>SO<sub>4</sub> (95-97%). Reagents FeSO<sub>4</sub>·7H<sub>2</sub>O and H<sub>2</sub>SO<sub>4</sub> (95-97%) were used to  
160 prepare the 0.01M Fe<sup>2+</sup> in a H<sub>2</sub>SO<sub>4</sub> solution. Input solution of pH 5.7 only consisted of  
161 Millipore MQ water (18.2 MΩ · cm). The solution of pH 7.6 was prepared with KH<sub>2</sub>PO<sub>4</sub> and  
162 NaOH and pH 9 solution with Na<sub>2</sub>B<sub>4</sub>O<sub>7</sub>·H<sub>2</sub>O.

163 Total concentrations of As, S and Fe in input and output solutions were analyzed by  
164 Inductively Coupled Plasma Atomic Emission Spectroscopy (ICP-AES, Thermo Jarrel-Ash  
165 with CID detector and a Perkin Elmer Optima 3200 RL). Detection limits for As, Fe and S  
166 were 1.3·10<sup>-6</sup>, 3.6·10<sup>-7</sup> and 3.1·10<sup>-6</sup> mol L<sup>-1</sup>, respectively. The accuracy uncertainty in ICP-AES  
167 measurements was estimated to be around 3%. Ferrous and total dissolved iron concentrations  
168 in output solutions were determined by colorimetry using the ferrozine method (modified after  
169 To et al., 1999) in a UV-VIS HP Spectrophotometer. Fe(III) was taken as the difference  
170 between Fe(tot) and Fe(II). The quality of the results was assured by measuring several  
171 standards, blanks and duplicates. Fe(tot) concentrations matched ICP-AES results within 5%.

172 Input and output solutions pH was measured at experimental temperature on an unstirred  
173 aliquot of solution using a Crison meter combination electrode with temperature  
174 compensation. pH calibration was made with standards of 2, 4, 7 and 9.21 buffer solutions.  
175 The reported accuracy is ±0.02 pH units.

176 Redox potential was measured by an Orion combination Pt/Ag-AgCl redox electrode.  
177 The measurements were referenced to the Standard Hydrogen Electrode (*SHE*). Reliable redox  
178 potential measurements could not be systematically obtained because of considerable drifting  
179 in the recorded values with time. This drifting could be caused by the low concentrations of  
180 aqueous species in the output solutions, by the presence of H<sub>2</sub>S<sub>(g)</sub> or by the absence of a clearly

181 dominant redox couple (Nordstrom, 2000). When it was possible to carry out the  
182 measurements the Eh was between 0.2 and 0.7 V.

### 183 2.3 Flow-through experiments

184 Dissolution experiments were carried out using non-stirred and stirred flow-through  
185 reactors (ca. 35 mL in volume) fully immersed in a thermostatic water-bath held at a constant  
186 temperature of 25, 50 or 70°C ( $\pm 1$ )°C. The reaction cells were composed of two chambers, a  
187 lower chamber of 33-mm inner diameter and an upper chamber of 26-mm inner diameter. The  
188 two chambers were separated by a fine (5  $\mu\text{m}$ ) nylon mesh, on which arsenopyrite powder was  
189 placed. In the stirred experiments, the magnetic bar was in contact with the powder. The  
190 dissolved oxygen concentrations in the experiments at dissolved oxygen conditions lower than  
191 those in equilibrium with a free atmosphere were achieved by enclosing the entire  
192 experimental set-up (input and output solutions, peristaltic pump, flow-through reactors and  
193 tubing) in a glove box purged with the corresponding O<sub>2</sub>/N<sub>2</sub> gas mixture (4.5% O<sub>2</sub> in N<sub>2</sub> for  
194 the experiment with 2.0 mg L<sup>-1</sup> of input dissolved oxygen and pure N<sub>2</sub> for the experiment with  
195 0.2 mg L<sup>-1</sup> of input dissolved oxygen). Input solutions were purged with the same gas  
196 mixtures filling the glove box on a regular basis to ensure the dissolved oxygen  
197 concentrations, which were checked before pumping into the reactors. Experimental oxygen  
198 partial pressure in the glove box was continuously monitored by an oxygen partial pressure  
199 detector with an accuracy of  $\pm 0.1$  % O<sub>2(g)</sub>. The concentration of dissolved oxygen in the  
200 reacting solution was measured at steady state in some representative experiments by  
201 luminescence-based oxygen sensor using a Hach HQ10 portable dissolved oxygen meter.  
202 To ensure that the rate of supply of oxidant (dissolved oxygen) into the cell (i.e., to the  
203 reacting mineral surface) is faster than its consumption by the rate of reaction, a series of  
204 experiments was carried out increasing the flow rate from 0.01-0.05 to 0.20-0.40 mL min<sup>-1</sup>

205 (see Table 1). Such a high increase in flow rate did not yield an increase in dissolution rate.  
 206 This fact suggests that the reaction rate is not affected by ionic transport limitations in the  
 207 surrounding solution.

208 Steady-state conditions were considered to be attained when differences in the metal  
 209 concentration of the output solution were within  $\pm 5\%$  for at least 200 h and four consecutive  
 210 leachate samples. After the experiments, the reacted samples were collected, rinsed with  
 211 double-distilled water, dried at room temperature and stored in closed microvials under  
 212 controlled atmosphere until examination by Scanning Electron Microscopy (SEM) and X-Ray  
 213 Photoelectron Spectroscopy (XPS), and determination of BET surface area. A schematic  
 214 sketch of the experimental setting and some more details of the experimental procedure can be  
 215 found elsewhere (Cama and Acero, 2005; Acero et al., 2007a,b; and Asta et al., 2010).

#### 216 2.4 Calculation of dissolution rates at pH < 6

217 The dissolution rate in steady state,  $Rate_{stst}$ , ( $\text{mol m}^{-2} \text{s}^{-1}$ ) was based on the release of As, Fe  
 218 and S according to the expression:

$$219 \quad (1) \quad Rate_{stst} = \frac{q(C_{j,out} - C_{j,inp})}{v_j A}$$

220 where  $q$  is the fluid volume flux through the system ( $\text{m}^3 \text{s}^{-1}$ ),  $C_{j,inp}$  and  $C_{j,out}$  are the  
 221 concentrations of component  $j$  (As, Fe or S) in the input and the output solutions, respectively  
 222 ( $\text{mol m}^{-3}$ ),  $v_j$  is the stoichiometry coefficient of  $j$  in the dissolution reaction, and  $A$  is the  
 223 reactive surface area ( $\text{m}^2$ ), which in this study is calculated using the final specific surface area  
 224 ( $\text{m}^2 \text{g}^{-1}$ ) and the final mass (g). The error associated with the calculated dissolution rates was  
 225 estimated by the Gaussian error propagation method (Barrante, 1974) to range from 12 to 25  
 226 % and it was dominated by the uncertainty of BET surface area measurements ( $\pm 10\text{-}15\%$ ).

## 2.5 Calculation of dissolution rates at pH > 6

At pH higher than 6, it is expected that Fe(II) released from arsenopyrite dissolution oxidizes quickly to Fe(III) (Singer and Stumm, 1970), and that Fe(III)-bearing phases precipitate on arsenopyrite surface, coating arsenopyrite grains as dissolution proceeds. This process is satisfactorily described by the shrinking core model (SCM) and was already applied to pyrite oxidation at basic pH (Nicholson et al., 1990). The model assumes spherical particle shape and, according to Wen (1968), the process of dissolution-coating can be divided into three successive steps: diffusion of reactant (O<sub>2(aq)</sub> in our case) from the infinite solution to the external surface of the coat, diffusion of O<sub>2(aq)</sub> through the coat volume and, finally, O<sub>2(aq)</sub>-induced dissolution of the unreacted core. O<sub>2(aq)</sub> consumption at the core surface gives rise to a concentration gradient across the coat.

To simplify the system, we assume that diffusion in water is much faster than through the solid coat and that the unreacted core shrinks much more slowly than the time needed to reach steady-state diffusion across the coat. This means that, except in the early steps of dissolution, the coat acts as the slow barrier for the whole process, and the dissolution rate decreases as the coat accumulates.

At the onset of the experiment, when no coat is still developed, surface reaction is the step that controls the overall process. For simplicity, it is assumed that the arsenopyrite dissolution rate (mol m<sup>-2</sup> s<sup>-1</sup>) is linearly dependent on oxygen activity:

$$(2) \quad \text{Rate} = k\rho a_{O_2}$$

where  $k$  is a mass transfer dissolution constant (m s<sup>-1</sup>),  $\rho$  is the molar density of arsenopyrite (37850 mol m<sup>-3</sup>), and  $a_{O_2}$  is the input O<sub>2(aq)</sub> activity. When the coating is developed, and the overall process is controlled by oxygen diffusion through the coating, arsenopyrite dissolution depends on the oxygen flux:

$$251 \quad (3) \quad \text{Rate} = \frac{1}{v_{ox}} D \frac{\partial C}{\partial r}$$

252 where  $v_{ox}$  is O<sub>2</sub> stoichiometry in the arsenopyrite dissolution reaction (see below), and D is the  
 253 effective diffusion coefficient of the coating (m<sup>2</sup> s<sup>-1</sup>). The O<sub>2</sub> concentration gradient can be  
 254 calculated as:

$$255 \quad (4) \quad \frac{\partial C}{\partial r} = \frac{C - C_s}{x}$$

256 Where C is the concentration of O<sub>2</sub> in the pore water, C<sub>s</sub> is the concentration on the core  
 257 surface, which is assumed zero, and x is the coating thickness. According to Wen (1968), the  
 258 time required to react a specified molar fraction of the initial arsenopyrite, X, can be  
 259 calculated on the basis of the step that controls the overall process:

260 - dissolution at the surface of the particle when no coating is developed (at the onset of the  
 261 experiment):

$$262 \quad (5) \quad t = \frac{v_{ox} r_0 \rho}{kC} [1 - (1 - X)^{1/3}]$$

263 - diffusion across the coat:

$$264 \quad (6) \quad t = \frac{v_{ox} r_0^2 \rho}{6DC} [1 - 3(1 - X)^{2/3} + (1 - X)]$$

265 where  $r_0$  is initial radius of the particle (m) and D is the effective diffusion coefficient of the  
 266 coating. The total time of the overall process is obtained by summing the duration of the two  
 267 steps. For each specific time, the molar fraction of arsenopyrite dissolved, X, was calculated  
 268 from the accumulated mass of solute (sulphate or arsenic) released during a time span divided  
 269 by the initial mass of arsenopyrite. The accumulative mass of solute was calculated as the

270 product of concentration per flow rate and per time increment, and the result added to the  
271 solute corresponding to the previous time span. Hence, the values of  $k$  and  $D$  were obtained  
272 for each experiment as their best fit in eqs. (5) and (6). As shown by eqs. (5) and (6), the  
273 values of  $k$  and  $D$  depend on the value of the initial radius  $r_0$ , which was estimated to be  
274 around 10  $\mu\text{m}$ , as corresponds to the average particle size obtained in the determination of  
275 granulometric distribution for the initial unreacted arsenopyrite (see section 2.1).

276

277

### 3 RESULTS

#### 278 3.1 Results based on solution chemistry

279 Variations with time of the output As, Fe and S concentrations in some representative flow-  
280 through experiments with different pH are depicted in Fig. 1. The duration of the experiments  
281 varied from 600 to 4500 h. In the experiments carried out at  $\text{pH} < 6$  steady-state conditions  
282 were attained after 300-1200 h and the duration of steady state lasted for more than 300 h. In  
283 the experiments carried out at  $\text{pH} 7.5$  steady state was not attained, and the output  
284 concentration decreased with time (Fig. 1c). Some of the experiments consisted of two to  
285 three stages with different DO concentrations while keeping pH, temperature and flow rate  
286 constant. The experimental conditions for each experiment are shown in Tables 1 and 2. The  
287 residence time in the reactors was between 1.5 and 20 h, depending on the flow rate (0.4-0.01  
288  $\text{mL min}^{-1}$ ). Calculated total dissolved mass of arsenopyrite throughout the experiments was  
289 usually less than 10 %. As shown in Fig. 1, concentrations of arsenic, iron and sulphur in the  
290 output solutions were highest at the start of the experiments, subsequently decreasing. Initially  
291 high concentrations were probably due to the dissolution of microparticles (Fig. 2a, b), with  
292 surface areas higher than those of the bulk samples, or to the existence of external, altered  
293 layers of the ground mineral (Lasaga, 1998; Acero et al., 2007a,b,c; Acero et al., 2009).

294 According to Borda et al. (2004) high concentrations could be due to preferential dissolution  
295 in the early stages of the experiments. The fact that a mineral dissolves apparently faster at the  
296 start of the experiment and the fact that steady states are attained after some time (in our case  
297 300 to 1200 h) highlight the need to carry out long term flow-through instead of short batch  
298 experiments to predict arsenopyrite dissolution in field scenarios where an extended  
299 interaction with solutions is expected.

300 From pH 1 to 4, sulphur concentrations in the output solutions were usually lower than  
301 As and Fe, yielding S/As lower than one, whereas output concentrations of As and Fe were  
302 practically the same, yielding Fe/As ratio close to one (Table 1). In a few experiments the  
303 Fe/As was higher than one (e.g.,  $1.3 < \text{Fe/As} < 1.8$ ). The deficit in aqueous sulphur when  
304 dissolving arsenopyrite in the pH range 2 to 4.5 in the absence of initial dissolved Fe(III) was  
305 also reported by McKibben et al. (2008), who also observed a deficit in As. Sulphur deficit in  
306 sulphide dissolution in acidic solutions has been described in earlier studies (Lochmann and  
307 Pedlik, 1995; Weisener et al., 2003, 2004; Malmström and Collin, 2004; Acero et al., 2007b);  
308 pyrite (Domènech et al., 2002); pyrrhotite (Janzen et al., 2000); galena (De Giudici and  
309 Zuddas, 2001; Cama and Acero, 2005; Cama et al., 2005). Acero et al. (2007b) during the  
310 collection of acidic output solutions detected a H<sub>2</sub>S<sub>(g)</sub> odour which resulted in aqueous sulphur  
311 deficit. Although it can not be ruled out, this odour was not detected during the collection of  
312 output solutions in our study. Therefore, for experiments at pH 1 to 4, the arsenopyrite  
313 dissolution was computed from the output As and Fe concentration. As discussed below,  
314 aqueous iron was partially depleted ( $\text{Fe/As} < 1$ ) at mildly acidic pH (4.4 to 5.8) and totally  
315 depleted at basic pH (7.5 to 9) ( $[\text{Fe}]_{\text{out}}$  was below the detection limit). Arsenic output  
316 concentrations were lower than sulphur in the neutral to basic pH range (Fig. 1c,d). Therefore,  
317 for experiments at pH higher than 6, the arsenopyrite dissolution was computed from the  
318 output SO<sub>4</sub> concentration.

319 The saturation state of the output solution at the end of each experiment was calculated  
320 using the PHREEQC code (Parkhurst, 1995) and WATEQ database (Ball and Nordstrom,  
321 2001). Data for scorodite (FeAsO<sub>4</sub>·2H<sub>2</sub>O) are those revised by Krause and Ettel (1988). At pH  
322 1-3, output solutions were undersaturated with respect to native sulphur, S-bearing phases and  
323 Fe-oxy-hydroxides. In the pH range of 4.4 to 5.8, the amount of Fe(II) measured was  $\leq 25\%$ .  
324 Hence, with 75% of Fe(III) solutions were supersaturated with respect to scorodite and Fe-  
325 oxy-hydroxide phases, such as amorphous Fe(OH)<sub>3</sub>, goethite and lepidocrocite. At pH > 7,  
326 although aqueous iron was depleted, calculations were run by using a very low iron  
327 concentration ( $10^{-7}$  mol L<sup>-1</sup>), yielding output solutions supersaturated with respect to several  
328 iron oxy-hydroxides. A comparison of SEM photographs of samples before and after reacting  
329 in acidic and basic pH reveals the precipitation of new Fe-bearing phases at neutral and basic  
330 pH (Fig. 2). The formation of secondary iron precipitates on reacted pyrite surfaces at basic  
331 pH has been reported in earlier works (Koslides and Ciminelli, 1992; Bonnissel-Gissinger et  
332 al., 1998; Pérez-López et al., 2007).

333 Figure 3 plots the variation of the arsenopyrite dissolution rate with pH at 25 °C and  
334 dissolved O<sub>2</sub> concentration of 8.7 mg L<sup>-1</sup> in this study and other reported works. According to  
335 our results from pH 1 to 5 the arsenopyrite dissolution rate scarcely changes with pH  
336 regardless of whether the results are based on As or Fe release (average log rate  $\sim -10.1 \pm 0.2$   
337 mol As m<sup>-2</sup> s<sup>-1</sup>; see Table 1). At pH 5-6 the dissolution rate is similar to that at pH < 4. In the  
338 pH range of 7.5 to 9 the dissolution rates increase.

339 The dependence of arsenopyrite dissolution rate on dissolved oxygen concentration was  
340 assessed at acid pH values (Table 1). Figure 4 plots the variation in the output of As, S and Fe  
341 concentrations as dissolved oxygen concentration increased or decreased in both multistage  
342 and single experiments. The output concentrations of As, S and Fe decreased by lowering the  
343 input dissolved oxygen concentration from 8.7 to 4.5 and 0.2 mg L<sup>-1</sup> (Fig. 4). The observed  
344 log DO-log rate dependence suggests that oxidation rate may be proportional to a fractional

345 power of the dissolved-oxygen activity (Fig. 5 a). Steady-state dissolved oxygen concentration  
346 in the reactor in all experiments was calculated based on arsenic and used to derive a rate-DO  
347 dependence in which the rate is proportional to a fractional power of the dissolved oxygen  
348 ( $R = k \cdot a_{O_2}^n$ ) where  $n$  is the reaction order with respect to dissolved oxygen. Remarkably, the  
349 measured data in the representative experiments (Table 1) were the same ( $\pm 0.1 \log DO$ ) as the  
350 calculated DO concentrations. The reaction order with respect to oxygen estimated from  
351 multiple linear regression of the rates was  $0.76 \pm 0.11$ . Alternatively, a non-linear fit to the rate  
352 – DO data was also possible (see Fig. 5 b). It was observed that when DO concentration  
353 exceeded  $1 \times 10^{-4}$  M, the rate became independent of DO concentration.

