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2       **NOVEL CATION SELECTIVE EXHAUSTIVE INJECTION-SWEEPING**  
3       **PROCEDURE FOR 5-NITROIMIDAZOLE DETERMINATION IN WATERS**  
4       **BY MICELLAR ELECTROKINETIC CHROMATOGRAPHY USING**  
5       **DISPERSIVE LIQUID-LIQUID MICROEXTRACTION**

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18      **Keywords:** 5-Nitroimidazoles, Dispersive liquid-liquid microextraction, Cation selective  
19      exhaustive injection, Sweeping, Micellar electrokinetic chromatography

21 **Abstract**

22 A novel method consisting on cation-selective exhaustive injection and sweeping (CSEI-  
23 sweeping) as on-line preconcentration followed by a micellar electrokinetic chromatography  
24 (MEKC) separation has been developed for the determination of 5-nitroimidazoles (5-NDZ) in  
25 environmental waters. Moreover, dispersive liquid-liquid microextraction (DLLME) has been  
26 proposed for first time as sample treatment technique prior to CSEI-sweeping-MEKC. DLLME  
27 was applied on 5 mL of sample. Dibromomethane (1156  $\mu$ L) and 2-butanol (1363  $\mu$ L) were  
28 employed as extractant and dispersive solvents, respectively. Salting-out effect was achieved by  
29 the addition of 16 % (w/v) NaCl to the samples. After DLLME and organic solvent evaporation,  
30 the residue was redissolved in a low conductivity solvent (5 mM phosphoric acid with 5% of  
31 methanol) and electrokinetically injected at 9.8 kV for 632 s in a bare fused-silica capillary (57.2  
32 cm, 50  $\mu$ m I.D.). Prior to the injection, the capillary was rinsed with 50 mM phosphate buffer pH  
33 2.5, followed by a plug of a higher conductivity buffer (100 mM phosphate pH 2.5, 50 mbar, 264  
34 s) and a plug of water (50 mbar, 2 s). Separation was carried out applying -30 kV at 20 °C in 44  
35 mM phosphate buffer pH 2.5, containing 8% tetrahydrofuran and 123 mM sodium dodecyl sulfate.  
36 Analytical signals were monitored at 276 nm. Validation was performed in river and well waters,  
37 obtaining satisfactory results in terms of linearity, precision ( % RSD generally lower than 10%)  
38 and trueness (recoveries higher than 70% in almost all cases). LODs ranged from 0.61 to 2.44  
39 ng/mL. The combination of this microextraction technique with the proposed capillary  
40 electrophoresis methodology supposes a simple, sensitive and cheap alternative for 5-NDZ  
41 analyses, in accordance with the aims of green chemistry.

42

43     **1. Introduction**

44         Metronidazole (MNZ) is an antibiotic which belongs to 5-nitroimidazole (5-NDZ) family. It  
45         is widely used in humans for treating diseases due to anaerobia microbes [1]. It is considered an  
46         essential antibiotic according to World Health Organization (WHO) [2]. Other drugs from the same  
47         family, such as tinidazole (TNZ), ornidazole (ORZ), carnidazole (CRZ), ipronidazole (IPZ),  
48         secnidazole (SCZ) and ternidazole (TRZ) have been also considered in the current work. The  
49         chemical structures of the selected compounds are shown in Figure 1. In spite of the  
50         effectiveness of 5-NDZs as antibacterial and antiprotozoal agents, their employment in the  
51         veterinary practice is restricted. Some reports attribute mutagenic, carcinogenic and genotoxic  
52         properties to 5-NDZs [3-4], and therefore their application have been banned in animals intended  
53         for human consumption in European Union (EU), United States (US) and China [5-7]. On the  
54         other hand, 5-NDZs possess high polarity and low biodegradability, which involves high  
55         bioaccumulation levels and, consequently, ecotoxicity [8]. The presence of antibiotics such as 5-  
56         NDZs in sewage water has been already reported [9], even though a few number of papers have  
57         been focused on their removal from wastewater treatment plants [10,11]. For all these reasons, 5-  
58         NDZ drugs are considered as emerging water micropollutants. The environmental risks caused  
59         by drugs have been studied in the past decade [12], however more research in this area is still  
60         required, including updated reports about their presence and levels in ecosystems. Thus,  
61         analytical methodology for detection and quantification of these residues in the environment is  
62         highly needed in order to evaluate the exposition of the environment to antibiotics and their risks  
63         [13].

64         Several methods have been proposed for 5-NDZ determination [14-17], however,  
65         applications to environmental water samples are reduced [18-20]. New contributions about 5-  
66         NDZ determination in aquatic environmental samples are desired, taking into account that their  
67         presence has been already reported in natural waters [21]. Traditionally, liquid chromatography  
68         (LC) coupled to mass spectrometry (MS) has been the most popular choice for monitoring 5-NDZ  
69         residues [22-23]. Capillary electrophoresis (CE) [24] or gas chromatography (GC) [25] are  
70         among the proposed alternatives to LC. Low solvent consumption, short analysis time and high  
71         efficiency are CE characteristics; however its use is limited due mainly to the poor sensitivity,  
72         especially when it is coupled to ultraviolet (UV) detection. In order to overcome this  
73         disadvantage, different preconcentration strategies have been developed, including on-line  
74         (sample stacking techniques) [26], in-line [27] and off-line (through sample pretreatment)  
75         procedures[28]. The most common on-line preconcentration methodologies include: acetonitrile

76 stacking [29], field-amplified sample stacking (FASS) and field-amplified sample injection (FASI)  
77 [30], sweeping [31], dynamic pH junction [32] and isotachophoretic stacking [33], although others  
78 less known strategies offer even higher sensitivity enhancement factors (SEF).

79 A decade ago, Quirino et al. proposed a novel on-line preconcentracion technique, based  
80 on cation-selective exhaustive injection and sweeping (CSEI-sweeping). It combines two on-line  
81 preconcentration techniques: field-enhanced sample injection (FESI) and sweeping. This  
82 combination allows achieving enhancement factors from a thousand- to almost a million-fold in  
83 relation to conventional CE [34]. CSEI-sweeping involves electrokinetic injection (FESI) of a high  
84 amount of charged cationic compounds, creating long analyte zones in the capillary with higher  
85 concentration than in the original sample solution [35]. After sample injection, background  
86 electrolyte (BGE) vials containing micelles are placed at both ends of the capillary and negative  
87 voltage is applied. Micelles focused the analytes in narrow bands by sweeping, and they are  
88 consequently separated by conventional micellar electrokinetic chromatography (MEKC). A  
89 procedure scheme is shown in Figure 2. To achieve FESI, analytes must be dissolved in a low-  
90 conductivity medium, which is not so obvious, especially for complex real samples showing high  
91 or moderate salinity. In such cases, a proper sample pretreatment is needed in order to reduce  
92 sample conductivity. Another drawback inherent to CSEI-sweeping-MEKC is the run time since it  
93 requires longer analysis times compared to conventional CE, considering capillary conditioning,  
94 injection and separation time [34]. However, most commercial CE instruments allow the  
95 automation of capillary preconditioning, sample injection and CE separation. Currently CSEI-  
96 sweeping-MEKC is a quite novel technique since it has not been much exploited yet, however it  
97 has been successfully employed e.g. for thedetermination of drugs of abuse in urine [36-37];  
98 methadone in serum [38]; herbicides in water samples [39]; and melamine and cyromazine in milk  
99 [40].