### 354           **3.2 Results based on arsenopyrite surface spectroscopy**

355           Atomic concentrations of iron, arsenic and sulphur at the reacted and unreacted  
356 arsenopyrite surfaces were determined by the XPS examination of the samples before and  
357 after the flow-through experiments and the obtained results are summarized in Table 3. The  
358 poor quality of the signal in the Fe2p region of the reacted arsenopyrite surfaces prevented the  
359 XPS identification of the iron surface species. The results show that, at acidic pH, arsenopyrite  
360 surface is enriched in arsenic and sulphur, whereas, at pH 4 to 9, the surface is enriched in iron  
361 and arsenic, consistently with the solution results.

362           An examination of the S2p spectra of arsenopyrite reacted at acidic pH and  $8.7 \text{ mg L}^{-1}$  of  
363 DO (Fig. 6a) indicates the existence of three possible species at binding energies (BE) of  
364 approximately 161.3-162.7, 163.5-164.1 and 168.6-169.1. eV. These binding energies were  
365 identified as S<sup>2-</sup>, polysulphides (S<sub>n</sub><sup>2-</sup> where  $n \geq 2$ ), and sulphate, respectively, according to the  
366 values reported in earlier studies (Buckley and Woods, 1985; Mycroft et al. 1990; Nesbitt and  
367 Muir, 1994, 1998; Pratt et al. 1994; Nesbitt et al., 1995; Hacquard et al. 1999). The detection  
368 of elemental S on the surfaces was practically impossible (i.e., this could not be confirmed or

369 ruled out) owing to the technical limitations of the equipment used in the acquisition of the  
370 spectra. The examination of As3d peaks (Fig. 6b) shows a major peak at approximately 45.3  
371 eV, which corresponds to As(V) (Nesbitt and Muir, 1998). The shoulder of the low binding  
372 energy side may indicate the contribution of As(-I) and As(II) species at binding energies of  
373 41.3-41.7 eV and 43.3 eV, respectively (Nesbitt et al., 1995, 1998; Buckley and Walker,  
374 1988). Spectra of arsenopyrite surface after reacting at acidic pH and low DO concentration  
375 (1.8 mg L<sup>-1</sup>) showed that reduced sulphur and arsenic signals increased and the amount of the  
376 most oxidized species (As(V) and sulphates) decreased (Figs. 6 c,d).

377 XPS results in the pH range of 4.5-5.5 (Figs. 6e, f) show changes with respect to the  
378 results obtained at acid pH. First, at this pH the amount of surface iron is higher than in the  
379 initial sample, which can be attributed to the presence of iron precipitates (e.g., Fe(III)-  
380 hydroxides). In the S2p spectra two species are identified with an energy binding of 164.7 and  
381 169.2 eV (Fig. 6e), which correspond to an intermediate oxysulphur (Schaufuss et al., 2000)  
382 and sulphate, respectively. In the case of As3d peak (Fig. 6f) the best fits of the spectra  
383 indicate the presence of a major contribution of As(V) and minor As(III) with peaks at 45.7  
384 and 44 eV, respectively (Fig. 6f). In the range of pH 7-9 (Fig. 6g, h), S2p spectra show that S(-  
385 I) oxidizes mainly to polysulphides, and the As3d spectra show a shoulder with binding  
386 energies in the range of 41.8-42 that may indicate the contribution of As(-I) and a major peak  
387 at approximately 45 eV, which corresponds to As(V) (Nesbitt and Muir, 1998) (Fig. 6g, h).

388 MicroRaman spectra are shown in Fig. 7. Samples reacted at acidic pH (1-3) show that  
389 the most intense peaks are at 472, 219 and 150 cm<sup>-1</sup> (Fig. 7a). These peaks are attributed to  
390 elemental sulphur, whereas the other noticeable peaks are due to iron oxide and As<sub>2</sub>O<sub>3</sub>. This  
391 seems to indicate that, under acidic conditions, native sulphur could be a dominant surface  
392 species along with a minor amount of a stable phase iron oxide. MicroRaman spectra of  
393 samples reacted at pH 7.5 (Fig. 7b) show peaks that are associated with hematite, iron  
394 oxy(hydroxide), probably goethite, as well as traces of As-O and As<sub>2</sub>O<sub>3</sub> (e.g., claudetite)

395 (Table 4). Thus, secondary predominant surface species are pH dependent (e.g., see Brookins,  
396 1988). It should be noted that native sulphur and iron oxy-hydroxides are very efficient Raman  
397 scatterers, i.e. the higher their crystallinity, the higher their Raman efficiency. Raman  
398 scattering involves hundreds of atomic layers below the surface.

399 Hence, taking into account the saturation index of the output solution and the Raman  
400 microanalysis of the solids one can conclude that iron hydroxides such as goethite may be  
401 responsible for Fe depletion at pH higher than 5. Sulphur detection by Raman spectroscopy  
402 shows the formation of elemental sulphur in significant amounts at acidic pH (1-3), although  
403 thermodynamic calculations indicate that it should not precipitate (solutions are  
404 undersaturated).

405 Finally, As depletion at pH > 4 could be attributed to arsenate sorption on the Fe(III)-  
406 hydroxide surface. At pH < 7 the surface is positively charged, and at 7 < pH < 9, although the  
407 net surface charge is negative, positively charged sites are still available (Stumm and Morgan,  
408 1996; Kosmulski, 2004). Accordingly, an estimation of the As sorption onto Fe-oxyhydroxide  
409 at pH range 4-9 was made on the basis of the Generalized Two layer-surface complexation  
410 model (Dzombak and Morel, 1990) using the PHREEQC code, a surface area of 600 m<sup>2</sup> g<sup>-1</sup>  
411 (as described by Dzombak and Morel, 1990 for hydrous ferric oxide) and a surface site density  
412 of 2.3 sites nm<sup>-2</sup> (Davies and Kent, 1990). Given the total amount of Fe-hydroxide, which was  
413 based on SO<sub>4</sub> released (<0.1 g), the maximum amount of As(III) and As(V) sorbed was less  
414 than 10<sup>-9</sup> mol As (at pH < 9.5). This arsenic amount was too low to account for the As that  
415 was retained in the experiments, which was higher than 1/5 Fe (i.e. 10<sup>-4</sup> mol As). Therefore,  
416 the formation of an As phase seems to be responsible for As depletion at pH > 4. Based on the  
417 stoichiometry of our results (e.g., 1 < Fe/As < 2.6; Table 3), scorodite and/or pharmacosiderite  
418 or pitticite could be responsible for As depletion. Beattie and Poling (1987) showed that, at pH  
419 values greater than 7, arsenopyrite oxidation results in the formation of secondary arsenic  
420 minerals such as pitticite [Fe<sub>2</sub>(AsO<sub>4</sub>)(SO<sub>4</sub>)OH·2H<sub>2</sub>O] and pharmacosiderite

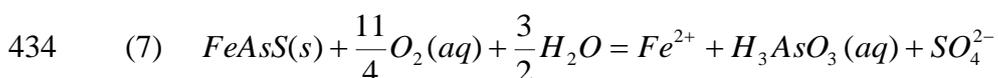
421 (2FeAsO<sub>4</sub>·Fe(OH)<sub>3</sub>·5H<sub>2</sub>O). Similarly, Hacquard et al. (1999) observed the formation of an  
 422 oxidation layer composed of Fe(III) arsenite and arsenate on the arsenopyrite surface after  
 423 reacting with a solution of pH 10.

## 424 4 DISCUSSION

### 425 4.1 Evolution of the arsenopyrite surface during dissolution

426 The arsenopyrite structural formula (FeAsS) has been described as Fe<sup>2+</sup>(AsS)<sup>2-</sup> with a minor  
 427 contribution of Fe<sup>3+</sup>(AsS)<sup>3-</sup> (Nesbitt et al., 1995; Shuey, 1975). Arsenic and sulphur form a  
 428 dianionic group by means of a covalent bond. During dissolution, iron is mainly released as  
 429 Fe<sup>2+</sup>, and arsenic and sulphur as S<sup>-1</sup> and As<sup>-1</sup>, respectively, which are not stable species in  
 430 solution with the result that oxidation probably occurs.

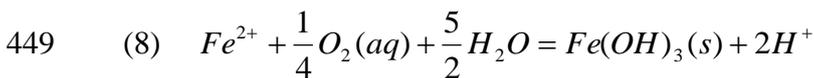
431 Inspection of the reacted samples confirmed that the arsenopyrite surface undergoes  
 432 critical variation as a function of pH, which influenced the overall arsenopyrite oxidative  
 433 dissolution. At pH < 4 the overall oxidative dissolution can be simplified as (Yu et al., 2007):



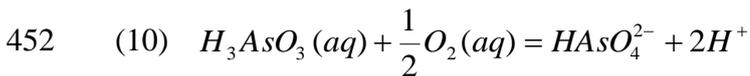
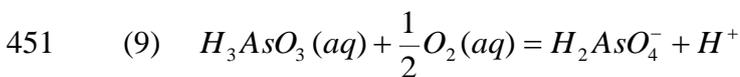
435 Hence, steady-state dissolution rates were obtained at pH < 4 based on As and Fe release  
 436 normalized with respect to final BET specific surface area (Table 1). The dissolution rates  
 437 obtained in the experiments carried out in H<sub>2</sub>SO<sub>4</sub> (pH 1 and 3) were the same, within error, as  
 438 the rates obtained in HCl. This agreement shows that the arsenopyrite rates obtained in this  
 439 study are applicable to acidic environments with sulphate as the main anionic species (e.g.,  
 440 systems affected by acid mine drainage). Furthermore, the effect of Fe(II) on arsenopyrite rate  
 441 was examined at pH 1 and 8.7 mg L<sup>-1</sup> of DO. The results showed that Fe(II) exerts little  
 442 influence on the arsenopyrite dissolution rate.

443 From pH 5 to 6, iron depletion was incomplete, yielding Fe/As and Fe/S aqueous ratios  
 444 lower than one (Table 1). Likewise, incomplete arsenic depletion was observed, giving 0.81  
 445 <S/As < 1.38. At pH ranging from 7 to 9, aqueous iron was completely depleted (below the  
 446 detection limit), and As depletion was higher than that at pH 5-6.

447 Therefore, at pH > 6 released ferrous iron rapidly oxidized to ferric iron, which  
 448 precipitated as Fe-solid phases:



450 and As(III) oxidized to As(V) (eq. 9 and 10) (Walker et al., 2006):



453 To simplify, the variable S/As aqueous ratio can be interpreted as variable proportions in  
 454 the formation of scorodite and Fe(OH)<sub>3</sub> phases:



456 The amorphous Fe(OH)<sub>3</sub>(s) may represent ferrihydrite, which could transform into hematite  
 457 and/or goethite as pH increased (Schwertmann and Murad, 1983). This is a simplification and  
 458 mixed phases as pharmacosiderite can also be possible.

459 The precipitated Fe(III) and As-bearing phases form a coat on the arsenopyrite grains as  
 460 dissolution proceeds. As stated above, this process was modelled using the shrinking core

461 model (SCM). Thus, the limiting process of arsenopyrite dissolution at pH > 6 is the diffusion  
462 of O<sub>2(aq)</sub> through the coating, resulting in the dissolution of the unreacted core of arsenopyrite.  
463 An important parameter that governs dissolution under the SCM is the stoichiometry of O<sub>2(aq)</sub>  
464 in the arsenopyrite dissolution reaction ( $v_{ox}$ ), which according to the addition of eqs. 8, 9 or 10  
465 (depending on pH) and 11 to eq. 7 is 15/4.

466 An example of application of the SCM that simulates the variation in sulphate  
467 concentration versus time at pH 7.5 and 0.27 mol m<sup>-3</sup> O<sub>2(aq)</sub> is depicted in Fig. 8. As expected,  
468 the early values are sensitive to k value, whereas the influence of D increases with time. The  
469 effect of coating can be observed in the plot. After 250 h only 0.12 mol % of initial  
470 arsenopyrite was dissolved. However, if no coating had been formed the dissolution of  
471 arsenopyrite would have been 0.24 mol %. The value of the effective diffusion coefficient of  
472 the coating (m<sup>2</sup> s<sup>-1</sup>) (D) estimated for the experiments that follow the SCM pattern varies  
473 within a range from 10<sup>-17</sup>-to 1.5·10<sup>-16</sup> m<sup>2</sup> s<sup>-1</sup>. No variation with temperature, pH or O<sub>2(aq)</sub>  
474 concentration is apparent. In pyrite dissolution experiments, Nicholson et al. (1988) estimated  
475 a D value around 3·10<sup>-16</sup> m<sup>2</sup> s<sup>-1</sup> for the diffusion coefficient of O<sub>2</sub> through the Fe-hydroxide  
476 coating, and Huminicki and Rimstidt (2009) reported a value of 3.6·10<sup>-15</sup> m<sup>2</sup> s<sup>-1</sup> for the  
477 diffusion coefficient of H<sub>2</sub>O<sub>2</sub>. Both values are comparable or higher than those obtained from  
478 our experiments. One reason for the discrepancy could be the dependence of D on r<sub>0</sub><sup>2</sup> (the  
479 value assumed for initial radius, eq. 6), which in our case is smaller than that reported in the  
480 pyrite case. Furthermore, the D values estimated here are 7 to 8 orders of magnitude lower  
481 than 2·10<sup>-9</sup> m<sup>2</sup> s<sup>-1</sup>, the diffusion coefficient of O<sub>2(aq)</sub> in free water at 25°C (Han and Bartels,  
482 1996). This difference is too large to be attributed to porosity and tortuosity of a porous  
483 medium. On the other hand, these D values are much higher than the diffusion coefficient  
484 values that are typical of solids (10<sup>-20</sup> to 10<sup>-34</sup> m<sup>2</sup> s<sup>-1</sup>, Levine, 1978), but could be similar to  
485 those of poorly crystalline solids as suggested by Nicholson et al. (1988).

486

487 **4.2 The effect of pH on dissolution rates**

488 We suggest that a hydrogen ion effect on the arsenopyrite dissolution rate in this pH range is  
489 very low (Table 1; Fig. 3). This result is consistent with the insignificant pH dependence of  
490 the arsenopyrite-rate reported by Yu et al (2007) in the pH range of 2-6 (e.g., nearly zero-order  
491 dependence on pH), but is in disagreement with the pH-arsenopyrite-rate dependence reported  
492 by McKibben et al. (2008) in the range of pH 2-4.5 ( $\text{Rate}=k \cdot (\text{M}_{\text{H}^+})^{0.3}$ ).

493 A slight increase of the dissolution rates is observed as pH increased. A similar basic pH-rate  
494 tendency was observed in pyrite oxidation by Williamson and Rimstidt (1994), Ciminelli and  
495 Osseo-Asare (1995a,b). In the case of arsenopyrite oxidation, Yu et al. (2007) obtained  
496 apparent arsenopyrite dissolution rates that reached a minimum near pH 7-8 followed by an  
497 increase at higher pH values.

498 **4.3 The effect of dissolved oxygen on dissolution rates**

499 The rates obtained at 25 °C at acidic pH are DO-dependent, decreasing when the  
500 dissolved oxygen concentration is diminished (Figs. 4 and 5). This type of dependence has  
501 been reported for arsenopyrite at acidic pH (Yu et al., 2007; McKibben et al., 2008) and for  
502 pyrite (Nicholson et al., 1988; Williamson and Rimstidt, 1994; Domènech et al, 2002). The  
503 reaction order of the arsenopyrite dissolution rate with respect to dissolved oxygen at acidic  
504 pH was 0.76. This value is higher than those reported by Yu et al. (2007) ( $0.45 \pm 0.05$ ) at pH  
505 5.9 and McKibben et al. (2008) ( $0.33 \pm 0.18$ ) at pH 2-4.5 for arsenopyrite, and by McKibben  
506 (1984), Williamson and Rimstidt (1994) and Domènech et al. (2002) for pyrite ( $0.4-0.5 \pm 0.04$ )  
507 at pH 2-10, suggesting that arsenopyrite dissolution is more strongly influenced by DO than  
508 pyrite dissolution. In contrast to these results, Walker et al. (2006) reported rate-DO  
509 independence in the pH range from 6.3 to 6.7.

510 An alternative, non-linear fit to the data was also possible (see Fig.5 b), considering  
 511 that when DO concentration was higher than  $1 \times 10^{-4}$  M, the rate varied barely, becoming thus  
 512 DO-independent. As it was reported by Nicholson *et al.* (1988) and Domènech *et al.* (2002)  
 513 for pyrite dissolution, this behaviour could be consistent with a surface mechanism that  
 514 involves equilibrium adsorption-desorption. Thus, according to the non-linear behaviour, the  
 515 dissolution rate-DO-dependence can be expressed as:

$$516 \quad (12) \quad R = r_m \frac{K \cdot [O_2]}{1 + K \cdot [O_2]}$$

517 where  $R$  is the dissolution rate ( $\text{mol m}^{-2} \text{s}^{-1}$ ),  $r_m$  ( $\text{mol m}^{-2} \text{s}^{-1}$ ) is the rate based on  
 518 maximum saturation of available surface sites,  $K$  ( $\text{L mol}^{-1}$ ) is the adsorption equilibrium  
 519 constant and  $[O_2]$  is the dissolved oxygen concentration ( $\text{mol L}^{-1}$ ). The best fit to the  
 520 arsenopyrite data was obtained with the following values:  $r_m = 1.2 \pm 0.1 \cdot 10^{-10} \text{ mol m}^{-2} \text{ s}^{-1}$  and  
 521  $K = 1 \cdot 10^4 \text{ L mol}^{-1}$ .

522

#### 523 4.4 The effect of temperature on the dissolution rate

524 The temperature dependence of dissolution rate generally follows the Arrhenius law:

$$525 \quad (13) \quad \text{Rate} = A e^{-E_{app}/RT}$$

526 where  $A$  is the pre-exponential factor,  $E_{app}$  is the apparent activation energy,  $R$  is the gas  
 527 constant and  $T$  is the temperature (K). To obtain experimentally the apparent activation energy  
 528 at different pH, experiments were carried out at 25, 50 and 70°C at acid pH by maintaining  
 529 constant both the pH and dissolved O<sub>2</sub> concentration (Table 1 and Fig. 9). The apparent  
 530 activation energy for arsenopyrite oxidation by oxygen at acid pH was  $18.5 \text{ kJ mol}^{-1}$ . Yu et al.  
 531 (2007) using short-term flow reactors (6-8 h) reported  $E_{app}$  values ( $\text{kJ mol}^{-1}$ ) of 43 (pH 1.8)  
 532 and 57 (pH 5.9). McKibben et al. (2008) carrying out short-term batch experiments at pH 2-  
 533 4.5 obtained a complicated non-Arrhenius behaviour similar to that reported by Rimstidt et al.

534 (1994) for arsenopyrite dissolution in iron-rich solutions (from 0 to 25 °C,  $E_{app} = 18 \text{ kJ mol}^{-1}$   
535 and from 25 to 60 °C,  $E_{app} = -6 \text{ kJ mol}^{-1}$ ). Our  $E_{app}$  values suggest that the overall arsenopyrite  
536 dissolution is a mixture of surface and diffusion controlled processes. Further research using  
537 rotating disk experiments would be necessary to distinguish between different reaction  
538 mechanisms.

#### 539 **4.5 Arsenopyrite dissolution rates**

540 A comparison of the arsenopyrite dissolution rates obtained in this study with previously  
541 reported values is shown in Fig. 3. Table 5 gives the dissolution rates by abiotic oxidation by  
542 DO and kinetic parameters obtained in our work and in reported studies of arsenopyrite  
543 (Walker et al., 2006; Yu et al., 2007; McKibben et al., 2008). The rates of arsenopyrite  
544 dissolution obtained in this work are approximately two-three orders of magnitude slower than  
545 those reported by Yu et al. (2007) and McKibben et al. (2008), but in good agreement to that  
546 reported by Walker et al. (2006). It is important to highlight that different experimental setups  
547 and experimental conditions were used in these studies. Note that sulphide dissolution rates  
548 obtained in very short-term experiments (< 50 h) are usually one order of magnitude faster  
549 (Yu et al., 2007; McKibben et al., 2008). Indeed, on the onset of our long-term experiments  
550 apparent dissolution rates were faster than steady-state rates. Thus, the advantage of the long-  
551 term experiments is that achievement of durable steady state (> 300 h) guarantees the reaction  
552 to proceed under steady mineral-solution conditions. In line with this explanation, the  
553 dissolution rates obtained in this work are in good agreement with the rates obtained for other  
554 sulphides such as pyrite (Domènech et al. 2002), chalcopyrite (Acero et al., 2007a, 2009),  
555 sphalerite (Acero et al., 2007b), galena (Acero et al., 2007c) and marcasite (Asta et al., 2010)  
556 based on long-term experiments. Therefore, we suggest that quantification of arsenopyrite  
557 oxidation at acidic mine wastes is appropriately obtained by the long-term experiments.

558

## 5 CONCLUSIONS

559 In the present study, the surface chemistry and arsenopyrite dissolution were studied by means  
560 of stirred and non-stirred long term flow through experiments at a range of pH 1-9, at  
561 temperatures from 25 to 70 °C at three different dissolved oxygen concentrations and under  
562 far-from equilibrium conditions.

563 On the one hand, we have shown that under the studied experimental conditions,  
564 arsenopyrite dissolution is strongly affected by dissolved oxygen and slightly affected by pH  
565 and temperature. At acidic pH, stirred and non-stirred experiments yield the same dissolution  
566 rate, and the rise in temperature yields an apparent activation energy of  $18.5 \pm 1.6 \text{ kJ mol}^{-1}$  that  
567 suggests that arsenopyrite dissolution is a mixture of surface and diffusion-controlled  
568 processes. On the other hand, the arsenopyrite surface chemistry analysis showed that, under  
569 acidic conditions arsenopyrite dissolution creates a sulphur-enriched surface layer. This layer  
570 is mainly made up by polysulphides and sulphates although the presence of elemental sulphur  
571 cannot be ruled out. However, these layers do not exert any passivating effect once the steady  
572 state is attained. At mildly acid to basic pH, precipitation of Fe phases takes place. Fe-coating  
573 grows on the mineral surface and prevents the diffusion of aqueous species through it. Thus,  
574 steady state is not attained and the output concentration (and As release) decreases with time.

575 These results are representative of the conditions found in remediated AMD sites using  
576 passive treatments where reacting limestone increases pH to ca. 7: Fe-layers passivate the  
577 surface of sulphides, and in addition, the high stability of Fe oxy(hydroxide) at circumneutral  
578 pH provides a considerable retention capacity of toxic metal(oid)s, such as arsenic, that are in  
579 non-admissible levels in AMD.

580 This aqueous and solid chemistry-based study represents a firm step towards  
581 understanding both the long-term behaviour of arsenopyrite dissolution and the implication  
582 that arsenopyrite oxidative dissolution can have in arsenic contamination into distinct waters,

583 showing the effects that dissolved oxygen and pH have on the complexity of arsenopyrite  
584 dissolution kinetics.

585

586

587       *Acknowledgments.* This research was supported by the project CTM2007-66724-C02-  
588 01/TECNO from the Spanish Government. MPA was financially supported by the Spanish  
589 Government with a PhD fellowship. PA was financially supported by the Spanish Government  
590 with a research contract from the “Juan de la Cierva” program. Albert Soler and Àngels  
591 Canals from the Mineralogical Dept. of Barcelona University supplied the arsenopyrite  
592 samples. We wish to express our gratitude to Javier Pérez, Vanessa Ouro, Rafel Bartrolí, Eva  
593 Pelegrí, Eva Prats, Ana Dominguez and Josep Elvira for their technical assistance. We are  
594 indebted to Lorenzo Calvo for his assistance in XPS analyses and to Xavier Llovet who  
595 assisted in the obtainment of EMP analyses at the Technical Services of Barcelona University.  
596 We also gratefully acknowledge the helpful comments and valuable suggestions made by Dr.  
597 M.A. McKibben of a previous version of this manuscript. We are grateful to Dr. Josep M.  
598 Soler, Dr. J.D. Rimstidt, two anonymous reviewers and associate editor Jeremy Fein for their  
599 insightful comments. We thank to George Von Knorring for improving the English style of  
600 this paper.

601

602

603

**6 REFERENCES**

- 604 Acero, P., Ayora, C., Torrentó, C., Nieto, J.M., 2006. The role of trace elements during  
605 schwertmannite precipitation and subsequent transformation into goethite and jarosite.  
606 *Geochim. Cosmochim. Acta* 70, 4130-4139.
- 607 Acero, P., Cama, J., Ayora, C., 2007a. Kinetics of chalcopyrite dissolution at pH 3. *Eur. J.*  
608 *Mineral.* 19, 173-182.
- 609 Acero, P., Cama, J., Ayora, C., 2007b. Sphalerite dissolution kinetics in acidic environment.  
610 *Appl. Geochem* 22, 1872-1883.
- 611 Acero, P., Cama, J., Ayora, C., 2007c. Rate law for galena dissolution in acidic environment.  
612 *Chem. Geol.* 245, 219-229.
- 613 Acero, P., Cama, J., Ayora, C., Asta, M.P. 2009. Chalcopyrite dissolution rate law from pH 1  
614 to 3. *Geol. Acta* 7, 389-397.
- 615 Anawar, H.M., Akai, J., Komaki, K., Terao, H., Yoshioka, T., Ishizuka, T., Safiullah, S., Kato,  
616 K., 2003. Geochemical occurrence of arsenic in groundwater of Bangladesh: sources and  
617 mobilization processes. *J. Geochem. Explor.* 77, 109-131.
- 618 Aposhian, H.V., Zakharyan, R.A., Avram, M.D., Sampayo-Reyes, A., Wollenberg, M.L.,  
619 2004. A review of the enzymology of arsenic metabolism and a new potential role of  
620 hydrogen peroxide in the detoxication of the trivalent arsenic species. *Toxicol. Appl.*  
621 *Pharmacol.* 198, 327-335.
- 622 Asta, M.P., Ayora C., Roman-Ross G., Cama, J., Acero P. , Gault, A.G., Charnock, J.M., and  
623 Bardelli, F. 2009. Natural attenuation of arsenic in the Tinto Santa Rosa acid stream  
624 (Iberian Pyritic Belt, SW Spain): the role of iron precipitates. *Chemical Geology*, 271, 1-  
625 12.
- 626 Asta, M.P., Cama, J., Acero, P. 2010. Dissolution kinetics of marcasite at acidic pH. *Eur. J.*  
627 *Mineral.* doi: 10.1127/0935-1221/2010/0022-1981.
- 628 Ball, J., Nordstrom, D.K., 2001. User's manual for WATEQ4F with revised thermodynamic  
629 database and test cases for calculating speciation of major, trace and redox elements in  
630 natural waters. U.S. Geological Survey Water-Resources Investigation Report 91-183.
- 631 Barrante, J.R., 1974. *Applied Mathematics for Physical Chemistry*. Prentice-Hall, Inc.
- 632 Beattie, M.J.V., Poling, G.W., 1987. A Study of the Surface Oxidation of Arsenopyrite Using  
633 Cyclic Voltammetry. *Int. J. Miner. Process.* 20(1-2), 87-108.
- 634 Bersani, D., Lottici, P.P., Montenero, A., 1999. Micro-raman investigation of iron oxide films  
635 and powders produced by sol-gel syntheses. *J. Raman Spectrosc.* 30, 355-360.
- 636 Bhattacharya, P., Chatterjee, D., Jacks, G., 1997. Occurrence of arsenic contaminated  
637 groundwater in alluvial aquifers from Delta Plains, eastern India: options for safe drinking  
638 water supply. *Int. J. Water. Resour Manage.* 13, 79-92.

- 639 Bhattacharya, P., Jacks, G., Jana, J., Sracek, A., Gustafsson, J.P., Chatterjee, D. 2001.  
640 Geochemistry of the Holocene alluvial sediments of Bengal Delta Plain from West Bengal,  
641 India: implications on arsenic contamination in groundwater. In *Groundwater Arsenic*  
642 *Contamination in the Bengal Delta Plain of Bangladesh* (ed. G. Jacks, P. Bhattacharya, and  
643 A.A. Khan), pp. 21-40. TRITA-AMI Report 3084, KTH Spec Pub, KTH.
- 644 Bhattacharya, P., Jacks, G., Ahmed, K.M., Khan, A.A., Routh, J., 2002. Arsenic in  
645 groundwater of the Bengal Delta Plain aquifers in Bangladesh. *Bull. Env. Cont. Toxicol.*  
646 69, 538-545.
- 647 Bhattacharya, P., Ahmed, K. M., Hasan, M.A., Broms, S., Fogelström, J., Jacks, G., Sracek,  
648 O., von Brömssen, M., Routh, J., 2006. Mobility of arsenic in groundwater in a part of  
649 Brahmanbaria district, NE Bangladesh. In *Managing Arsenic in the Environment: From*  
650 *Soil to Human Health* (ed. E. S. R. Naidu, G. Owens, P. Bhattacharya, P.Nadebaum), pp.  
651 95-115. CSIRO, Melbourne, Australia.
- 652 Bonnissel-Gissingier, P., Alnot, M., Ehrhardt, J. J., and Behra, P., 1998. Surface oxidation of  
653 pyrite as a function of pH. *Environ. Sci. Technol.* 32, 2839-2845.
- 654 Borda, M. J., Strongin, D.R., Schoonen, M.A., 2004. A vibrational spectroscopic study of the  
655 oxidation of pyrite by molecular oxygen. *Geochim. et Cosmochim. Acta* 68, 1807-1813.
- 656 Breed, A.W., Harrison, S.T.L., Hansford, G.S., 1997. A preliminary investigation of the ferric  
657 leaching of a pyrite/arsenopyrite flotation concentrate. *Miner. Eng.* 10, 1023-1030.
- 658 Bonnissel-Gissingier, P., Alnot, M., Ehrhardt, J.J., Behra, P., 1998. Surface oxidation of pyrite  
659 as a function of pH. *Environ. Sci. Technol.* 32, 2839-2845.
- 660 Brookins, D.G., 1988. Eh-pH diagrams for geochemistry. Springer-Verlag, Berlin.
- 661 Buckley, A.N., Walker, G.W., 1988. The surface-composition of arsenopyrite exposed to  
662 oxidizing environments. *Appl. Surf. Sci.* 35, 227-240.
- 663 Buckley, A.N., Woods, R., 1985. X-Ray photoelectron spectroscopy of oxidized pyrrhothite  
664 surfaces. I Expose to air. *Appl. Surf. Sci.* 22, 280-287.
- 665 Bunnell, J.E., Finkelman, R.B., Centeno, J.A., Selinus, O., 2007. Medical geology: a globally  
666 emerging discipline. *Geol. Acta* 5, 273-281.
- 667 Cama, J., Acero, P., 2005. Dissolution of minor sulphides present in a pyritic sludge at pH 3  
668 and 25 °C. *Geol. Acta* 3, 15-26.
- 669 Cama, J., Acero, P., Ayora, C., Lobo, A., 2005. Galena surface reactivity at acidic pH and  
670 25°C based on flow-through and in situ AFM experiments. *Chem. Geol.* 214, 309-330.
- 671 Casiot, C., Leblanc, M., Bruneel, O., Personne, J.C., Koffi, K., Elbaz-Poulichet F.O., 2003.  
672 Geochemical processes controlling the formation of As-rich waters within a tailings  
673 impoundment (Carnoulès, France). *Aquat. Geochem.* 9, 273-290.
- 674 Casiot, C., Lebrun, S., Morin, G., Bruneel, O., Personne, J.C., Elbaz-Poulichet, F.O., 2005.  
675 Sorption and redox processes controlling arsenic fate and transport in a stream impacted  
676 by acid mine drainage. *Sci. Total Environ.* 347, 122-130.
- 677 Ciminelli, V.S.T., Osseo-Asare, K., 1995a. Kinetics of pyrite oxidation in sodium carbonate  
678 solutions. *Metall. Mater. Trans. B* 26, 209-218.

- 679 Ciminelli V.S.T., Osseo-Asare K., 1995b. Kinetics of Pyrite Oxidation in Sodium-Hydroxide  
680 Solutions. Metall. Mater. Trans. B 26, 677-685.
- 681 Corkhill, C.L., Vaughan, D.J., 2009. Arsenopyrite oxidation - A review. Appl. Geochem. 24,  
682 2342-2361.
- 683 Costa, M.C., Botelho Do Rego, A.M., Abrantes, L.M., 2002. Characterization of a natural and  
684 an electro-oxidized arsenopyrite: A study on electrochemical and X-ray photoelectron  
685 spectroscopy. Int. J. Miner. Process. 65, 83-108.
- 686 Craw, D., Falconer, D., Youngson, J.H., 2003. Environmental arsenopyrite stability and  
687 dissolution: theory, experiment, and field observations. Chem. Geol. 199, 71-82.
- 688 Davies, J.A., Kent, D.B., 1990. Surface complexation modelling in aqueous geochemistry.  
689 Rev. Mineral. Geochem. 23, 177-260.
- 690 De Giudici, G., Zuddas, P., 2001. In situ investigation of galena dissolution in oxygen  
691 saturated solution: evolution of surface features and kinetic rate. Geochim. Cosmochim.  
692 Acta 65, 1381-1389.
- 693 Domènech, C., de Pablo, J., Ayora, C., 2002. Oxidative dissolution of pyritic sludge from the  
694 Aznalcollar mine (SW Spain). Chem. Geol. 190, 339-353.
- 695 Dowling, C.B., Poreda, R.J., Basu, A.R., Peters, S.L., 2002. Geochemical study of arsenic  
696 release mechanisms in the Bengal Basin groundwater. Water Resour. Res. 38, 1-18.
- 697 Dzombal, D.A., Morel, F.M.M., 1990. Surface complexation modeling. Wiley-Interscience,  
698 New York.
- 699 Ferguson, J.E., 1990. The heavy elements-chemistry, environmental impact and health effects.  
700 Pergamon Press.
- 701 Flynn, E.J., Solin, S.A., Papatheodorou, G.N., 1976. Vibrational excitations of As<sub>2</sub>O<sub>3</sub>. II.  
702 crystalline phases. Phys. Rev. B 13, 1752-1758.
- 703 Frau, F., Ardaù, C., 2003. Geochemical controls on arsenic distribution in the Baccu Locci  
704 stream catchment (Sardinia, Italy) affected by past mining. Appl. Geochem. 18, 1373-  
705 1386.
- 706 Hacquard, E., Bessiere, J., Alnot, M., Ehrhardt, J.J., 1999. Surface spectroscopic study of the  
707 adsorption of Ni(II) on pyrite and arsenopyrite at pH 10. Surf. Interface Anal. 27, 849-860.
- 708 Han, P., Bartels, D.M., 1996. Temperature dependence of oxygen diffusion in H<sub>2</sub>O and D<sub>2</sub>O.  
709 J. Phys. Chem., 100, 5597-5602.
- 710 Hernández-García, M.E., Custodio, E., 2004. Natural baseline quality of Madrid Tertiary  
711 detrital aquifer groundwater (Spain): A basis for aquifer management. Environ. Geol. 46,  
712 173-188.
- 713 Huminicki, D. M. C. and Rimstidt, J. D., 2009. Iron oxyhydroxide coating of pyrite for acid  
714 mine drainage control. Applied Geochemistry 24, 1626-1634.
- 715 Janzen, M.P., Nicholson, R.V., Scharer, J.M., 2000. Pyrrhotite reaction kinetics: Reaction  
716 rates for oxidation by oxygen, ferric iron, and for nonoxidative dissolution. Geochim.  
717 Cosmochim. Acta 64, 1511-1522.