100 Solid phase extraction (SPE) has demonstrated to be a good option for sample clean-up  
101 prior to CSEI-sweeping-MEKC [41-42]. Other techniques have been also employed, such as  
102 cloud point extraction [43] or hollow fiber based liquid-liquid-liquid microextraction (HF-LLME)  
103 [44]. Dispersive liquid-liquid microextraction (DLLME) is a novel microextraction method  
104 introduced by Rezaee et al. [45]. Due to its simplicity, low cost and low solvent consumption, it  
105 has been very popular as sample pretreatment during last decade. DLLME consists on a quick  
106 injection of an organic solvent mixture (water-immiscible organic extractive solvent plus water-  
107 miscible organic dispersive solvent) into an aqueous sample causing a cloudy dispersion.  
108 Dispersive solvent enhances the exchange surface between the extractive solvent and the

109 aqueous sample matrix, assisting the analyte extraction. It is a miniaturized sample treatment  
110 technique which has found its major application field in water analyses with satisfactory results  
111 [46]. To the best of our knowledge, DLLME has never been employed for sample treatment prior  
112 to a CSEI-sweeping-MEKC procedure.

113 In this work, a novel CSEI-sweeping-MEKC method has been developed for 5-NDZ  
114 determination in water samples. In addition, DLLME has been successfully coupled to CSEI-  
115 sweeping-MEKC, since DLLME yields low conductivity extracts compatible with the requirements  
116 of FESI. The combination of DLLME as sample treatment and CE is a green alternative for 5-  
117 NDZ analyses, considering the low consumption of reagents and sample and the low impact of  
118 the buffers employed as separation media. Also DLLME represents a quick procedure which  
119 compensates the time needed for the on-line preconcentration and separation method.

120

## 121 **2. Materials and Methods**

122

### 123 **2.1 Chemicals and reagents**

124 All reagents were analytical reagent grade, unless indicated otherwise, and solvents were  
125 HPLC grade. Ultrapure water (Milli-Q plus system, Millipore, Bedford, MA, US) was used  
126 throughout the work. Sodium chloride (NaCl), sodium hydroxide (NaOH), sodium dihydrogen  
127 phosphate (NaH<sub>2</sub>PO<sub>4</sub>) and orthophosphoric acid (H<sub>3</sub>PO<sub>4</sub>) (85%) were obtained from Panreac-  
128 Química (Madrid, Spain); methanol (MeOH) and 2-butanol were supplied by VWR International  
129 (West Chester, PA, US). Acetonitrile (MeCN), dibromomethane, isopropanol, sodium dodecyl  
130 sulfate (SDS) and oxalic acid were purchased from Sigma-Aldrich (St. Louis, MO, USA). Formic  
131 acid (98-100%) and tetrahydrofuran (THF) were supplied by Merck (Darmstadt, Germany).

132 Analytical standards of CRZ ([2-(2-methyl-5-nitro-imidazol-1-yl)ethyl]thiocarbamic acid o-  
133 methyl ester), ORZ (1-(3-chloro-2-hydroxypropyl)-2-methyl-5-nitroimidazole), MNZ (1-(2-  
134 hydroxyethyl)-2-methyl-5-nitroimidazole) and TNZ (1-(2-ethylsulfonylethyl)-2-methyl-5-nitro-  
135 imidazole) were supplied by Sigma-Aldrich (St. Louis, MO, US), while IPZ (2-isopropyl-1-methyl-  
136 5-nitroimidazole), SCZ ( $\alpha$ ,2-dimethyl-5-nitro-1H-imidazole-1-ethanol hemihydrate) and TRZ (1-(3-  
137 hydroxypropyl)-2-methyl-5-nitroimidazole) hydrochloride were purchased from Witega (Berlin,  
138 Germany).

139 0.2  $\mu$ m nylon membrane filters (Pall Corp, MI, US) were used for sample filtration.

140

### 141 **2.2 Standard solutions preparation**

142 Individual stock standard solutions of every 5-NDZ containing 1.00 g/L were prepared by  
143 dissolving each pure compound in MeCN. These solutions were stored in dark bottles at -20°C  
144 and equilibrated to room temperature before use. They were stable for at least six months.  
145 Intermediate standard solutions of 2.00 mg/L of each 5-NDZ, except for TNZ (4.01 mg/L), were  
146 prepared by mixing aliquots of each individual stock standard solution and diluting with MeCN.  
147 These solutions were stored at 4°C and exposure to direct light was avoided. They were stable  
148 for at least three months.

149 Fresh working standard solutions of lower concentrations were daily prepared in injection  
150 medium.

151

### 152 **2.3 Instrumentation and software**

153 CE experiments were carried out with an Agilent 7100 CE System (Agilent Technologies,  
154 Waldbronn, Germany) equipped with a diode-array detector. Data were collected using the  
155 software supplied with the HP ChemStation (Version B.02.01). Separations were performed in a  
156 57.2 cm x 50 µm internal diameter (i.d.), uncoated fused-silica capillary with an optical path  
157 length of 150 µm (bubble cell capillary from Agilent Technologies) and an effective length of 48.7  
158 cm.

159 A pH-meter (Crison model pH 2000, Barcelona, Spain) with a resolution of ±0.01 pH unit,  
160 a centrifuge (Universal 320 model from Hettich, Leipzig, Germany), an evaporator with nitrogen  
161 (System EVA-EC from VLM GmbH, Bielefeld, Germany) and a vortex (Genie 2 model from  
162 Scientific Industries, Bohemia, USA) were also used.

163 UNSCRAMBLER® v 9.8 software [47] was used for evaluating the obtained results.

164

### 165 **2.4 Preparation of samples**

166 Natural water samples from different sources were considered in this study. River water  
167 samples were collected from Riofrío river (Riofrío, Granada, Spain), being the sampling point  
168 located after a fish farm drain, and from Genil urban river (Granada, Spain). Besides, water  
169 samples from a well placed in a cattle area (La Serena, Badajoz, Spain) were also studied. Water  
170 samples were kept at 4°C and equilibrated to room temperature before analysis.