- 718 Koslides, T., Ciminelli, V.S.T., 1992. Pressure Oxidation of Arsenopyrite and Pyrite in  
719 Alkaline-Solutions. *Hydrometallurgy* 30, 87-106.
- 720 Kosmulski, M., 2004. pH dependent surface charging and points of zero charge II. Update. *J.*  
721 *Colloid and Interface Sci.* 275, 214-224.
- 722 Krause, E., Ettel, V.A., 1988. Solubility and stability of scorodite, FeAsO<sub>4</sub>·2H<sub>2</sub>O: New data  
723 and further discussion. *Am. Mineral.* 73, 850-854.
- 724 Lasaga, A.C., 1998. *Kinetic Theory in the Earth Sciences*. Princeton University Press, New  
725 Jersey.
- 726 Lazareva, E.V., Shuvaeva, O.V., Tsimbalist, V.G., 2002. Arsenic speciation in the tailings  
727 impoundment of a gold recovery plant in Siberia. *Geochem.: Explor. Environ. Anal.* 2,  
728 263-268.
- 729 Lee, J.S., Chon, H.T., 2006. Hydrogeochemical characteristics of acid mine drainage in the  
730 vicinity of an abandoned mine, Daduk Creek, Korea. *J. Geochem. Explor.* 88, 37-40.
- 731 Lee, P.K., Kang, M.J., Choi, S.H., Touray, J.C., 2005. Sulphide oxidation and the natural  
732 attenuation of arsenic and trace metals in the waste rocks of the abandoned Seobo tungsten  
733 mine, Korea. *Appl. Geochem.* 20, 1687-1703.
- 734 Lengke, M.F., Sanpawanitchakit, C., Tempel, R.N., 2009. The oxidation and dissolution of  
735 arsenic-bearing sulfides. *Can. Mineral.*, 47, 593-613.
- 736 Levine, I.N., 1978. *Physical Chemistry*. McGraw Hill, New York.
- 737 Lochmann, J., Pedlik, M., 1995. Kinetic Anomalies of Dissolution of Sphalerite in Ferric  
738 Sulphate-Solution. *Hydrometallurgy* 37, 89-96.
- 739 Malmström, M.E., Collin, C., 2004. Sphalerite weathering kinetics: effect of pH and particle  
740 size. *Proc. 11th Symp. Water-Rock Interaction*, 849-852.
- 741 Martens, W., Frost, R.L., Klopogge, J.T., 2003a. Raman spectroscopy of synthetic erythrite,  
742 partially dehydrated erythrite and hydrothermally synthesized dehydrated erythrite. *J.*  
743 *Raman Spectrosc.* 34, 90-95.
- 744 Martens, W., Frost, R.L., Klopogge, J.T., Williams, P.A., 2003b. The basic copper arsenate  
745 minerals olivenite, cornubite, cornwarlite, and clinoclase: An infrared emission and raman  
746 spectroscopic study. *Am. Mineral.* 88, 501-508.
- 747 McArthur, J.M., Ravencroft, P., Safiullah, S., Thirlwall, M.F., 2001. Arsenic in groundwater:  
748 testing pollution mechanism for sedimentary aquifers in Bangladesh. *Water Resour. Res.*  
749 37, 109-117.
- 750 McGuire, M.M., Banfield, J.F., Hamers, R.J., 2001a. Quantitative determination of elemental  
751 sulphur at the arsenopyrite surface after oxidation by ferric iron: mechanistic implications.  
752 *Geochem. Trans.* 2, 25-29.
- 753 McGuire, M.M., Jallad, K.N., Ben-Amotz, D., Hamers, R.J., 2001b. Chemical mapping of  
754 elemental sulphur on pyrite and arsenopyrite surfaces using near-infrared raman imaging  
755 microscopy. *Appl. Surf. Sci.* 178, 105-115.

- 756 McKibben, M.A., 1984 Kinetics of aqueous oxidation of pyrite by ferric ion, oxygen and  
757 hydrogen peroxide from pH 1-4 and 20-40 °C. Ph.D. Thesis, Pennsylvania State  
758 University, U.S.A.
- 759 McKibben, M.A., Tallant, B.A., Del Angel, J.K., 2008. Kinetics of inorganic arsenopyrite  
760 oxidation in acidic aqueous solutions. *Appl. Geochem.* 23, 121-135.
- 761 Mikhlin, Y.L., Romanchenko, A.S., Asanov, I.P., 2006. Oxidation of arsenopyrite and  
762 deposition of gold on the oxidized surfaces: A scanning probe microscopy, tunneling  
763 spectroscopy and XPS study. *Geochim. Cosmochim. Acta* 70, 4874-4888.
- 764 Mycroft, J.R., Bancroft, G.M., McIntyre, N.S., Lorimer, J.W., Hill, I.R., 1990. Detection of  
765 Sulphur and Polysulphides on Electrochemically Oxidized Pyrite Surfaces by X-Ray  
766 Photoelectron-Spectroscopy and Raman-Spectroscopy. *J. Electroanal. Chem.* 292, 139-  
767 152.
- 768 Nesbitt, H.W., Muir, I.J., 1994. X-Ray Photoelectron Spectroscopic Study of a Pristine Pyrite  
769 Surface Reacted with Water-Vapor and Air. *Geochim. Cosmochim. Acta* 58, 4667-4679.
- 770 Nesbitt, H.W., Muir, I.J., 1998. Oxidation states and speciation of secondary products on  
771 pyrite and arsenopyrite reacted with mine waste waters and air. *Mineral. and Petrol.* 62,  
772 123-144.
- 773 Nesbitt, H.W., Muir, L.J., Pratt, A.R., 1995. Oxidation of Arsenopyrite by Air and Air-  
774 Saturated, Distilled Water, and Implications for Mechanism of Oxidation. *Geochim.*  
775 *Cosmochim. Acta* 59(9), 1773-1786.
- 776 Nicholson, R.V., Gilham, R.V., Reardon, E.J., 1988. Pyrite oxidation in carbonate-buffered  
777 solutions:1. experimental kinetics. *Geochim. Cosmochim. Acta* 52, 1077-1085.
- 778 Nicholson, R.V., Gilham, R.V., Reardon, E.J., 1990. Pyrite oxidation in carbonate-buffered  
779 solutions: 2. rate control by oxide coatings. *Geochim. Cosmochim. Acta* 54, 395-402.
- 780 Nickson, R., McArthur, J., Burgess, W., Ahmed, K.M., Ravenscroft, P., Rahman, M., 1998.  
781 Arsenic poisoning of groundwater in Bangladesh. *Nature* 395, 338.
- 782 Nordstrom, D., 2000. Aqueous Redox Chemistry and the Behaviour of Iron in Acid Mine  
783 Waters. In *Proceedings of the 18 Workshop on Monitoring Oxidation-Reduction Processes*  
784 *for Groundwater Restoration.* (eds. R.T. Wilkin, R.D. Ludwig and R.G. Ford), pp. 43-47.  
785 Environmental Protection Agency, EPA /600/R-02/002.
- 786 Parkhurst, D., 1995. User's guide to PHREEQC: A computer program for speciation, reaction  
787 path, advective-transport, and inverse geochemical calculations. US Geological Survey  
788 Water Resources Investigation Report, 95-4227, Lakewood, Colorado, 143 pp.
- 789 Pérez-López, R., Cama, J., Nieto, J.M., Ayora, C., 2007. The iron-coating role on the  
790 oxidation kinetics of a pyritic sludge doped with fly ash. *Geochim. Cosmochim. Acta* 71,  
791 1921-1934.
- 792 Pfeifer, H.R., Häussermann, A., Lavanchy, J.C., Halter, W., 2007. Distribution and behavior  
793 of arsenic in soils and waters in the vicinity of the former gold-arsenic mine of Salanfe,  
794 Western Switzerland. *J. Geochem. Explor.* 93, 121-134.

- 795 Pratt, A.R., Nesbitt, H.W., Muir, I.J., 1994. Generation of Acids from Mine Waste - Oxidative  
796 Leaching of Pyrrhotite in Dilute H<sub>2</sub>SO<sub>4</sub> Solutions at pH 3.0. *Geochim. Cosmochim. Acta*  
797 58, 5147-5159.
- 798 Richardson, S., Vaughan, D.J., 1989. Arsenopyrite - a Spectroscopic Investigation of Altered  
799 Surfaces. *Mineral. Mag.* 53, 223-229.
- 800 Rosman, T.G., Uddin, A.N., Burns, F.J., 2004. Evidence that arsenic acts as a carcinogen in  
801 skin cancer. *Toxicol. Appl. Pharmacol.* 198, 394-404.
- 802 Routh, J., Bhattacharya, P., Jacks, G., Ahmed, K.M., Khan, A.A., Rahman, M.M., 2000.  
803 Arsenic geochemistry of Tala groundwater and sediments from Satkhira District,  
804 Bangladesh. *Eos. Trans. Am. Geophys. Union* 81, 550.
- 805 Ruitenberg, R., Hansford, G.S., Reuter, M.A., Breed, A.W., 1999. The ferric leaching kinetics  
806 of arsenopyrite. *Hydrometallurgy* 52, 37-53.
- 807 Salzsauler, K.A., Sidenko, N.V., Sherriff, B.L., 2005. Arsenic mobility in alteration products  
808 of sulphide-rich, arsenopyrite-bearing mine wastes, snow lake, manitoba, Canada. *Appl.*  
809 *Geochem.* 20, 2303-2314.
- 810 Sánchez-Rodas, D., Gomez-Ariza, J.L., Giraldez, I., Velasco, A., Morales, E., 2005. Arsenic  
811 speciation in river and estuarine waters from southwest Spain. *Sci. Total Environ.* 345,  
812 207-217.
- 813 Sarmiento, A.M., Casiot, C., Nieto, J.M., Elbaz-Poulichet, F., Olias, M., 2005. Seasonal  
814 variations in Fe and As speciation and mobility in waters affected by acid mine drainage in  
815 the Odiel river basin (Huelva, Spain). *Geogaceta* 37, 115-118.
- 816 Sarmiento, A.M., Oliveira, V., Gómez-Ariza, J.L., Nieto, J.M., Sánchez-Rodas, D., 2007. Diel  
817 cycles of arsenic speciation due to photooxidation in acid mine drainage from the Iberian  
818 pyrite belt (SW, Spain). *Chemosphere* 66, 677-683.
- 819 Schaufuss, A.G., Nesbitt, H.W., Scaini, M.J., Hoechst, H., Bancroft, M.G., Szargan, R., 2000.  
820 Reactivity of surface sites on fractured arsenopyrite (FeAsS) toward oxygen. *Am. Mineral.*  
821 85, 1754-1766.
- 822 Schwertmann, U., Murad, E., 1983. The effect of pH on the formation of goethite and  
823 hematite from ferrihydrite. *Clays Clay Miner.* 31, 277-284.
- 824 Shuey, R.T., 1975. Arsenopyrite FeAsS. In *Semiconducting Ore Minerals*. Elsevier, New  
825 York, USA.
- 826 Singer, P.C., Stumm, W., 1970. Acidic mine drainage: the rate determining step. *Science*, 167,  
827 1121-1123.
- 828 Smedley, P.L., Kinniburgh, D.G., 2002. A review of the source, behaviour and distribution of  
829 arsenic in natural waters. *Appl. Geochem.* 17, 517-568.
- 830 Smedley, P. L., Knudsen, J., Maiga, D., 2007. Arsenic in groundwater from mineralized  
831 Proterozoic basement rocks of Burkina Faso. *Appl. Geochem.* 22, 1074-1092.
- 832 Stumm, W. and Morgan, J.J (1996) *Aquatic chemistry. Chemical Equilibria and Rates in*  
833 *Natural Waters*. Wiley Interscience, New York.

- 834 Tallant, B.A., McKibben, M.A., 2005. Arsenic mineral kinetics: Arsenopyrite oxidation.  
835 Geochim. Cosmochim. Acta 69, A820-A820.
- 836 To, T., Nordstrom, K.D., Cunningham, K., Ball, J., and McCleskey, B., 1999. New method for  
837 the direct determination of dissolved Fe(III) concentration in acid mine waters. Environ.  
838 Sci. Technol. 33, 807-813.
- 839 Verplanck, P.L., Mueller, S.H., Goldfarb, R.J., Nordstrom, D.K., Youcha, K., 2007.  
840 Geochemical controls of elevated arsenic concentrations in groundwater, Ester Dome,  
841 Fairbanks district, Alaska. Chem. Geol. 255, 160-172.
- 842 Walker, F.P., Schreiber, M.E., Rimstidt, J.D., 2006. Kinetics of arsenopyrite oxidative  
843 dissolution by oxygen. Geochim. Cosmochim. Acta 70, 1668-1676.
- 844 Weisener, C., Smart, R., Gerson, A., 2003. Kinetics and mechanisms of the leaching of low  
845 Fe sphalerite. Geochim. Cosmochim. Acta 67, 823-830.
- 846 Weisener, C., Smart, R., Gerson, A., 2004. A comparison of the kinetics and mechanism of  
847 acid leaching of sphalerite containing low and high concentrations of iron. J. Miner.  
848 Process. 74, 239-249.
- 849 Welch, A.H., Stollenwerk, K.G.E., 2003. Arsenic in Ground Water: Geochemistry and  
850 Occurrence. Kluwer Academic Publishers.
- 851 Wen, C.Y., 1968. Noncatalytic solid-fluid reaction models. Ind. Eng. Chem. 60, 34-54.
- 852 White, A. F. and Peterson, M. L., 1990. Role of reactive-surface-area characterization in  
853 geochemical kinetic models. In: Melchior, D. C. and Bassett, R. L. Eds.) *ACS Symposium*  
854 *Series 416, Chemical Modeling of Aqueous Systems II*. American Chemical Society,  
855 Los Angeles, California.
- 856 White, A. & Brantley, S., 2003. The effect of time on the weathering of silicate minerals: why  
857 do weathering rates differ in the laboratory and field? Chem. Geol., 202, 479-506.
- 858 Williams, M., 2001. Arsenic in mine waters: an international study. Environ. Geol. 40, 267-  
859 278.
- 860 Williamson, M.A., Rimstidt, J.D., 1994. The Kinetics and Electrochemical Rate-Determining  
861 Step of Aqueous Pyrite Oxidation. Geochim. Cosmochim. Acta 58, 5443-5454.
- 862 Yu, Y., Zhu, Y., Williams-Jones, A.E., Gao, Z., Li, D., 2004. A kinetic study of the oxidation  
863 of arsenopyrite in acidic solutions: implications for the environment. Appl. Geochem. 19,  
864 435-444.
- 865 Yu, Y., Zhu, Y., Gao, Z., Gammons, C.H., Li, D., 2007. Rates of arsenopyrite oxidation by  
866 oxygen and Fe(III) at pH 1.8-12.6 and 15-45 °C. Environ. Sci. Technol. 41, 6460-6464.
- 867

**TITLES OF TABLES**

Table 1. Experimental conditions and arsenopyrite dissolution rates based on steady-state values at acidic pH.

Table 2. Experimental conditions and dissolution rates obtained by the SCM at neutral and basic pH. D is the diffusion coefficient for O<sub>2</sub> through the coating.

Table 3. Results obtained from X-ray Photoelectron Spectroscopy (XPS) determinations on the initial and reacted arsenopyrite samples. Surface stoichiometry is represented by molar ratios.

Table 4. Results obtained from micro-Raman spectra on the initial and reacted arsenopyrite samples (AsFeS means arsenopyrite).

Table 5. Arsenopyrite dissolution rate laws obtained in earlier studies and in the present work.

**FIGURE CAPTIONS**

Fig. 1. Variation in total iron (grey rhombi), arsenic (black rhombi) and sulphur (circles) as a function of time in arsenopyrite representative experiments at different pH and 8.7 mgL<sup>-1</sup> of input O<sub>2</sub>-dissolved and 25 °C. Initial concentrations for the experiments are not depicted for the sake of significance of the vertical scale at basic pH, and iron concentration is not depicted because it was below detection limit.

Fig. 2. SEM images of (a) freshly ground arsenopyrite with attached microparticles on the surfaces; (b) reacted arsenopyrite sample at pH 1 and 25°C, microparticles on the grains mostly dissolved; (c) reacted arsenopyrite sample at pH 5.5, precipitates upon grains of arsenopyrite; and (d) arsenopyrite grain with precipitates on surface after reacting at pH 7.5.

Fig. 3. Comparison of arsenopyrite oxidative dissolution rates versus pH obtained at 25°C and atmospheric conditions in the present study and in reported works (see also Table 5 for the experimental set up details). St-St means steady state conditions and SCM corresponds to the dissolution rates obtained using the Shrinking Core Model (SCM).

Fig. 4. Variation in output As, S and Fe concentration as a function of time and dissolved oxygen at 25°C and at pH 1 (a), at pH 3 (b). In (a) and (b) vertical lines delineate the different stages with different input DO concentration at each stage.

Fig. 5. Arsenopyrite dissolution rate dependence on dissolved oxygen at pH 1 and 25°C. Log arsenopyrite dissolution rate vs. log DO in the reactor calculated (a); dissolution rate vs. dissolved oxygen (b).

Fig. 6. Curve fitted S2p and As3d spectra of arsenopyrite representative samples dissolved at 25°C, 8.7 mg L<sup>-1</sup> input DO and pH 3 (a) and (b); at 1.8 mg L<sup>-1</sup> input DO and pH 3 (c) and (d); 8.7 mg L<sup>-1</sup> input DO and pH 5.6 (e) and (f) and ); 8.7 mg L<sup>-1</sup> input DO and pH 7.

Fig. 7. MicroRaman spectra of arsenopyrite samples reacted at  $8.7 \text{ mg L}^{-1}$  DO at  $25 \text{ }^\circ\text{C}$  and at pH 3 (a); and pH 7 (b).

Fig. 8. Evolution of the molar fraction of arsenopyrite dissolved,  $X$ , versus time in the experiment ASP-22 (see Table 2). The curves are the plot of the process controlled by the surface dissolution step ( $k$  control; eq. 2), by the diffusion across the coating step ( $D$  control; eq. 3), and by the SCM model (the addition eq. 2 and 3).

Fig. 9. Arrhenius plot of arsenopyrite dissolution rates obtained at acidic pH (a) and  $8.7 \text{ mg L}^{-1}$  input DO. The apparent activation energy is obtained in  $\text{kJ mol}^{-1}$ .

Figure 1a

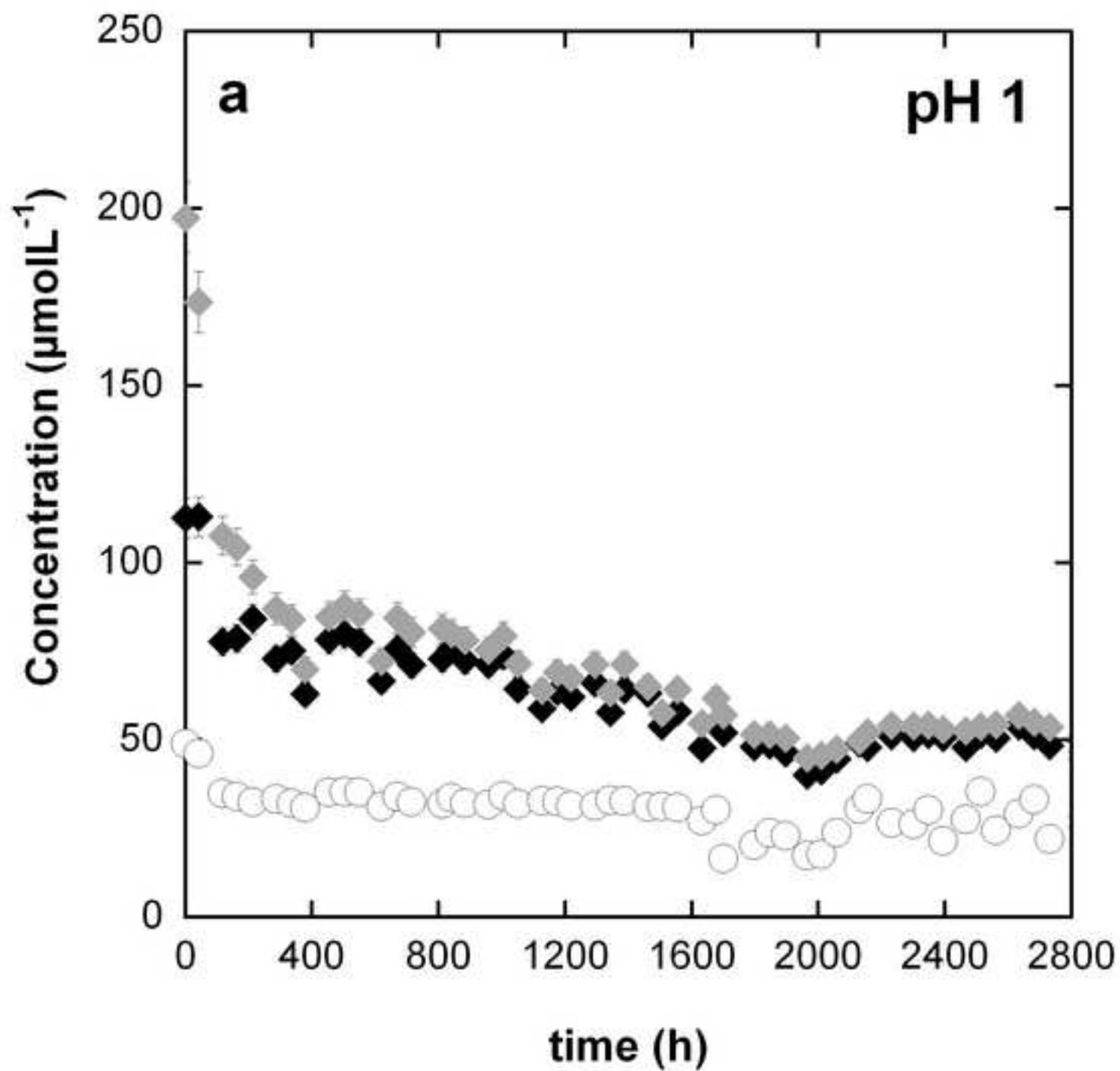


Figure 1b

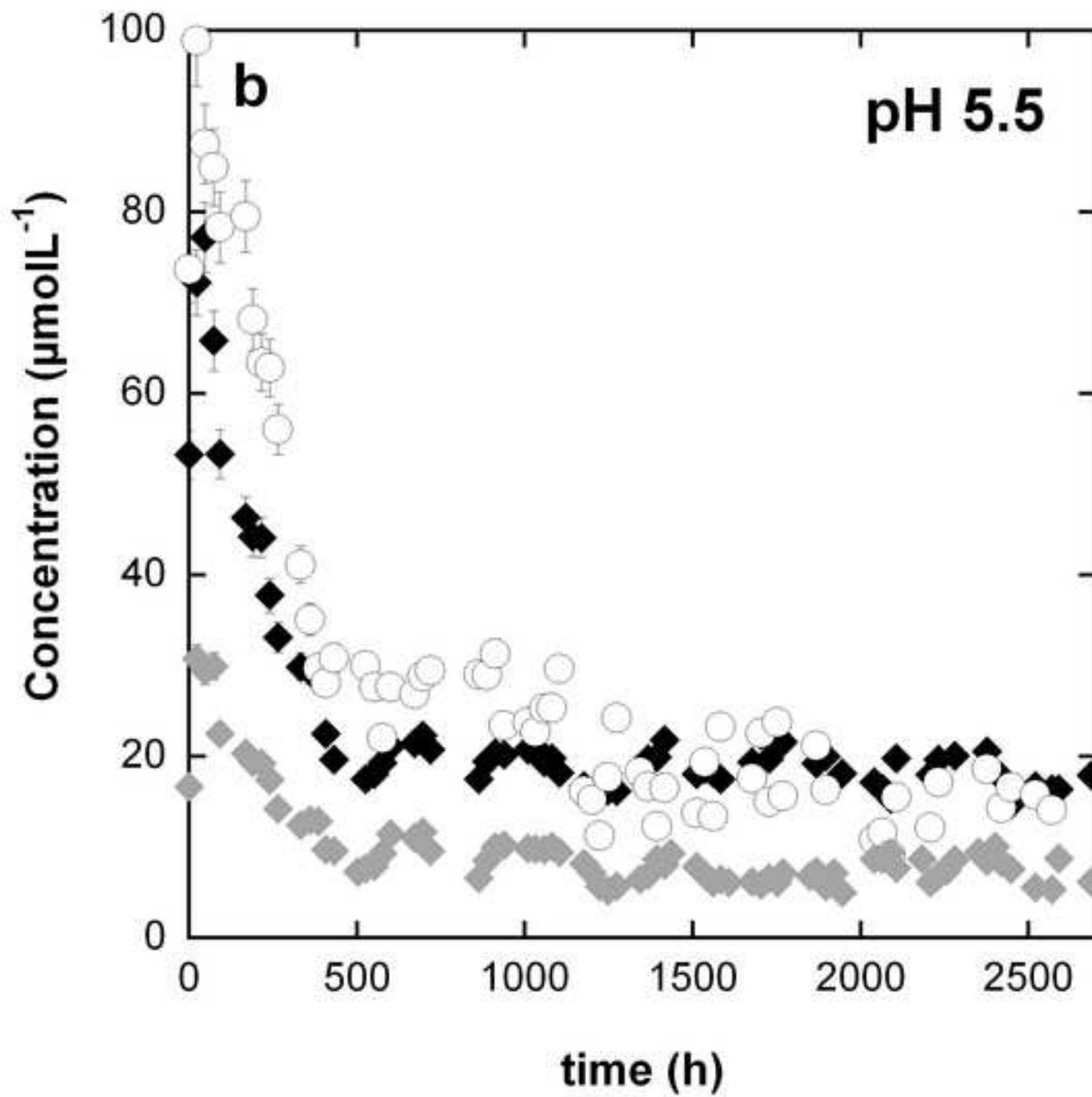


Figure 1c

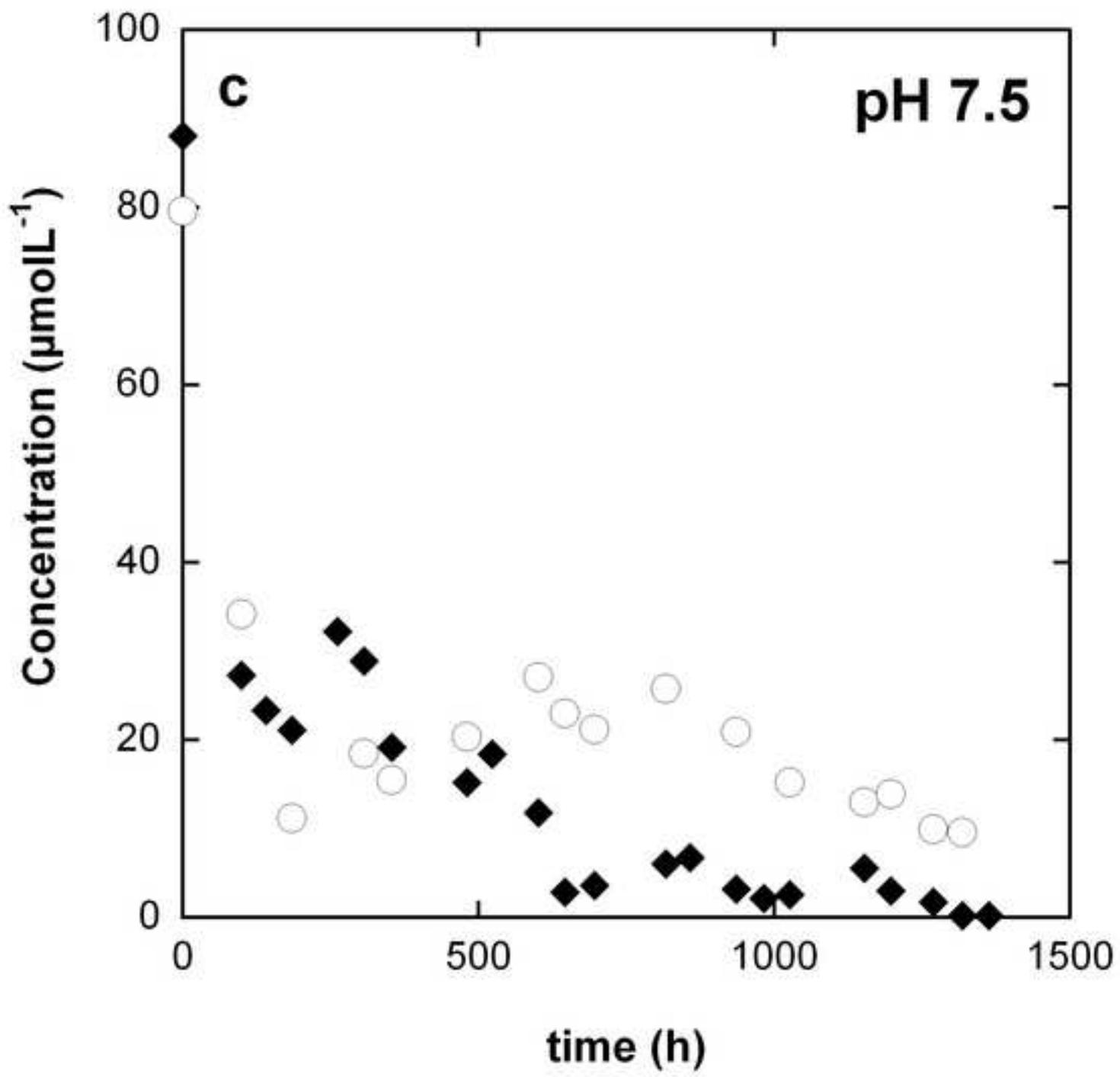


Figure 1d

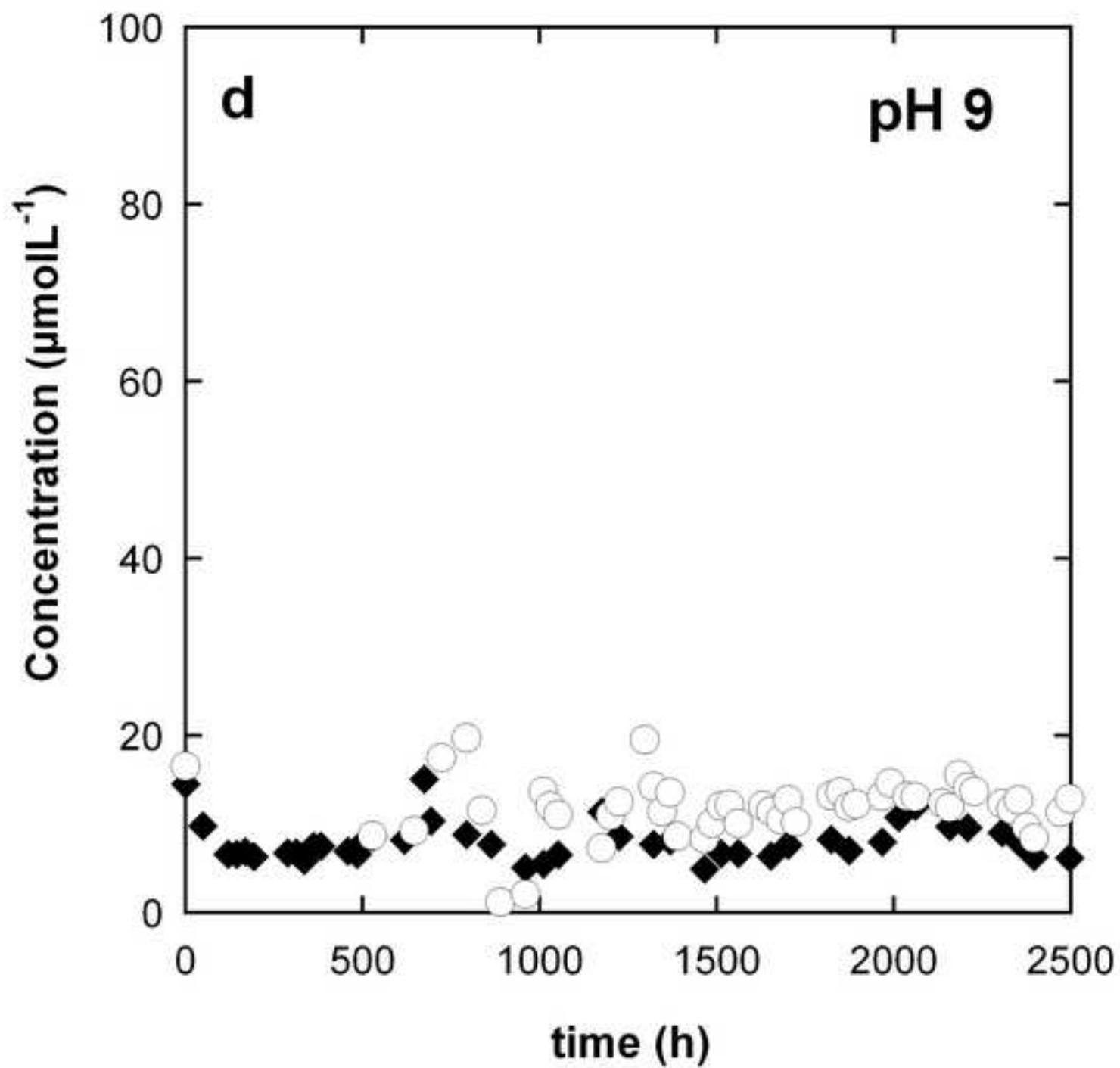


Figure 2

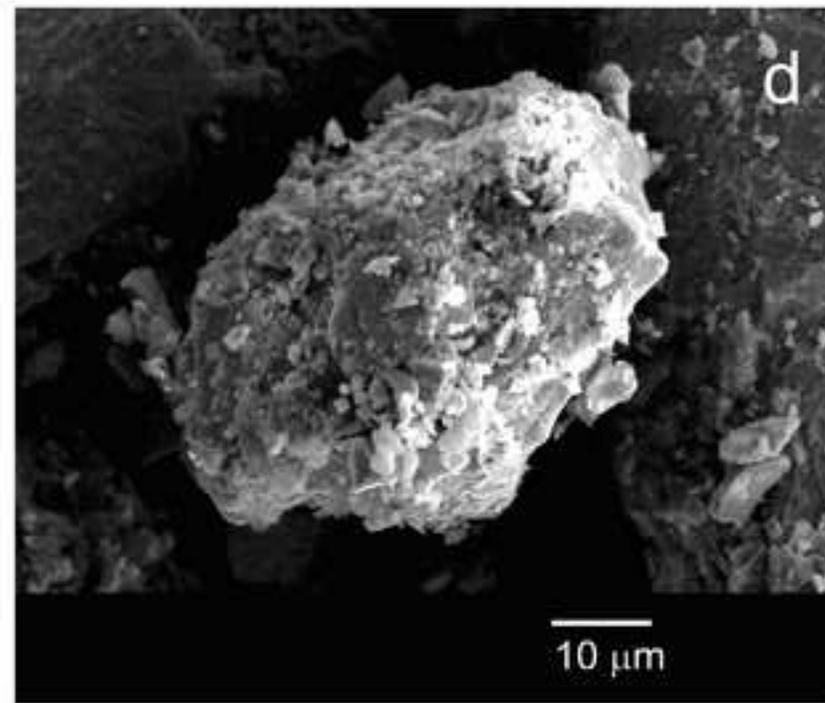
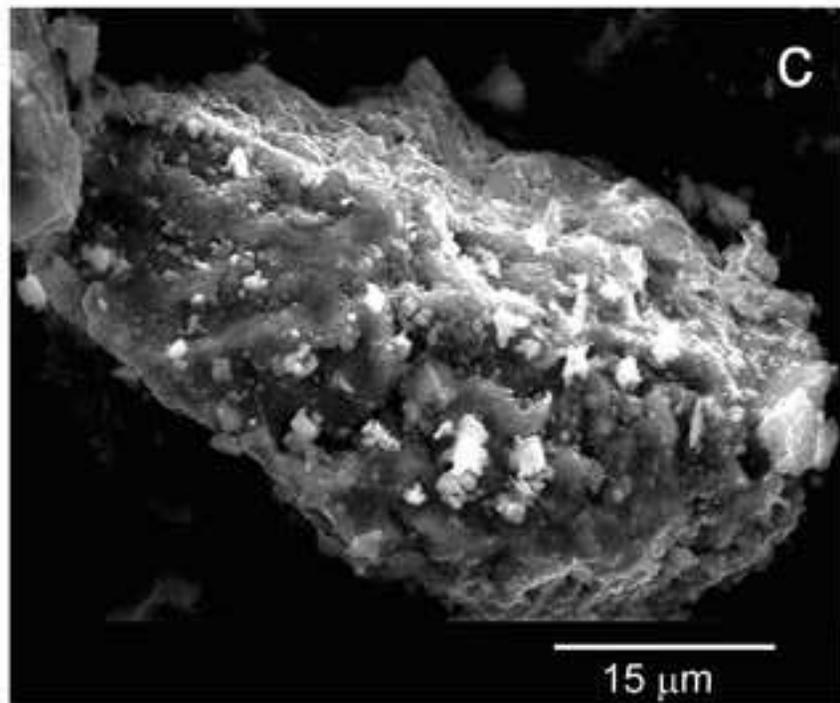
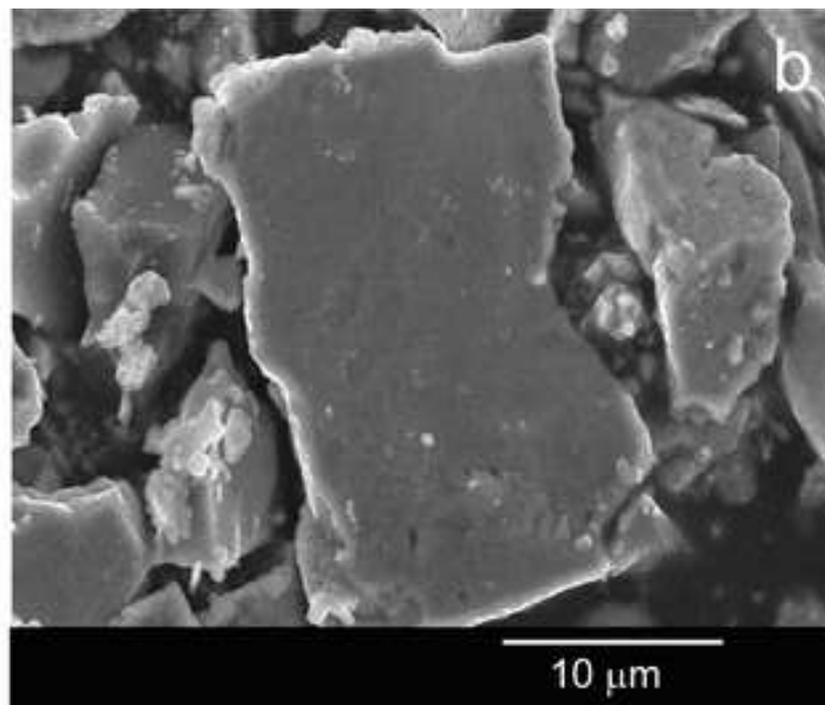
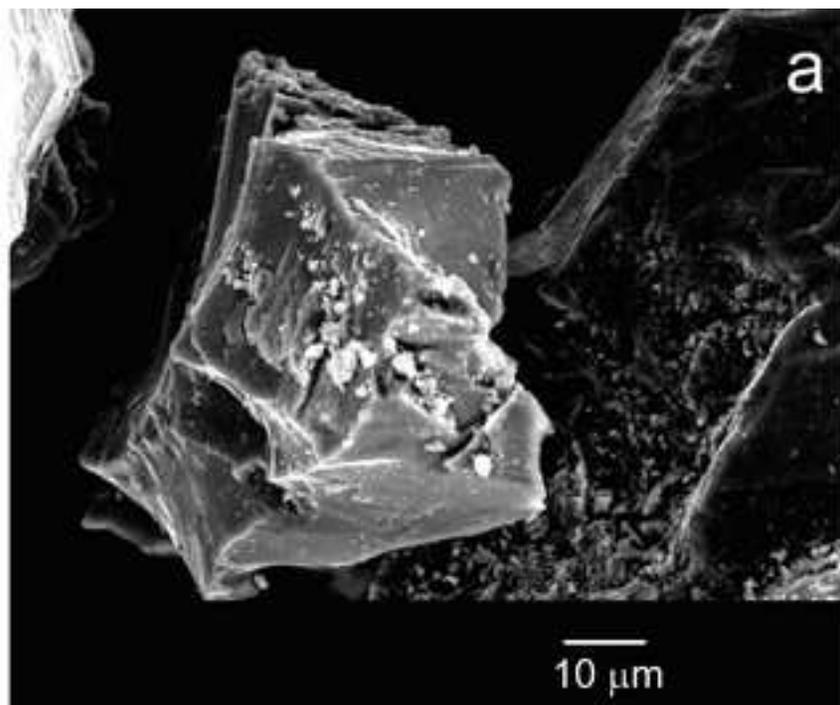