171 The employed DLLME procedure for water sample treatment has been previously  
172 reported by our group [18]. The sample treatment was applied to 5.00 mL aliquots of each water  
173 sample contained in a 15 mL centrifuge tube with a conical bottom. Analyte extraction was  
174 assisted by a salting out effect, adding 0.800 g of NaCl (16% w/v) to each sample. Salt was

175 dissolved in water samples by vortexing before carrying out the extraction. 5-NDZ extraction took  
176 place through the quick injection of an organic solvent mixture into the water sample consisting  
177 on 1156  $\mu$ L of dibromomethane as extraction solvent and 1363  $\mu$ L of 2-butanol as dispersive  
178 solvent. Injection was carried out with a syringe coupled to a needle with a flat point and it caused  
179 a cloudy solution. Afterwards, the sample tube was vortexed for 30 s and it was centrifuged for 5  
180 min at 9000 rpm. Phase separation occurred, obtaining the organic layer as sediment. It was  
181 carefully collected with a syringe coupled to a needle with a flat point and it was placed into a  
182 glass vial. It is important to avoid the collection of aqueous phase because of its high salt content.  
183 After the organic phase collection, it was evaporated to dryness under gentle nitrogen current and  
184 it was redissolved with 1.20 mL of an acidic aqueous solution (5.00 mM of orthophosphoric acid  
185 containing 5% of MeOH), vortexed for 2 min and filtered. Finally, 1.00 mL of the extract was  
186 analysed.

187

## 188 **2.5 CSEI-sweeping-MEKC procedure**

189 Water sample analysis was performed in an uncoated fused-silica capillary (57.2 cm x 50  
190  $\mu$ m i.d.). A new capillary was conditioned for the first time with NaOH 1 M for 15 min at 1 bar of  
191 pressure and 20°C. Afterwards and under the same conditions, capillary was rinsed with  
192 ultrapure water for 5 min and with a low conductivity buffer (LCB) for 15 min. LCB consisted of a  
193 50 mM phosphate buffer (NaH<sub>2</sub>PO<sub>4</sub>/H<sub>3</sub>PO<sub>4</sub>, pH = 2.5). Every day, capillary was washed with LCB  
194 for 15 min at 1 bar of pressure and 20°C. At the end of each day, capillary was washed with  
195 deionized water for 5 min at 5 bars followed by air flush for 5 min at 5 bars and at working  
196 temperature. Between runs, capillary was rinsed with NaOH 0.1 M at 3 bars for 2 min followed by  
197 ultrapure water at 3 bars for 0.5 min and lastly it was rinsed with LCB at 3 bars for 3 min.  
198 Subsequently, capillary was flushed with a high conductivity buffer (HCB) for 264 s at 50 mbars  
199 (capillary was filled a 31.5% of its total length) and 20°C. HCB consisted of a 100 mM phosphate  
200 buffer (NaH<sub>2</sub>PO<sub>4</sub>/H<sub>3</sub>PO<sub>4</sub>, pH = 2.5). Afterwards, inlet electrode was submerged into a vial  
201 containing ultrapure water for 5 s in order to wash it. Finally, an ultrapure water plug was injected  
202 into the capillary by applying a pressure of 50 mbar for 2 s. After capillary conditioning, samples  
203 were electrokinetically injected at 9.8 kV (normal mode) for 632 s. Electrophoretic separation was  
204 performed under a voltage of -30 kV programing a voltage ramp from 0 to -30 kV for 0.5 min at  
205 the beginning of the run. Separation temperature was 20°C. Background electrolyte (BGE)  
206 consisted of a phosphate buffer (44.0 mM, pH = 2.5) containing 123 mM of SDS and 8.00% of  
207 tetrahydrofuran (THF). UV detection was carried out at 276 nm, except for CRZ (244 nm).

208

209 **3. Results and discussion**

210

211 **3.1. Optimization of CSEI-sweeping-MEKC method**

212 Initially, separation was performed in a 48.5 cm x 50  $\mu$ m capillary. It was rinsed with LCB  
213 solution (25 mM phosphate buffer  $\text{NaH}_2\text{PO}_4/\text{H}_3\text{PO}_4$ , pH = 2.5) at 3 bars for 3 min at working  
214 temperature. Afterwards, HCB solution (150 mM phosphate buffer  $\text{NaH}_2\text{PO}_4/\text{H}_3\text{PO}_4$ , pH = 2.5)  
215 was flushed for 150 s into the capillary at 50 mbars and at work temperature. Inlet electrode was  
216 washed with ultrapure water for 5 s. A water plug (1 s at 50 mbars) was introduced prior to the  
217 sample injection. The sample, dissolved in 3 mM  $\text{H}_3\text{PO}_4$  solution, was injected for 600 s at 10 kV.  
218 5-NDZ separation was performed at -30 kV and 20°C, employing a buffer solution of 20 mM  
219 phosphate (pH = 2.5) containing 150 mM of SDS as BGE. Analytical signals were monitored at  
220 276 nm during the method optimization. Standard solutions of 70 ng/mL of each studied analyte  
221 (IPZ, ORZ, SCZ, TRZ and MNZ), except for CRZ (210 ng/mL), were used in this study.

222 Variables related to the CSEI-sweeping-MEKC methodology have been divided in three  
223 groups for their optimization: parameters affecting separation (buffer concentration, organic  
224 percentage in the buffer and surfactant concentration); chemical variables related to the capillary  
225 conditioning (concentration of LCB, concentration of HBC and capillary length filled with HCB);  
226 and instrumental variables related to the injection (voltage and injection time). These groups of  
227 variables were optimized through experimental designs in order to consider their interactions.  
228 Other parameters, such as separation pH, separation buffer nature, injection media, buffer  
229 concentration in the outlet vial during the sample injection, influence of a water plug before the  
230 sample injection and separation voltage were optimized univariately.

231

232 **3.1.1. Separation buffer pH and nature**

233 Separation buffer pH was studied in a narrow range (between 2 and 3). At lower pH  
234 values, electroosmotic flow (EOF) was very low and consequently worse analyte stacking effect  
235 was observed. At higher pH values, a stronger EOF was produced, and as consequence, EOF  
236 was higher than the electrophoretic velocity of the micelles. It resulted in analyte migration  
237 towards the cathode instead of from the inlet vial (cathode) to the outlet vial (anode). Considering  
238 the evaluated range, longer analysis times were observed at pH 3, while poorer stacking effect  
239 was observed at pH 2. Thus, separation pH was fixed to 2.5.

240           Different buffer natures as phosphate, oxalate and formate were studied. All the buffers  
241          were prepared at a concentration of 20 mM. Oxalate and formate buffers were prepared from  
242          their respective acids and adjusting the pH with NaOH, while phosphate buffer was prepared by  
243          mixing NaH<sub>2</sub>PO<sub>4</sub> and H<sub>3</sub>PO<sub>4</sub>. Phosphate buffer was chosen as the optimum because of the  
244          oxalate buffer showed bad reproducibility while formate buffer gave the lowest peak heights.