Figure 3

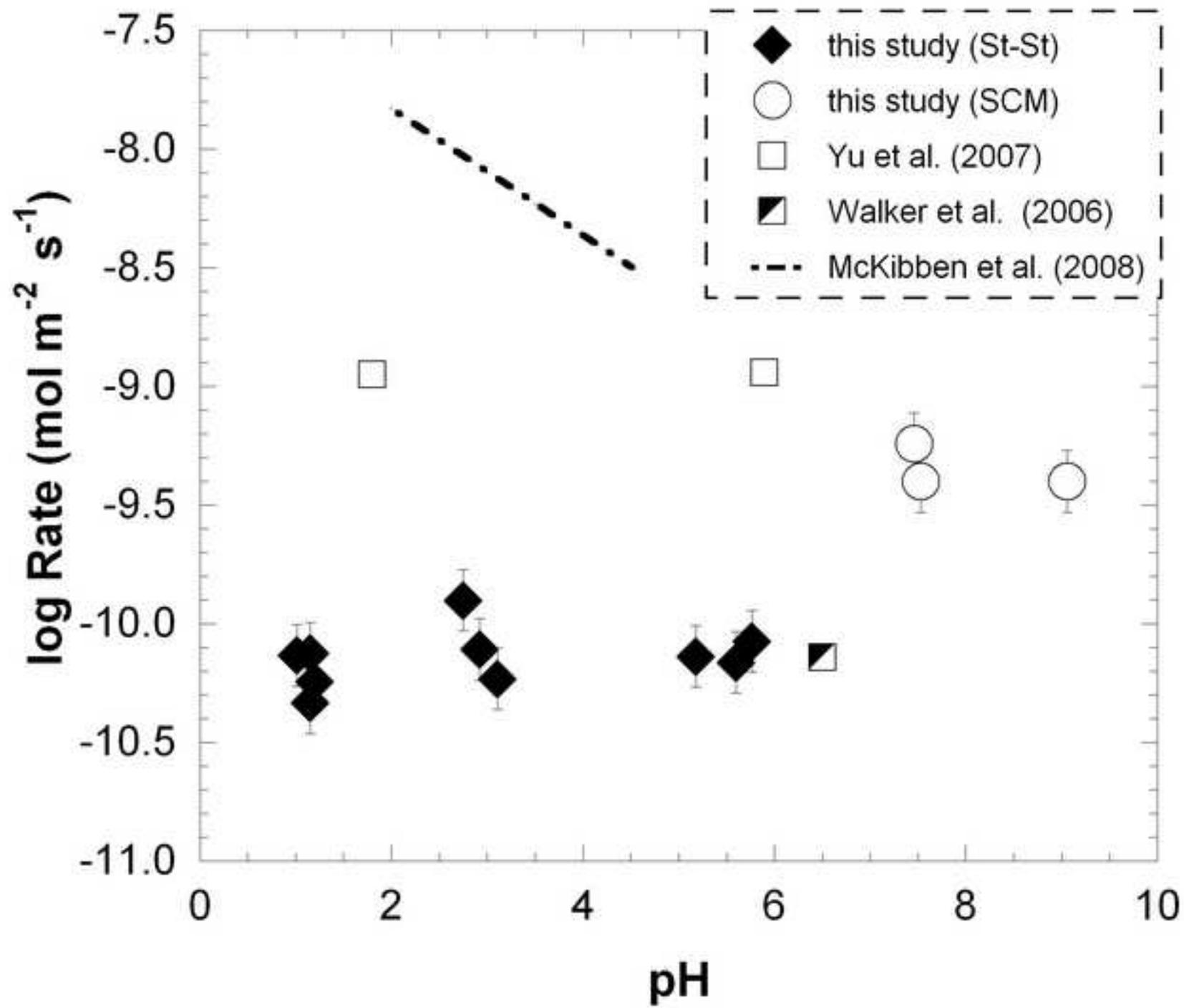


Figure 4a

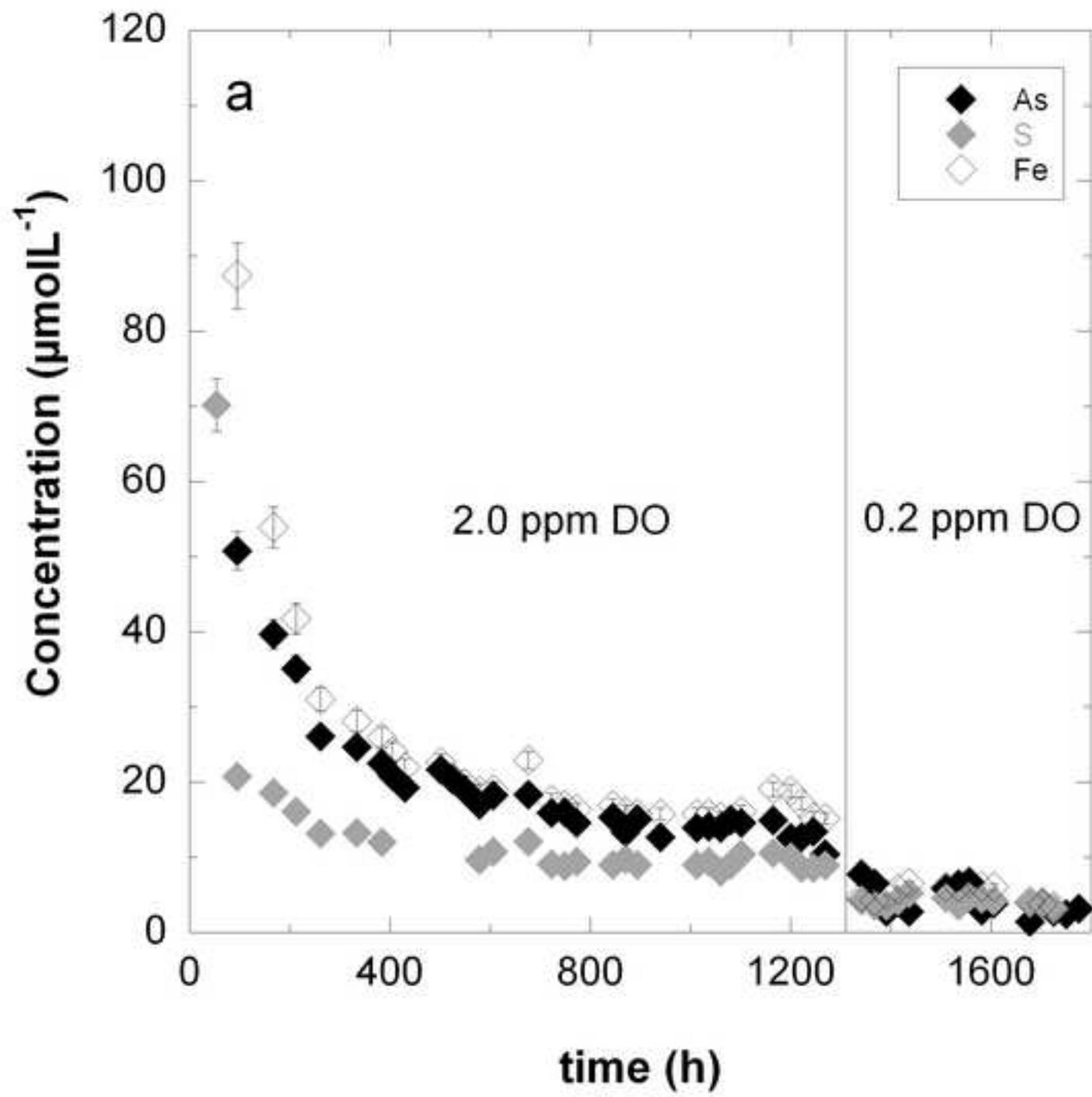


Figure 4b

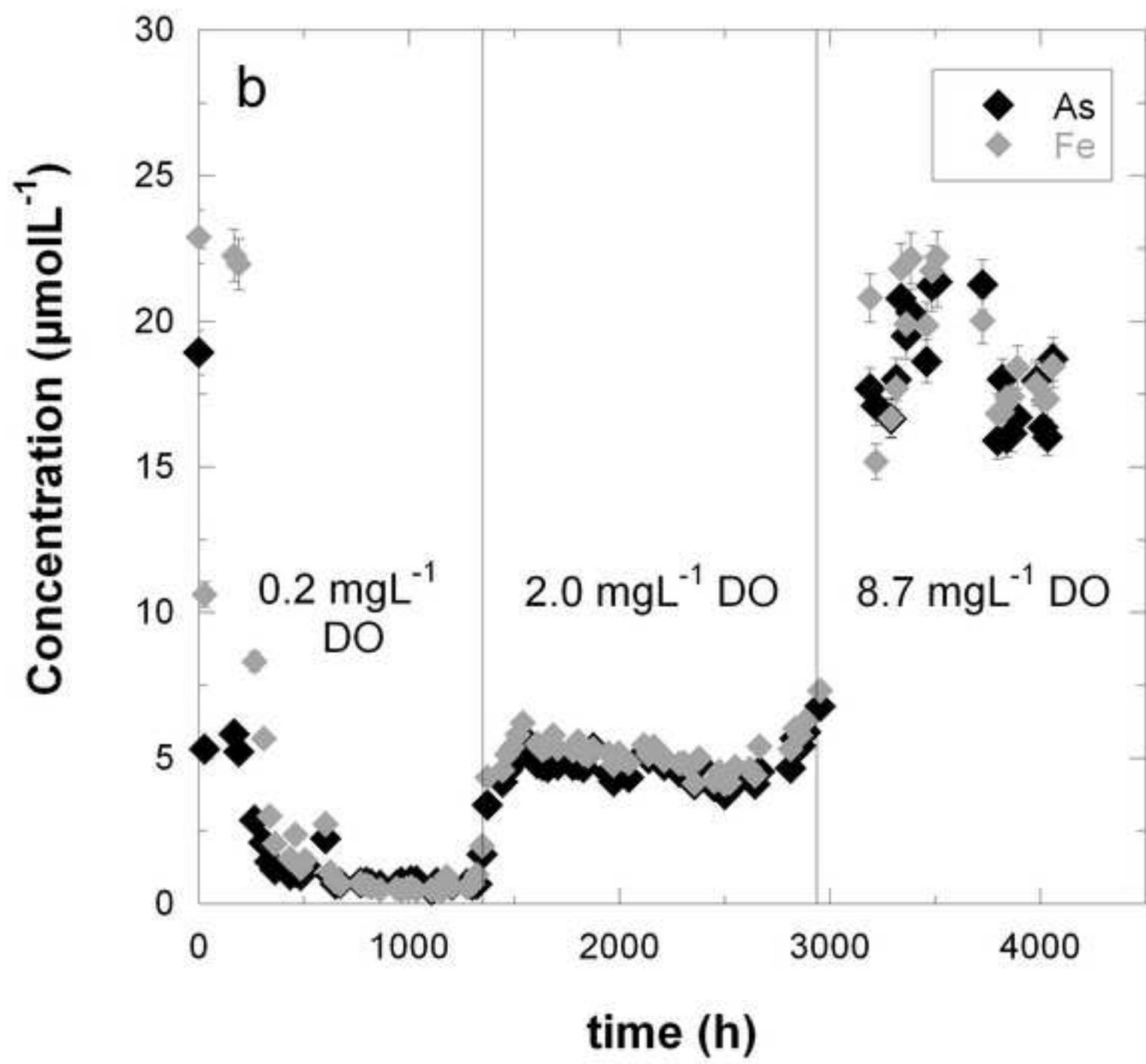


Figure 5a

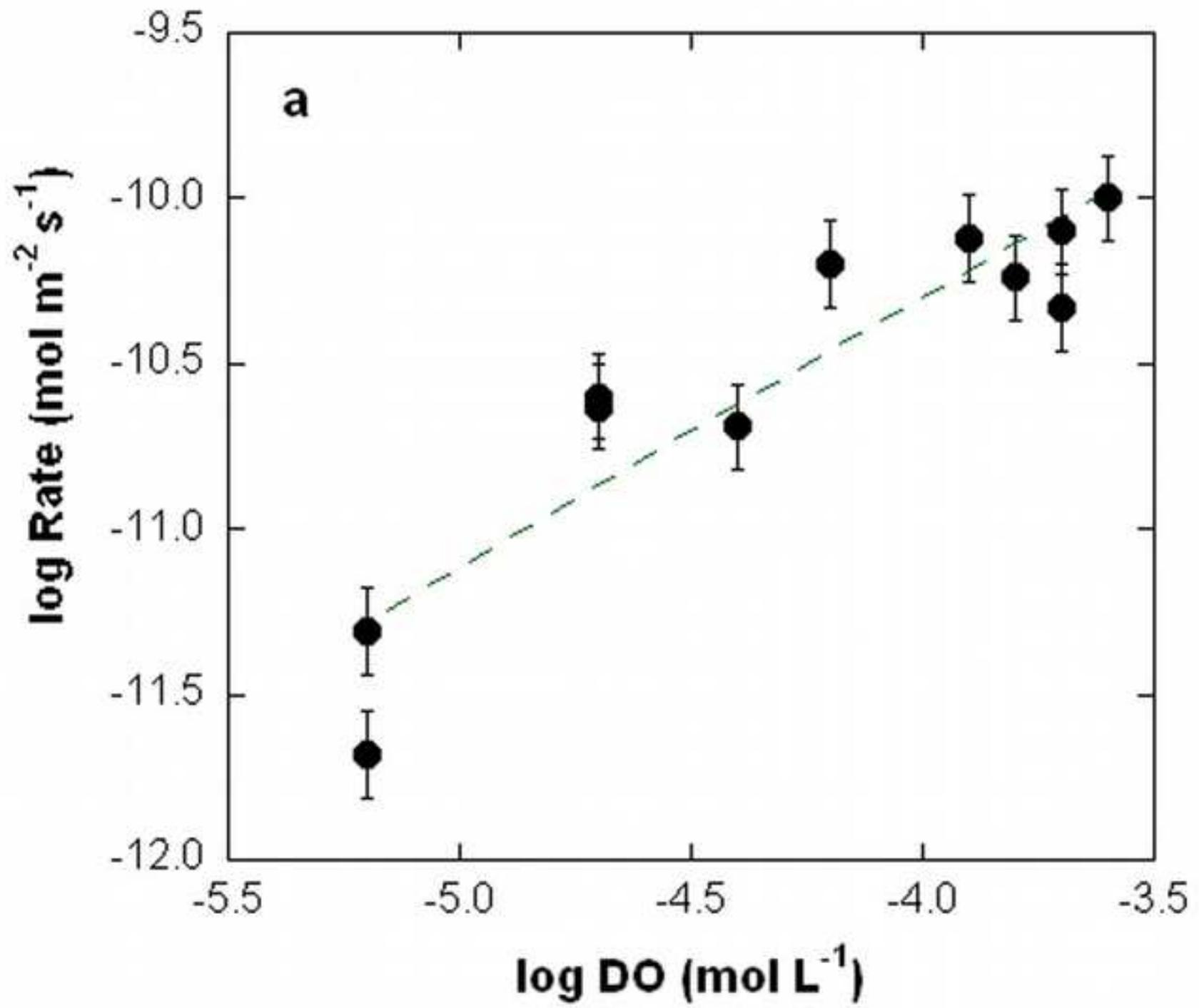


Figure 5b

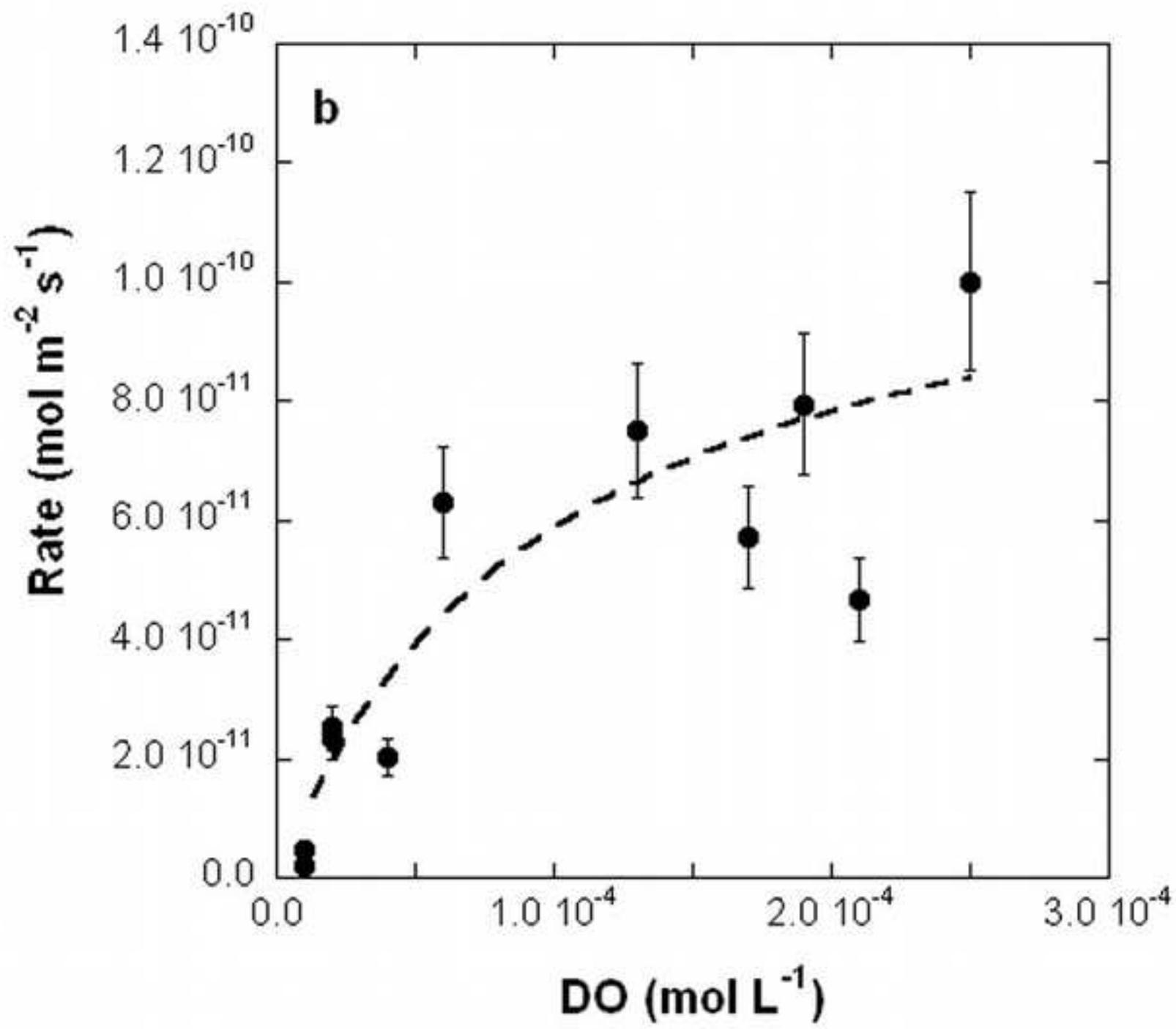


Figure 6

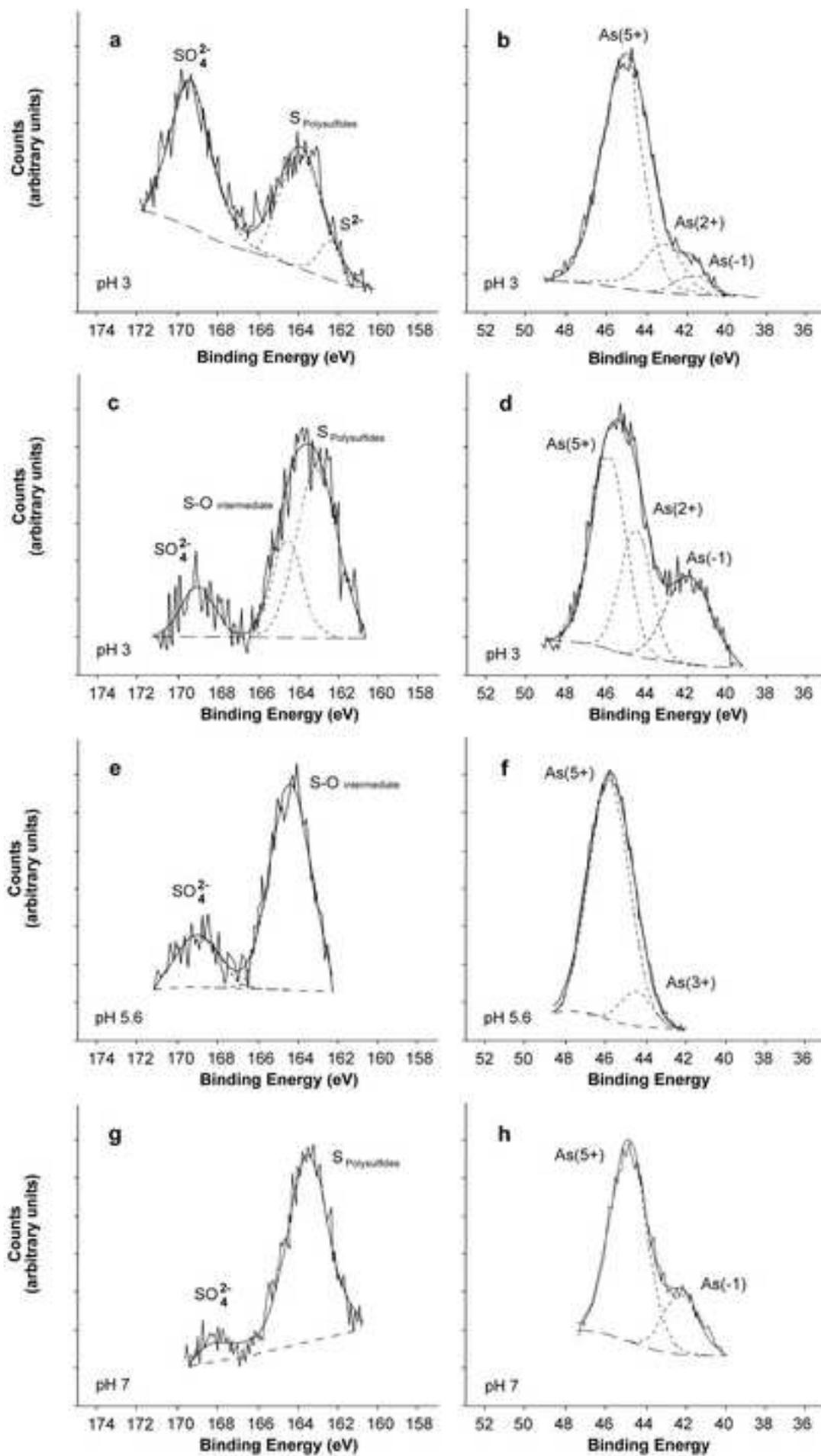


Figure 7

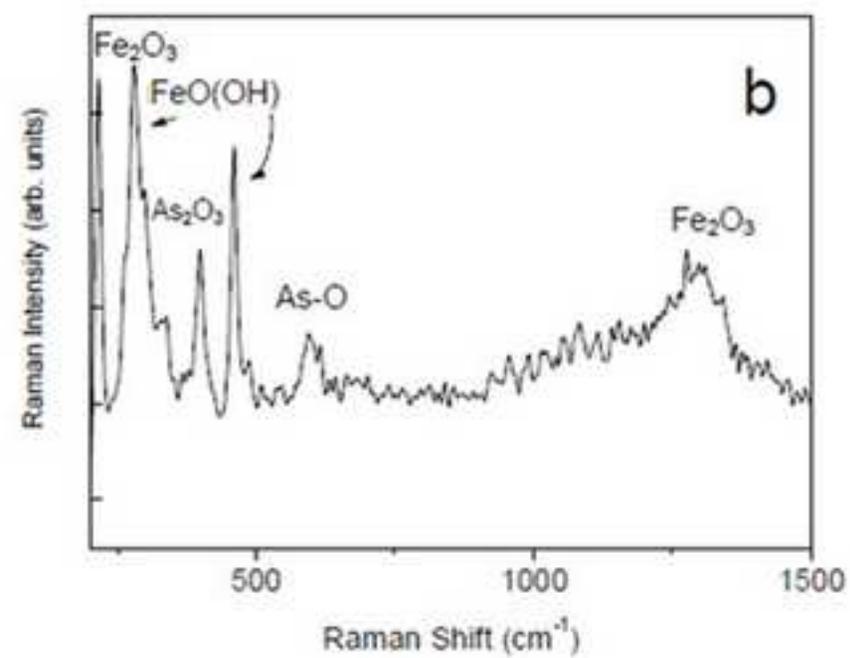
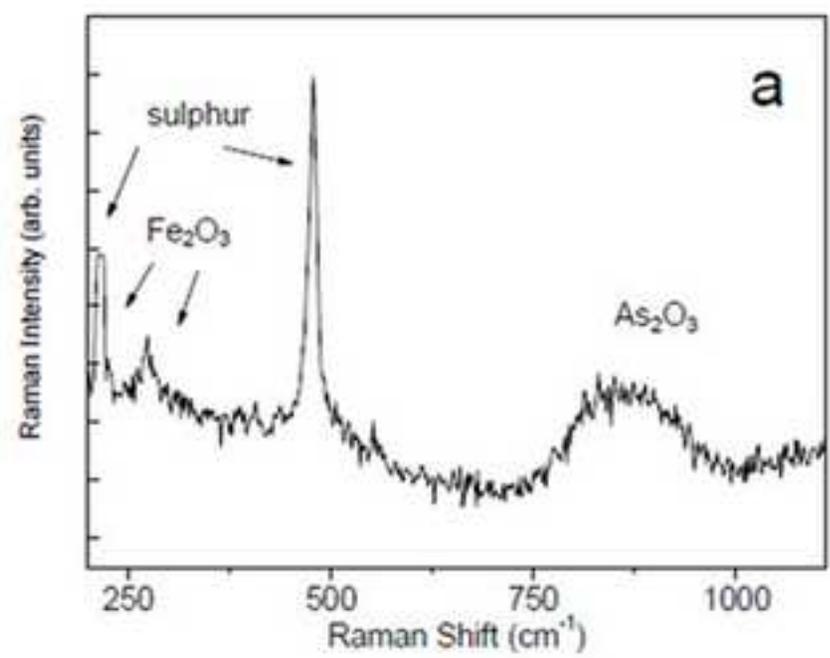