245           In order to increase peak resolution, different organic solvents were evaluated as  
246          modifiers of the separation buffer. (10% of MeCN, MeOH, isopropanol and THF). In all cases,  
247          similar peak resolution was shown and the only improvement observed was in terms of peak  
248          height. Signal increase was more appreciable when THF was added to the separation buffer.

249           Negatively charged micelles are required in CSEI-sweeping methodologies. In this work,  
250          SDS has been selected, considering it has been the most employed surfactant in the application  
251          of this technique [36-40].

### 252 253          **3.1.2. Separation buffer concentration**

254           A central composite design was employed for the multivariate optimization of surfactant  
255          concentration, phosphate buffer concentration and percentage of THF contained in BGE. Ranges  
256          for each parameter were established as follows: from 10.23 to 70.77 mM for phosphate buffer  
257          concentration, from 49.32 mM to 200.7 mM for SDS concentration and between 0 and 15.07% for  
258          THF volume contained in the separation buffer. Equation 1 represents the employed response  
259          function (R.F.). Parameters in R.F. have been normalized respect to the maximum value shown  
260          in the experimental set. The selected R.F. considers the most critical parameters for the proposed  
261          5-NDZ separation. Lowest resolution was obtained between SCZ and TRZ peaks, so the  
262          normalized resolution ( $R_{norm}^{Peaks\ SCZ-TRZ}$ ) among both peaks was included in R.F. Other terms  
263          such as normalized CRZ and MNZ theoretical plates are also included in the R.F. (equation 1). It  
264          is due because preliminary studies indicated that CRZ presented the lowest sensitivity while MNZ  
265          is the most important 5-NDZ drug due to its wide used.

266           $R.F. = Plates_{norm.}^{CRZ} + Plates_{norm.}^{MNZ} + R_{norm}^{Peaks\ SCZ-TRZ}$    Equation 1.

267           In the corresponding analysis of variance (ANOVA), a second-degree quadratic model  
268          was assumed. Lack-of-fit was no significant at a confident level of 95.0% ( $p$ -value > 0.05). In this  
269          case, interaction between THF percentage and SDS concentration was significant ( $p$ -value =  
270          0.0131) as well as the quadratic interactions ( $p$ -value > 0.05). The optimum values for the studied  
271          variables were: 44 mM phosphate buffer, 123 mM SDS and 8% THF. These values were  
272          established for further experiments.

273

274 **3.1.3. Separation voltage and temperature**

275 Separation was carried out at negative polarity. Voltage values from -25 kV to -30 kV  
276 were evaluated. Values below -30 kV provided longer analysis times without any improvement on  
277 peak resolution. According to that, -30 kV was established as optimum. Due to the high voltage  
278 employed, low separation temperature was desired. Temperature was set at 20°C in order to  
279 avoid capillary heating because of Joule effect.

280

281 **3.1.4. Water plug prior sample injection**

282 Some reports attribute an improvement on sensitivity to a water plug introduced into the  
283 capillary prior to sample injection. Before sample injection, water plugs injected under pressure at  
284 50 mbars for 1, 2, 3 and 4 s were tested. In this case, a water plug did not involve an increase on  
285 sensitivity so it was suggested to consider a water plug for 2 s prior to sample injection because it  
286 produced a better injection reproducibility [48].

287

288 **3.1.5. Chemical parameters of CSEI**

289 Phosphate buffer was selected as LCB and HCB solution. Concentration of both  
290 solutions, together with the capillary length filled with HCB solution, were evaluated by a central  
291 composite design. Studied experimental domains ranged from 9.82 to 50.18 mM for LCB  
292 concentration, from 74.43 to 200.57 mM for HCB concentration, and finally, the injection time of  
293 HCB solution (under a pressure of 50 mbars) was studied from 60.18 to 241.82 s. The considered  
294 R.F is indicated in equation 2.

295 
$$R.F. = \frac{H^{MNZ}}{1+|1-Sym^{MNZ}|} \quad \text{Equation 2.}$$

296 In this R.F., only the symmetry ( $Sym^{MNZ}$ ) and height ( $H^{MNZ}$ ) of MNZ peak have been  
297 included. According to that, analyte stacking into the capillary has been evaluated considering  
298 MNZ peak as representative. A higher stacking effect involves higher sensitivity in terms of peak  
299 height, without any losing of peak symmetry. In the proposed experimental design, lack of fit was  
300 no significant ( $p > 0.05$ ) at a confident level of 95.0%. The obtained optimum values from the  
301 surface response were 50 mM phosphate in LCB, 100 mM phosphate in HCB and 190 s of HCB  
302 injection time (capillary was filled a 31.5% of its total length). The influence of LCB and HCB  
303 concentrations on R.F. was found to be significant at a confidence level of 95 % ( $p$ -value =

304 0.0030 and *p*-value = 0.0163, respectively) as well as quadratic HCB concentration interaction (*p*-  
305 value = 0.0344).

306

### 307 **3.1.6. Instrumental parameters for FESI performance**

308 A central composite design was again employed to optimize voltage and injection time.  
309 Variables were continuously studied from 5 to 15 kV for the injection voltage and from 5 to 15 min  
310 for the injection time. Equation 3 was established as R.F. in order to improve CRZ and MNZ peak  
311 signals. For that reason, CRZ and MNZ normalized peak height ( $[H_{norm.}^{CRZ} + H_{norm.}^{MNZ}]$ ) were  
312 included as R.F. terms. Because of shapeless peaks were shown under some of the employed  
313 injection conditions, a term that represents the number of symmetrical peaks  
314 ( $n_{gaussians\ integrable\ peaks}$ ) was also included in the R.F. The number of analytes ( $n_{analytes}$ )  
315 studied in this experimental design was six.

316 
$$R.F. = [H_{norm.}^{CRZ} + H_{norm.}^{MNZ}] \cdot \frac{n_{gaussian\ defined\ peaks}}{n_{analytes}} \quad \text{Equation 3.}$$

317 Lack-of-fit was no significant at a confident level of 95.0%. The maximum of the response  
318 surface corresponds to an injection voltage of 9.8 kV and an injection time of 10.53 min. The  
319 response surface for the described experimental design can be found in Figure 3.

320

### 321 **3.1.7. Sample injection medium**

322 Low conductivity sample matrices benefit the analyte injection into the capillary by FESI,  
323 obtaining an important signal enhancement. However, the presence of a certain amount of  
324 protons in sample matrices is needed for obtaining charged analytes, and consequently obtaining  
325 a satisfactory injection by FESI. The needed proton concentration is determined by the analyte  
326  $pK_a$  values. Because of that, a phosphoric acid solution was proposed as injection media. An  
327 organic solvent was added to the injection solution in order to decrease its conductivity. MeCN,  
328 MeOH and THF at concentrations of 10% (v/v) were evaluated. Although peak signal  
329 improvement was expected, results did not show this behaviour. It could be due to the increase of  
330 the injection solution viscosity when an organic solvent is added, involving lower analyte mobility  
331 and a lower amount of injected analytes, resulting in lower peak signals. However, the addition of  
332 organic additives was considered to guarantee its low conductivity and a satisfactory sample  
333 injection as well as better peak symmetries. Better peak symmetries were observed when MeOH  
334 was employed so its percentage in the injection buffer was evaluated in the range between 2 and  
335 20%. Slight differences were noticed in these experiments, so 5% was considered as optimum, in

336 order to avoid a high solution viscosity. Acid concentration was studied in the range between 1  
337 and 10 mM. Signals did not show any variation when acid concentration was higher than 3 mM.  
338 At lower concentrations, peak signal depended on each analyte. In this case, a concentration of 5  
339 mM was chosen as optimum because higher concentrations could result in high conductivity  
340 samples while lower concentration could be detrimental for charging the analytes.

341 In order to obtain satisfactory reproducibility for sample injection, the nature of the  
342 solution placed in the outlet vial during sample injection to close the circuit was evaluated. Better  
343 reproducibility was showed when HCB was used as solution contained into the outlet vial.

344 Figure 4 shows an electropherogram of standard samples analysed under the proposed  
345 CSEI-sweeping-MEKC method considering the optimum conditions.

346

347

348

### 349 **3.2. Method characterization for water samples**

350 For the first time, DLLME is proposed as sample treatment coupled to CSEI-sweeping-  
351 MEKC. In this work, DLLME was applied to different water matrices using conditions previously  
352 established [18]. As it is usual, separation method was optimized employing standard solutions of  
353 the analytes. However, for the method characterization with spiked matrices, analyte behaviours  
354 showed slight variations in terms of peak migration times. Lower peak resolution was shown for  
355 analytes in water samples after DLLME treatment, so a longer capillary was used. A capillary of  
356 57.2 cm length was employed for method characterization on real water samples. Considering  
357 that, the capillary must be filled with HCB solution for 264 s, reaching the same experimental  
358 conditions (31.5 % of the total length filled with HCB) that those proposed for this novel CSEI-  
359 sweeping-MEKC strategy.

360 ORZ determination in water samples was not possible due to the presence of matrix  
361 interference at the same migration time. However, new sample injection conditions compared to  
362 preliminary ones led to evaluate a new 5-NDZ. TNZ presents lowest  $pK_a$ , being injected under the  
363 actual injection conditions, although it was not injected under the previously optimized conditions.

364 All 5-NDZs were monitored at 276 nm, except CRZ which presents at maximum UV  
365 absorption at 244 nm.

366

#### 367 **3.2.1. Calibration curves and analytical performance characteristics**

368 Matrix-matched calibration curves for the studied analytes (CRZ, IPZ, SCZ, TRZ, MNZ  
369 and TNZ) were established using river water from Riofrío river (Granada, Spain) as  
370 representative matrix to characterize the present method. River water samples were spiked at  
371 2.00, 4.00, 8.00, 14.0 and 20.0 ng/mL for all analytes except to TNZ (4.01, 8.02, 16.0, 28.1 and  
372 40.1 ng/mL). Five aqueous samples were spiked at the same concentration level. They were  
373 processed following the previously described procedure and injected and analysed according to  
374 the developed CSEI-Sweeping-MEKC methodology. Peak area was considered as a function of  
375 the analyte concentration on the sample. A blank sample was treated and no interferences were  
376 co-migrating with any 5-NDZ. Figure 5 shows an electropherogram of a spiked river water sample  
377 at 14.0 ng/mL for each compound except for TNZ (28.1 ng/mL).

378 Statistical parameters calculated by least-square regression and the performance  
379 characteristics of the DLLME-CSEI-sweeping-MECK-UV method for water samples are shown in  
380 Table 1. Limits of detection (LODs) were calculated as  $3 \times S/N$  as well as by Long and  
381 Winefordner [49] and Clayton criteria [50]. Quantification limits (LOQs) were calculated as  
382  $10 \times S/N$ . Taking into account that the off-line preconcentration factor due to the DLLME applied as  
383 sample treatment resulted in only 4.2 times, all 5-NDZ compounds were able to be quantified at  
384 the low ppb levels using the proposed methodology, in spite of the poor sensitivity attributed to CE-  
385 UV methods.

### 387 **3.2.2. Precision study**

388 The precision of the method was evaluated in terms of repeatability (intra-day precision)  
389 and intermediate precision (inter-day precision) by the application of the proposed DLLME-  
390 MEKC-sweeping-MEKC-UV method to water from Riofrío river spiked at three different  
391 concentration levels (4.00, 8.00 and 20.0 ng/mL) for all analytes (CRZ, IPZ, SCZ, TRZ and MNZ)  
392 except for TNZ (8.02, 16.0 and 40.1 ng/mL). Repeatability was studied for seven samples  
393 (experimental replicates) while intermediate precision was assessed for five consecutive days  
394 analysing one sample each day. No instrumental replicates for samples were considered. FESI  
395 involves an exhaustive injection and an important depletion is produced in the vial as a  
396 consequence of a single injection [34]. The results expressed as RSD (%) of the peak areas are  
397 summarized in Table 2. Satisfactory results were obtained in terms of precision, being RSD (%)  
398 lower than 10% in almost cases for all 5-NDZ drugs.

### 400 **3.3.3. Trueness assessment**

401 Trueness assays were carried out over different spiked water samples. Water samples  
402 from Riofrío river (Granada, Spain) and Genil river (Granada, Spain) and well water samples  
403 (Badajoz, Spain) were evaluated. Water samples were spiked at three different concentration  
404 levels (4.00, 8.00 and 20.0 ng/mL) for all analytes (CRZ, IPZ, SCZ, TRZ and MNZ) except to TNZ  
405 (8.02, 16.0 and 40.1 ng/mL). Seven samples from Riofrío river for each concentration level were  
406 treated following the proposed method. In the case of Genil river water samples and well water  
407 samples, five samples were required for each evaluated concentration level. A blank of each type  
408 of water was processed, and an interference associated to the matrix was co-migrating with CRZ  
409 in the case of Genil river and well water samples. Recovery values could not be established for  
410 this compound in the above-mentioned matrices. For the rest of 5-NDZ drugs, the obtained  
411 results are presented on Table 3. In general, recoveries over 70% were obtained.