Figure 8

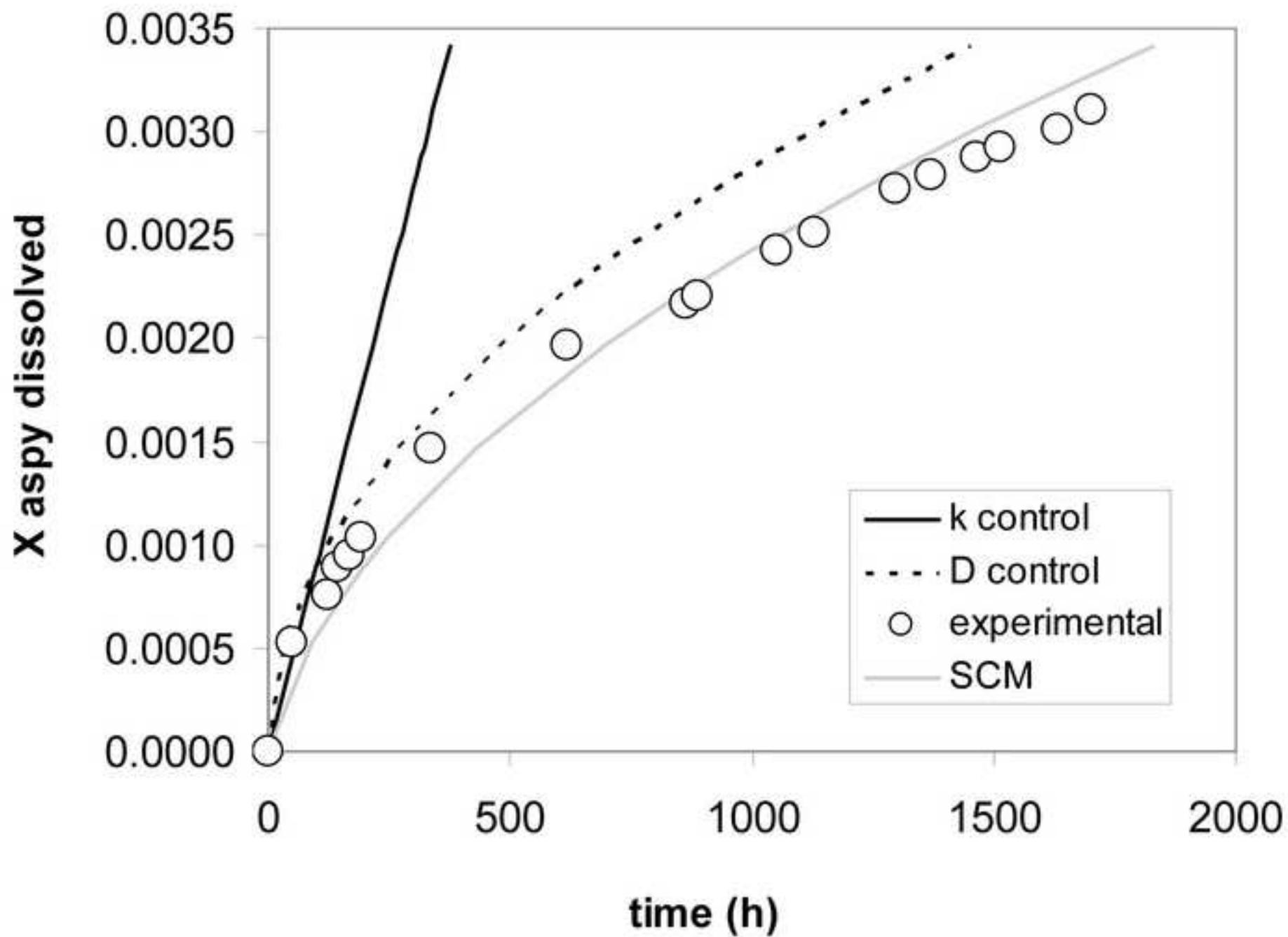


Figure 9

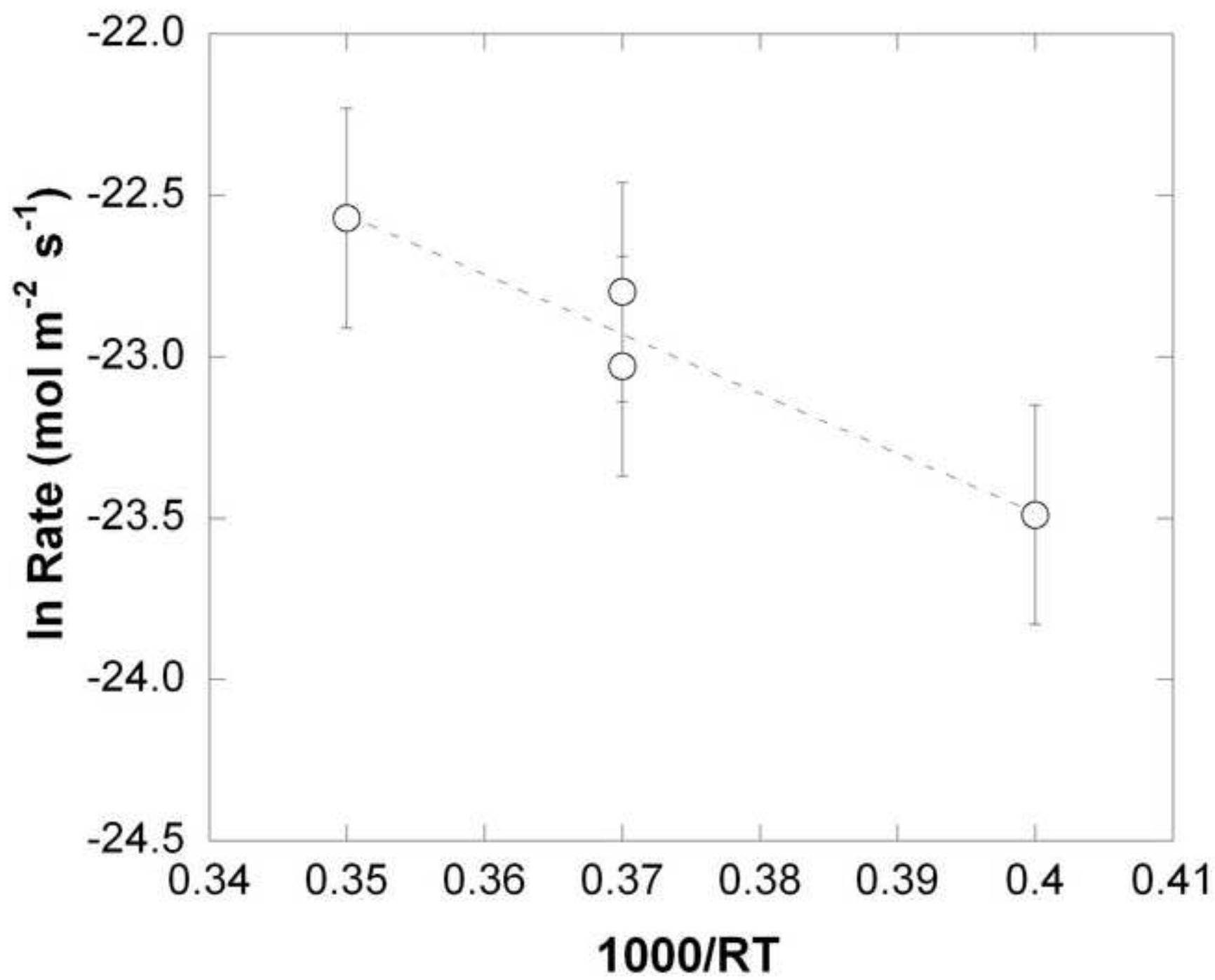


Table 1

Experiment	Stirring	Electrolyte	Flow rate (mL min <sup>-1</sup> )	pH	logDO		T (°C)	S	Fe (µM)	As	Stoichiometry			Final BET (m <sup>2</sup> g <sup>-1</sup> )	initial mass (g)	final mass (g)	R <sub>As</sub>	log R <sub>As</sub>	R <sub>Fe</sub>	log R <sub>Fe</sub>
					input (mol L <sup>-1</sup> )	calculated (mol L <sup>-1</sup> )					Fe/As	Si/As	Fe/S							
ASP-1	no	H <sub>2</sub> SO <sub>4</sub>	0.040	1.0	-3.6	3.7	25	99877	32.76	29.38	1.1	-	-	0.54	0.5014	0.4875	7.4·10 <sup>-11</sup>	-10.1	8.2·10 <sup>-11</sup>	-10.1
ASP-2	no	H <sub>2</sub> SO <sub>4</sub>	0.040	2.7	-3.6	3.8	25	963	41.34	34.66	1.1	-	-	0.40	0.5028	0.4962	1.3·10 <sup>-10</sup>	-9.9	1.4·10 <sup>-10</sup>	-9.9
ASP-3	no	HCl	0.037	1.2	-3.6	3.8	25	18.37	38.43	34.50	1.1	0.5	2.1	0.67	0.5039	0.4912	5.7·10 <sup>-11</sup>	-10.2	6.2·10 <sup>-11</sup>	-10.2
ASP-4	no	HCl	0.038	3.1	-3.6	3.7	25	18.37	30.96	29.67	1.0	0.6	1.7	0.65	0.5076	0.4930	5.9·10 <sup>-11</sup>	-10.2	6.1·10 <sup>-11</sup>	-10.2
ASP-5	yes	HCl	0.041	1.1	-3.6	3.9 <sup>(a)</sup>	25	27.76	53.97	50.33	1.1	0.5	1.9	0.63	0.8002	0.7330	7.5·10 <sup>-11</sup>	-10.1	8.1·10 <sup>-11</sup>	-10.1
ASP-7	no	HCl	0.034	1.3	-4.2	-4.4	25	4.34	9.16	7.89	1.2	0.6	1.2	0.43	0.5034	0.5013	2.0·10 <sup>-11</sup>	-10.7	2.4·10 <sup>-11</sup>	-10.6
ASP-8	no	HCl	0.036	3.1	-4.2	-4.8 <sup>(b)</sup>	25	22.69	22.51	17.42	1.3	1.3	1.0	0.60	0.7991	0.7911	2.2·10 <sup>-11</sup>	-10.7	2.8·10 <sup>-11</sup>	-10.6
ASP-9	no	HCl	0.038	1.1	-4.2	-4.7 <sup>(b)</sup>	25	22.93	18.89	15.64	1.2	1.0	0.8	0.49	0.8068	0.7992	2.5·10 <sup>-11</sup>	-10.6	3.0·10 <sup>-11</sup>	-10.5
ASP-10	no	Fe <sup>2+</sup> &H <sub>2</sub> SO <sub>4</sub>	0.038	1.2	-4.2	-3.7	25	124520	10070	21.49	-	-	-	0.61	0.4937	0.4829	4.7·10 <sup>-11</sup>	-10.3	-	-
ASP-11	no	HCl	0.014	1.1	-5.2	-5.4 <sup>(c)</sup>	25	9.81	7.83	4.38	1.8	1.7	0.8	0.33	0.8090	0.8075	2.1·10 <sup>-12</sup>	-11.7	3.7·10 <sup>-12</sup>	-11.4
ASP-12-a	yes	HCl	0.046	1.2	-4.2	-4.7 <sup>(b)</sup>	25	8.89	16.38	14.80	1.1	0.6	1.8	0.48	0.8239	0.8121	2.3·10 <sup>-11</sup>	-10.6	2.6·10 <sup>-11</sup>	-10.6
ASP-12-b	yes	HCl	0.048	1.3	-5.2	-5.4 <sup>(c)</sup>	25	3.89	4.38	3.03	1.1	1.3	1.1	0.48	0.8239	0.8085	4.9·10 <sup>-12</sup>	-11.3	7.0·10 <sup>-12</sup>	-11.2
ASP-50-1	no	HCl	0.033	1.1	-3.6	-3.9	50	29.12	54.27	51.84	1.0	0.6	1.9	0.46	0.5016	0.4742	1.3·10 <sup>-10</sup>	-9.9	1.3·10 <sup>-10</sup>	-9.9
ASP-50-2	no	HCl	0.035	2.9	-3.6	-3.9	50	48.38	71.67	78.72	1.0	0.6	1.5	0.51	0.8037	0.7763	1.2·10 <sup>-10</sup>	-9.9	1.1·10 <sup>-10</sup>	-10.0
ASP-70-1	no	HCl	0.037	1.0	-3.6	-3.7	70	28.01	59.06	49.19	1.2	0.9	2.1	0.41	0.5011	0.4785	1.4·10 <sup>-10</sup>	-9.8	1.7·10 <sup>-10</sup>	-9.8
ASP-13	no	HCl	0.038	5.2	-3.6	-3.8	25	36.90	25.93	33.72	0.8	1.1	0.7	0.63	0.5066	0.4702	7.3·10 <sup>-11</sup>	-10.1	5.6·10 <sup>-11</sup>	-10.3
ASP-14	no	DDW	0.036	5.5	-3.6	-3.7	25	15.45	4.06	18.03	0.2	0.8	0.3	0.67	0.5023	0.4851	3.2·10 <sup>-11</sup>	-10.5	7.3·10 <sup>-12</sup>	-11.1
ASP-15	yes	DDW	0.035	5.8	-3.6	-4.1	25	91.24	35.05	66.12	0.5	1.4	0.4	0.59	0.8039	0.7754	8.4·10 <sup>-11</sup>	-10.1	4.5·10 <sup>-11</sup>	-10.3
ASP-16-a	no	DDW	0.040	5.6	-5.2	-5.6	25	4.80	0.62	1.30	0.5	2.2	0.1	0.61	0.8033	0.8024	1.7·10 <sup>-12</sup>	-11.8	8.4·10 <sup>-13</sup>	-12.1
ASP-16-b	no	DDW	0.033	5.6	-3.6	-4.0	25	71.11	15.24	60.41	0.3	1.2	0.2	0.61	0.8033	0.7875	6.9·10 <sup>-11</sup>	-10.2	1.7·10 <sup>-11</sup>	-10.8
ASP-17	no	DDW	0.030	4.4	-3.6	-3.9	70	61.51	23.98	47.43	0.5	1.3	0.4	0.65	0.8002	0.7796	4.6·10 <sup>-11</sup>	-10.3	2.3·10 <sup>-11</sup>	-10.6
ASP-18	no	HCl	0.343	1.3	-3.6	-3.6	25	3.19	8.20	9.86	1.2	0.4	3.1	0.61	0.7664	0.7444	1.0·10 <sup>-10</sup>	-10.0	1.2·10 <sup>-10</sup>	-9.9
ASP-19	no	HCl	0.176	1.2	-3.6	-3.6	25	6.03	14.09	10.80	0.8	0.4	1.8	0.46	0.7065	0.6772	1.0·10 <sup>-10</sup>	-10.0	1.3·10 <sup>-11</sup>	-9.9
ASP-20	no	HCl	0.380	3.0	-4.2	-4.2	25	<i>b.d.l.</i>	9.37	8.30	0.9	-	-	0.60	0.8380	0.8092	1.1·10 <sup>-10</sup>	-10.0	1.2·10 <sup>-10</sup>	-9.9
ASP-21	no	HCl	0.403	1.0	-4.2	-4.2	25	<i>b.d.l.</i>	2.62	2.55	1.0	-	-	0.34	0.8171	0.8076	6.3·10 <sup>-11</sup>	-10.2	6.5·10 <sup>-11</sup>	-10.2

*b.d.l.*: below detection limit

*n.c.*: not calculated

(a) Log DO measured in the Steady State under the same experimental conditions =  $-3.8 \pm 0.1$

(b) Log DO measured in the Steady State under the same experimental conditions =  $-4.7 \pm 0.1$

(c) Log DO measured in the Steady State under the same experimental conditions =  $-5.5 \pm 0.1$

Table 2

Experiment	Stirring	Electrolyte	Flow rate (mL min <sup>-1</sup> )	pH	logDO		T (°C)	initial mass (g)	final mass	R <sub>s</sub> (mol m <sup>-2</sup> s <sup>-1</sup> )	log R <sub>s</sub>	D (m <sup>-2</sup> s <sup>-1</sup> )
					input (mol L <sup>-1</sup> )	calculated						
ASP-22	no	KH <sub>2</sub> PO <sub>4</sub>	0.034	7.5	-3.6	-3.6	25	0.5027	0.5005	4.0·10 <sup>-10</sup>	-9.4	1.0·10 <sup>-17</sup>
ASP-23	yes	KH <sub>2</sub> PO <sub>4</sub>	0.036	7.5	-3.6	-3.7	25	0.8040	0.7957	5.6·10 <sup>-10</sup>	-9.2	7.0·10 <sup>-16</sup>
ASP-24	no	Na <sub>2</sub> B <sub>4</sub> O <sub>7</sub>	0.031	9.1	-3.6	-3.7	25	0.5064	0.5014	4.0·10 <sup>-10</sup>	-9.4	1.5·10 <sup>-16</sup>
ASP-25	yes	Na <sub>2</sub> B <sub>4</sub> O <sub>7</sub>	0.032	8.9	-3.6	-3.6	25	0.8040	0.7925	Do not fit SCM		Do not fit SCM

Table 3

Sample	pH	S	Fe (at. %) <sup>†</sup>	As	Fe/S	As/S	Fe/As
<i>Initial FeAsS</i>		33.0	33.0	33.0	1	1	1
ASP-25-1*	1.0	35.9	26.2	36.7	0.73	1.02	0.71
ASP-25-3	1.2	37.5	27.5	34.2	0.73	0.91	0.80
ASP-25-2*	2.7	35.1	20.7	42.1	0.59	1.20	0.49
ASP-25-4	3.1	34.7	28.7	35.6	0.83	1.03	0.81
ASP-25-23	5.8	17.2	41.4	38.4	2.40	2.23	1.08
ASP-25-11	9.1	33.8	47.4	18.9	1.40	0.56	2.51

<sup>†</sup> Estimated normalizing out the rest of elements (oxygen and adventitious carbon)

\* Experiments carried out in H<sub>2</sub>SO<sub>4</sub> solutions

Table 4

Raw	pH 3	pH 7
AsFeS <sup>1,2</sup>	AsFeS <sup>1,2</sup>	AsFeS <sup>1,2</sup>
AsFeS <sup>1,2</sup>	As <sub>2</sub> O <sub>3</sub> <sup>3,4</sup>	As <sub>2</sub> O <sub>3</sub> + Fe <sub>2</sub> O <sub>3</sub> <sup>3,4,5</sup>
AsFeS <sup>1,2</sup>	AsFeS <sup>1,2</sup>	AsFeS <sup>1,2</sup>
AsFeS <sup>1,2</sup>	Fe <sub>2</sub> O <sub>3</sub> <sup>5</sup>	AsFeS <sup>1,2</sup>
SiO <sub>2</sub>	AsFeS+S <sup>1,2,5</sup>	As <sub>2</sub> O <sub>3</sub> <sup>3,4</sup>
AsFeS <sup>1,2</sup>	AsFeS + As <sub>2</sub> O <sub>3</sub> <sup>3,4</sup>	As <sub>2</sub> O <sub>3</sub> <sup>3,4</sup>
AsFeS <sup>1,2</sup>	AsFeS <sup>1,2</sup>	AsFeS <sup>1,2</sup>
AsFeS <sup>1,2</sup>	Fe <sub>2</sub> O <sub>3</sub> +AsO <sub>4</sub> +S <sup>3,4,5,6,7</sup>	AsFeS <sup>1,2</sup>

(1) McGuire et al. (2001b);

(2) Costa et al. (2002);(3) Martens et al. (2003a);

(4) Flynn et al. (1976); (5) Bersani et al. (1999);

(6) Mycroft et al. (1990);(7) Martens et al. (2003b).

Table 5

	Experiment type	pH	Temp (°C)	DO (mgL <sup>-1</sup> )	Activation Energy (kJ mol <sup>-1</sup> )	Dissolution Rate (mol m <sup>-2</sup> s <sup>-1</sup> )	Duration experiment (h)
Walker et al. (2006)	Mixed flow reactor	6.3-6.7	25	0.3-17	-	$r = 10^{-10.14 \pm 0.03}$	30
McKibben et al. (2008)	Batch reactor	2-5	10-40	-	-	$r = 10^{-6.11} [\text{O}_{2(\text{aq})}]^{0.33(\pm 0.18)} a_{\text{H}_2\text{O}}^{0.27(\pm 0.09)}$	-
Yu et al. (2007)	Mixed flow reactor	1.8-6.4	15-45	0.2-24.6	57	$r = 10^{-22.11 \pm 57/T} [\text{O}_{2(\text{aq})}]^{0.45(\pm 0.05)}$	6-8
<i>This study</i>	Flow through reactor	1-6	25-70	0-8.7	18.5	$r = 10^{-7.41(\pm 0.47)} [\text{O}_{2(\text{aq})}]^{0.76(\pm 0.11)} a_{\text{H}_2\text{O}}^{-0.12(\pm 0.07)}$	600-4500