412

#### 413 **4. Conclusions**

414 In conclusion, a novel CSEI-sweeping-MEKC has been developed for 5-NDZ  
415 determination. Several of the involved chemical and instrumental variables have been  
416 chemometrically optimized through experimental designs.

417 For the first time, DLLME has been coupled to this CE-based methodology. DLLME is a  
418 miniaturized sample clean-up treatment with a low solvent consumption. Low conductivity  
419 samples were obtained after the DLLME procedure, accomplishing with the proposed CE  
420 requirements. So, DLLME has shown to be suitable as sample treatment prior to CSEI-sweeping-  
421 MEKC analyses. The proposed DLLME/CSEI-sweeping-MEKC has been successfully applied to  
422 the determination of 5-NDZ residues in environmental waters. River and well waters have been  
423 also tested. In spite of low CE-UV sensitivity, the developed CSEI-sweeping-MEKC method  
424 results in a high on-line preconcentration factor, reaching detection limits lower than 2.44 ng/mL  
425 for all the studied analytes. Satisfactory results have been achieved for repeatability and  
426 intermediate precision studies, obtaining RSD lower than 10 % in most cases. The combination of  
427 DLLME with the evaluated CE technique supposes a cheap and green alternative for monitoring  
428 5-NDZ residues in waters.

429

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## Figure captions

Figure 1.- Chemical structures of evaluated 5-NDZ compounds. \*pKa values have been obtained from Scifinder ® database.

Figure 2.- CSEI-Sweeping-MEKC procedure scheme. Adapted from reference [34]. Steps: A, capillary is rinsed with a low conductivity buffer (LCB), followed by a plug of a higher conductivity buffer (HCB) and a water plug; B, electrokinetic injection at positive polarity, being cationic analytes stacked at the interface between the water zone and the HCB zone; C, cationic analytes are stacked at the HCB zone because of the long injection, but not at the water or matrix zone; D, background electrolyte is placed in both ends of the capillary and a negative voltage is applied; E, ordinary MEKC separation takes place.

Figure 3.- Estimated response surface for voltage and injection time associated to FESI.

Figure 4.- Electropherograms of standard analyte solutions obtained by the proposed CSEI-Sweeping-MEKC procedure using an extended path capillary (48.5 cm x 50  $\mu$ m). Signals are monitored at 276 nm. A. 10 ng/mL of each analyte. B. 75 ng/mL of each analyte. Peaks (1) CRZ, (2) IPZ, (3) ORZ, (4) SCZ, (5) TRZ, (6) MNZ.

Figure 5.- Electropherogram of a spiked river water sample at 14 ng/mL of each compound, except for TNZ concentration (28.1 ng/mL) . Peaks (1) CRZ, (2) IPZ, (3) SCZ, (4) TRZ, (5) MNZ, (6) TNZ. Separation current is showed.

Table 1. Statistical and performance characteristics of the DLLME-CSEI-sweeping-MEKC-UV method for 5-NDZ determination in river water sample.

Analyte	Linear range (ng/mL)	R <sup>2</sup>	Analytical resolution (γ <sup>-1</sup> ) (ng/mL in sample)	LOD (ng/mL in sample)			LOQ (ng/mL in sample)
				S/N = 3	Long	Clayton	
CRZ	4.38-20.0	0.980	1.01	1.31	1.04	2.24	4.38
IPZ	2.05-20.0	0.998	0.23	0.61	0.23	0.57	2.05
SCZ	4.60-20.0	0.995	0.42	1.38	0.41	1.00	4.60
TRZ	2.10-20.0	0.995	0.46	0.63	0.44	1.05	2.10
MNZ	2.79-20.0	0.985	0.72	0.84	0.69	1.69	2.79
TNZ	8.14-40.1	0.980	1.86	2.44	1.87	4.26	8.14

Table 2.- Precision study for the proposed method for the determination of 5-NDZs in river water sample.

	Repeatability (% RSD; n = 7)			Intermediate precision (% RSD; n = 5)		
	Level 1	Level 2	Level 3	Level 1	Level 2	Level 3
CRZ	7.60	8.13	6.43	23.6	8.13	9.86
IPZ	15.9	11.3	4.32	24.3	3.15	8.40
SCZ	15.5	6.46	7.64	11.4	10.9	7.85
TRZ	4.88	7.70	6.91	10.7	3.80	5.00
MNZ	4.32	7.12	7.26	6.93	6.45	4.92
TNZ	3.91	7.00	9.80	4.13	3.80	5.58

Level 1: 4.00 ng/mL for CRZ, IPZ, SCZ, TRZ and MNZ; 8.02 ng/mL for TNZ.

Level 2: 8.00 ng/mL for CRZ, IPZ, SCZ, TRZ and MNZ; 16.0 ng/mL for TNZ.

Level 3: 20.0 ng/mL for CRZ, IPZ, SCZ, TRZ and MNZ; 40.1 ng/mL for TNZ.

Table 3.- Recovery (R) percentages for 5-NDZs in river and well water samples using DLLME as sample treatment.

	Riofrío river water (% R; n = 7)			Genil river water (% R; n = 5)			La Serena well water (% R; n = 5)		
	Level 1	Level 2	Level 3	Level 1	Level 2	Level 3	Level 1	Level 2	Level 3
CRZ	83.3	73.0	82.8	-	-	-	-	-	-
IPZ	87.9	95.1	93.2	72.0	86.1	73.1	66.0	68.5	63.7
SCZ	68.6	73.6	97.2	78.9	91.3	82.0	84.9	83.1	74.0
TRZ	85.3	96.9	93.8	90.5	76.7	76.9	74.0	72.9	69.3
MNZ	76.9	83.8	72.0	80.3	81.5	74.6	68.7	60.9	67.3
TNZ	85.1	91.8	100.0	87.6	87.1	88.1	81.9	73.2	79.3

Level 1: 4.00 ng/mL for CRZ, IPZ, SCZ, TRZ and MNZ; and 8.02 ng/mL for TNZ.

Level 2: 8.00 ng/mL for CRZ, IPZ, SCZ, TRZ and MNZ; and 16.0 ng/mL for TNZ.

Level 3: 20.0 ng/mL for CRZ, IPZ, SCZ, TRZ and MNZ; and 40.1 ng/mL for TNZ.

Figure 1

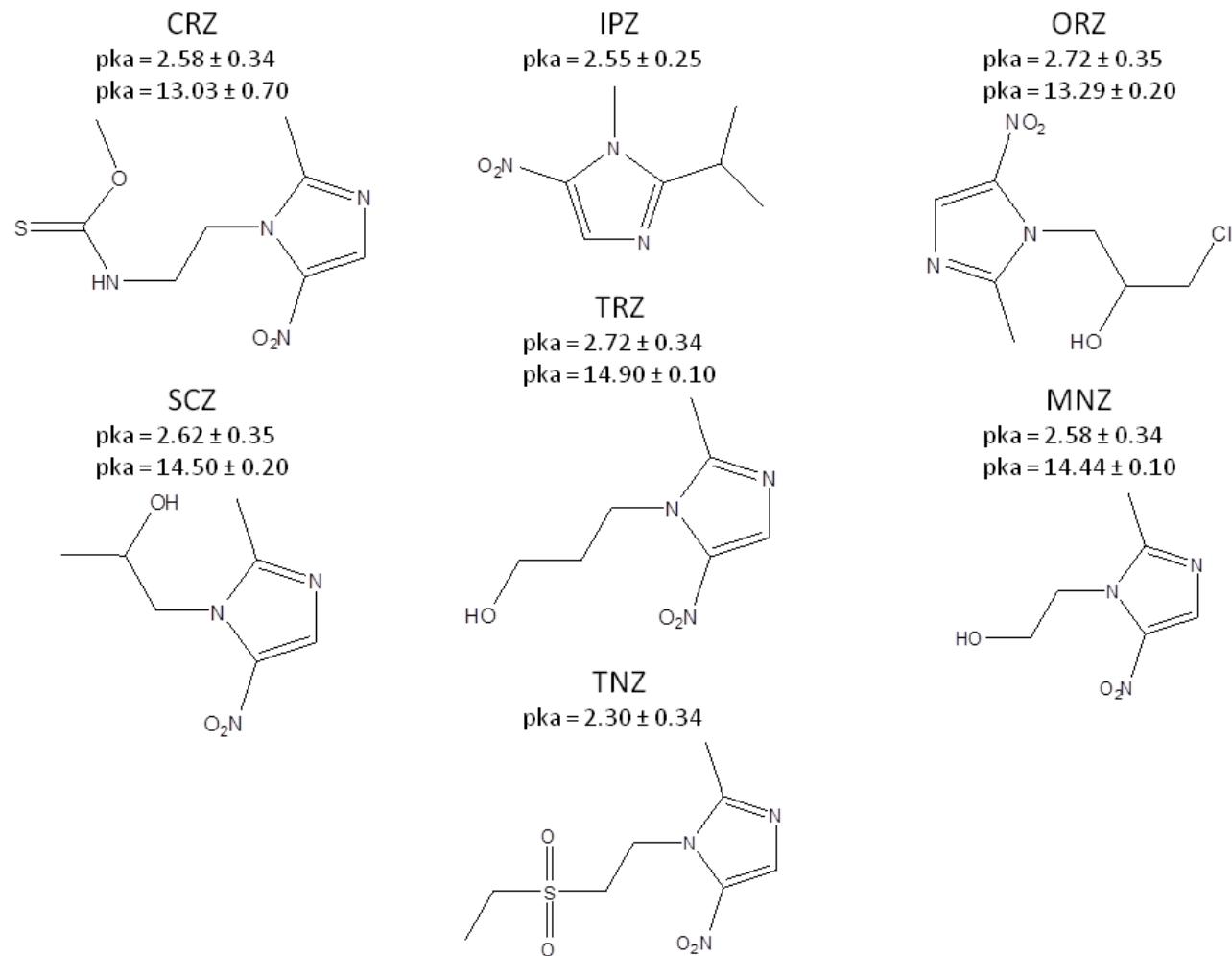


Figure 2

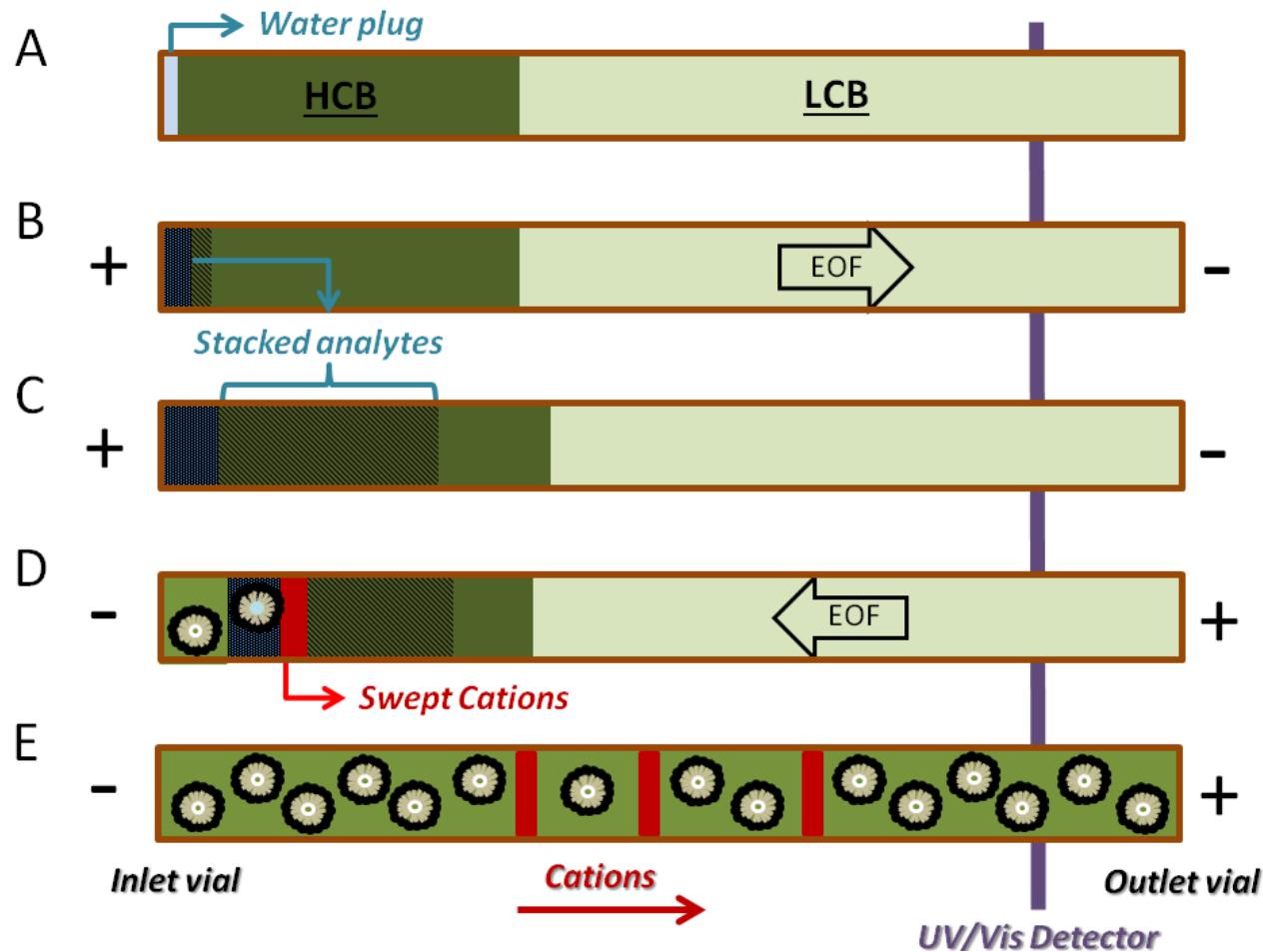


Figure 3

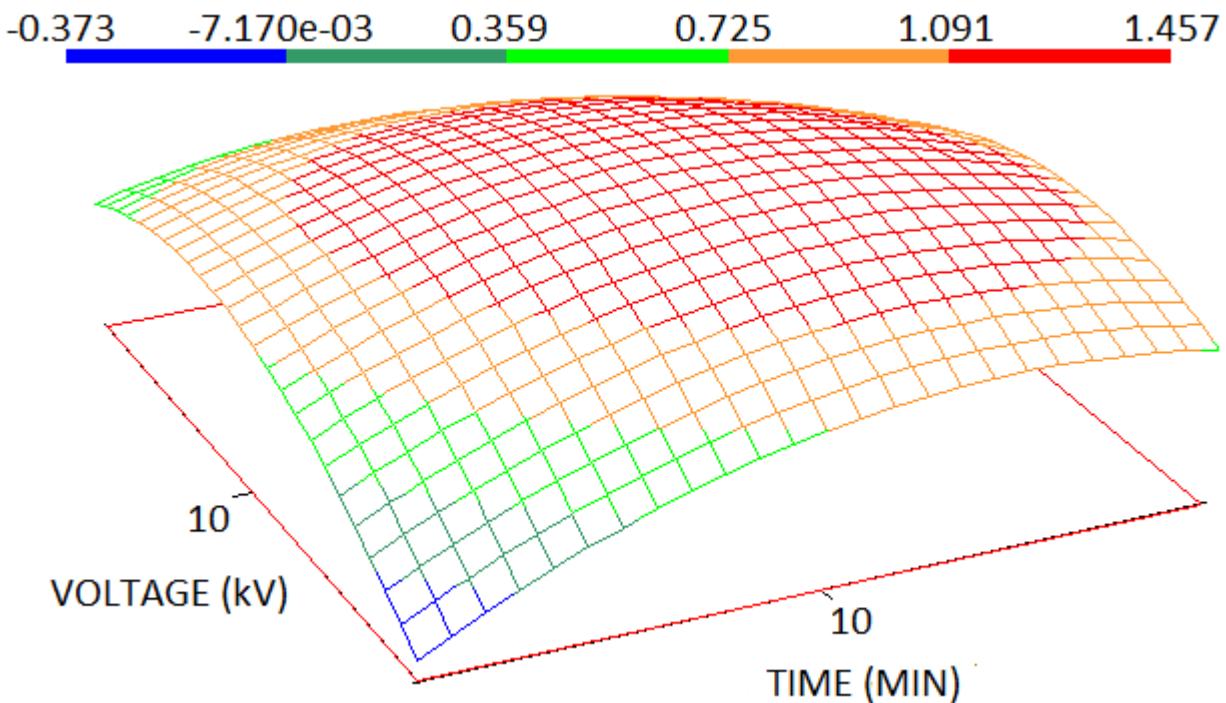


Figure 4.

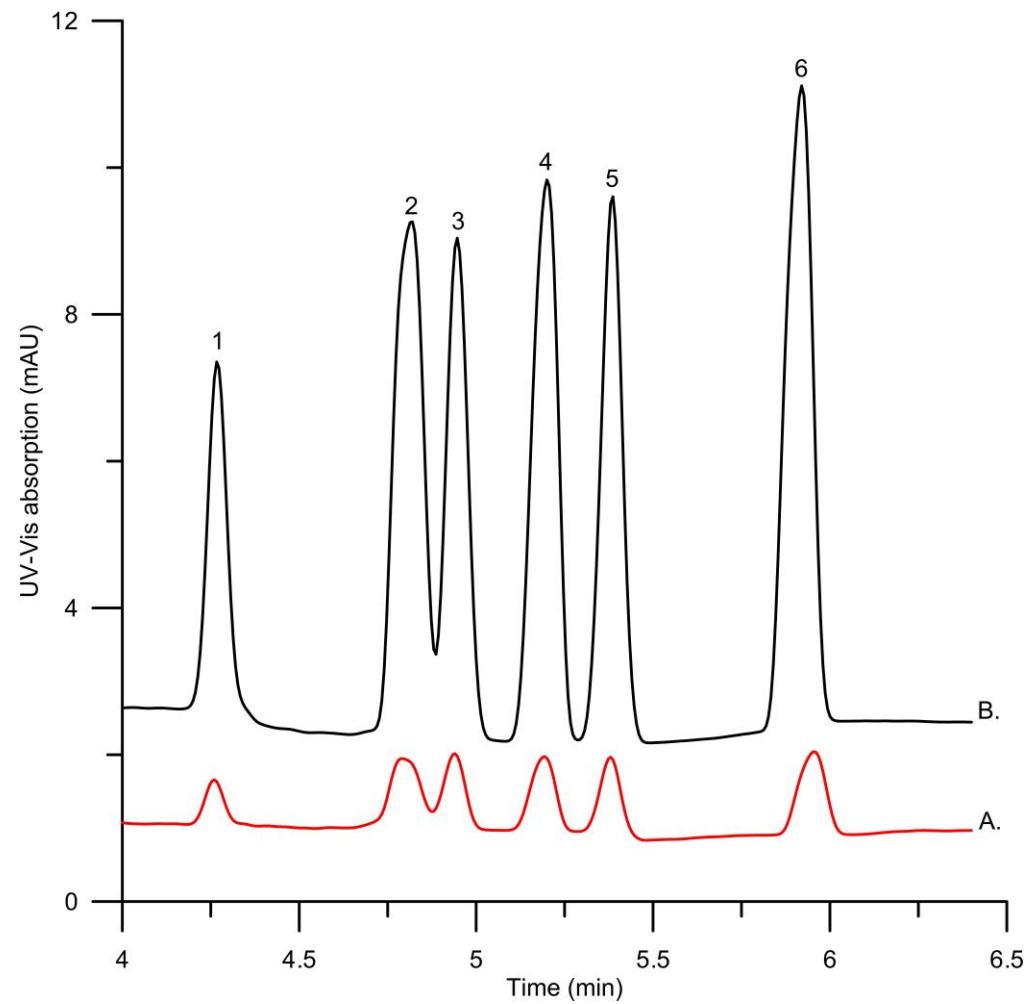


Figure 5

